

**Title page**

**Soil chemistry changes beneath decomposing cadavers over a one-year period**

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## Highlights

- decomposing cadavers affect below ground soil chemistry
- cadavers cause significant increases of ammonium, nitrogen, phosphorous and potassium in the first two month
- nitrate significantly increases after eight months
- pH increased significantly at first and then decreased significantly at the end of the experiment
- chemical markers can be assigned to groups and may have the potential to date the time since death (post-mortem interval)

1 **Abstract**

2 Decomposing vertebrate cadavers release large, localized inputs of nutrients. These  
3 temporally limited resource patches affect nutrient cycling and soil organisms. The impact of  
4 decomposing cadavers on soil chemistry is relevant to soil biology, as a natural disturbance,  
5 and forensic science, to estimate the post-mortem interval. But cadaver impacts on soils are  
6 rarely studied, making it difficult to identify common patterns.

7 We investigated the effects of decomposing pig cadavers (*Sus scrofa*) on soil chemistry (pH,  
8 ammonium, nitrate, nitrogen, phosphorous, potassium and carbon) over a one-year period in  
9 a spruce-dominate forest. Four treatments were applied, each with five replicates: two  
10 treatments including pig cadavers (placed on the ground and hung one metre above ground)  
11 and two controls (bare soil and bags filled with soil placed on the ground i.e. “fake pig”  
12 treatment).

13 In the first two months (15-59 days after the start of the experiment), cadavers caused  
14 significant increases of ammonium, nitrogen, phosphorous and potassium ( $p < 0.05$ ) whereas  
15 nitrate significantly increased towards the end of the study (263-367 days;  $p < 0.05$ ). Soil pH  
16 increased significantly at first and then decreased significantly at the end of the experiment.  
17 After one year, some markers returned to basal levels (i.e. not significantly different from  
18 control plots), whereas others were still significantly different. Based on these response  
19 patterns and in comparison with previous studies, we define three categories of chemical  
20 markers that may have the potential to date the time since death: early peak markers (EPM),  
21 late peak markers (LPM) and late decrease markers (LDM).

22 The marker categories will enhance our understanding of soil processes and can be highly  
23 useful when changes in soil chemistry are related to changes in the composition of soil  
24 organism communities.

25 **Keywords:** cadaver decomposition; soil nutrients; decomposition markers; disturbance;  
26 post-mortem interval (PMI)

## 27 **1. Introduction**

28 The vast majority of decomposing organic material in terrestrial ecosystems is either plant-  
29 derived or faecal matter, while cadavers only contribute marginally (ca. 1%) (Carter et al.,  
30 2007). However, although cadaver decomposition contributes quantitatively minimally to total  
31 ecosystem nutrient cycling, it can have a locally significant, although temporally limited,  
32 impact on the soil environment (Parmenter and MacMahon, 2009). Cadavers are nutrient-  
33 rich (Barton et al., 2013) and during decomposition, they release large amounts of water and  
34 breakdown products including proteins, fats and carbohydrates, which enter the underlying  
35 soil (Dent et al., 2004) and have a major impact on soil organisms (Carter et al., 2007;  
36 Szelecz et al., 2016, 2014). Understanding these effects is relevant for both soil ecology and  
37 forensic taphonomy and may help us develop new tools for the estimation of a post-mortem  
38 interval (PMI) i.e. the time elapsed since death (Carter et al., 2010; Haglund and Sorg, 1997).

39 Major transitions in the decomposition process are apparent on the cadaver and lead  
40 to the division into different decomposition stages i.e. fresh, bloated, active decay, advanced  
41 decay, dry and remains (Payne, 1965). Nevertheless, decomposition is a time-continuous  
42 process with overlapping and not clear-cut stages (Goff, 2009). Various abiotic and biotic  
43 factors can influence decomposition and accordingly its impact on soils. These factors  
44 include or may include temperature (Carter and Tibbett, 2006; Carter et al., 2008), moisture  
45 (Carter et al., 2010), pH (Haslam and Tibbett, 2009), soil type (Tumer et al., 2013), season  
46 (Meyer et al., 2013), access by insects (Campobasso et al., 2001), vertebrate scavenging  
47 (DeVault et al., 2003), associated material e.g. clothing (Matuszewski et al., 2014), burial  
48 (Forbes, 2008), trauma (open wounds) (Carter and Tibbett, 2008), size, age and type of  
49 carcass (Spicka et al., 2011; Stokes et al., 2013; Towne, 2000).

50 A range of decomposition studies exist, differing in experimental design (e.g. cadaver  
51 types, whole bodies or only parts, buried or placed on the soil surface). These studies show  
52 effects on soil pH (Aitkenhead-Peterson et al., 2012; Benninger et al., 2008), the  
53 concentration of ammonium (Meyer et al., 2013; Stokes et al., 2009a), nitrates (Anderson et

54 al., 2013; Meyer et al., 2013), total nitrogen (Anderson et al., 2013; Parmenter and  
55 MacMahon, 2009), total carbon (Hopkins et al., 2000; Macdonald et al., 2014), phosphorous  
56 (Macdonald et al., 2014; Towne, 2000), potassium (Aitkenhead-Peterson et al., 2012; Stokes  
57 et al., 2013), magnesium (Aitkenhead-Peterson et al., 2012) and calcium (Aitkenhead-  
58 Peterson et al., 2012; Melis et al., 2007) (Table 1 summarizes the results from the  
59 aforementioned studies that are relevant for this work). However, for some of these  
60 variables, knowledge remains very limited and the movement of carrion nutrients into soils is  
61 still an overlooked pathway (Barton et al., 2016),

62 We therefore investigated the impact of pig cadavers on selected soil chemical  
63 markers over a one-year period to include seasonal variation and to monitor the changes in  
64 soil chemistry beyond the peak decay stages. We compared the effects on soil chemistry of  
65 pig cadavers that were placed directly on the ground and pig cadavers that were hung one  
66 metre aboveground and contrasted them with two controls (bare soil and bags filled with  
67 soil). Our specific goals were to assess: 1) if changes in soil chemistry could be related to  
68 certain decomposition changes or time points and 2) if significant differences could be found  
69 between hanging and ground pigs.

## 70 **2. Material and Methods**

### 71 **2.1. Study site and experimental design**

72 The experiment was conducted in a small spruce (*Picea abies*) forest near Neuchâtel,  
73 Switzerland (47°01'05.01 N, 6°52'27.76 E, 775m a.s.l.). The study site is almost flat and  
74 covered an area of 1200 m<sup>2</sup>. Mean temperature and total precipitation (measured in-field with  
75 a Decagon Em50 digital data logger) were 10.2 °C and 978 mm. Further details are given in  
76 Szelecz et al. (2016) (Fig. 1, p. 407). The topsoil consisted of a litter layer (spruce needles  
77 and mosses), a fragmentation layer and a humification layer (O horizon, up to 1 cm) and an  
78 umbric horizon with a dark brown colour (A horizon, 1-17 cm) (Supplementary Material Fig.  
79 S1).

80 In total, 20 plots (ca. 4 m distant from each other) with four treatments (five replicates each)  
81 were set up randomly: 1) control (bare soil), 2) fake pigs (cotton bags filled with soil of the  
82 same size and weight as the pig cadavers for microclimatic effects), 3) ground pigs  
83 (cadavers directly placed on the ground for microclimatic and cadaveric fluids effects), and 4)  
84 hanging pigs (cadavers hanging 1 m above ground for cadaveric fluids effects).  
85 Ten domestic pigs (*Sus scrofa*), 8 females and 2 males, 10 weeks old, were bought from a  
86 local farm. They were sedated with Stresnil® (Azaperone) and euthanized with T61®  
87 (embutramide) by a veterinarian, immediately transported to the experimental site, weighed  
88 and placed on the plots. The average cadaver weight was 27.8 kg ± 0.8 kg (SE). All  
89 cadavers were placed in cages (140 cm x 95 cm) surrounded by wire mesh fences to keep  
90 scavengers and larger animals away. The experimental area was surrounded by an electric  
91 fence for additional protection. Control and fake pig plots were marked with bamboo sticks  
92 connected with cords. Wire mesh fences and cages could be opened at one side for soil  
93 sampling and weighing the cadavers. Cadavers were weighed just before placing and on  
94 every sampling day until D 331 using a digital hanging scale. Accordingly, soil from inside the  
95 fake pig bags was removed to match the weight loss of the pig cadavers.

## 96 **2.2. Decomposition stages and sampling**

97 Decomposition stages were estimated using the definitions provided by Payne (1965) for  
98 arthropod-exposed carrions. From the first day of cadaver placement (July, 01, 2013) until  
99 the beginning of the dry stage, each pig cadaver was examined daily to record the state of  
100 decomposition (including photographs and written reports) according to physical  
101 characteristics and arthropods present. After the beginning of the dry stage, the cadavers  
102 were examined at longer intervals (> 9 days).

103 On 11 sampling days from June 2013 until July 2014, a total of 220 soil samples (11 days x 4  
104 treatments x 5 replicates) were collected. Samples were initially taken shortly before the  
105 placing of the cadavers (D0), then on days 8, 15, 22, 36, 59, 84, 123, 263, 331 and 367  
106 (hereafter: D8, D15, D 22 asf.). A wooden rectangular frame (140 cm x 95 cm) with x (letters

107 A-N) and y (numbers 1-8) coordinates was placed on the ground at each site. At each  
108 sampling date, 10 points were randomly chosen from the x-y coordinates, excluding points  
109 outside of the surface directly impacted by the ground and hanging pig cadavers. These  
110 subsamples were taken with a bulb planter (6 cm diameter) to a depth of 10 cm, pooled and  
111 mixed to obtain one soil sample from each plot at each sampling day. Samples were stored  
112 at 4 °C until further processing.

### 113 **2.3. Chemical analyses**

114 Soil water pH was measured with a pH metre (Metrohm, 827 pH lab) after diluting the sample  
115 in water in a 1:2.5 proportion (Pansu and Gautheyrou, 2006). Ammonium ( $\text{NH}_4^+$ ) and nitrate  
116 ( $\text{NO}_3^-$ ) analyses were performed directly after sampling using colorimetric determination  
117 (Biochrom Libra S11 Spectrophotometer) (Scheiner, 2005). Total nitrogen (N) and carbon (C)  
118 were determined using a CHN analyser (Thermo Finnigan Flash EA 1112) on dry, ground  
119 soil. Bioavailable phosphorus ( $\text{P}_{\text{bio}}$ ) content was determined by colorimetric analysis  
120 (Biochrom Libra S11 Spectrophotometer) according to the Olsen method (Olsen et al., 1954).  
121 Potassium ( $\text{K}^+$ ) contents were determined using inductively coupled plasma optical emission  
122 spectrometry (Perkin-Elmer Optima 3300 DV ICP-OES) preceded by a cation exchange  
123 capacity extraction (CEC, cobaltihexamine method). All analyses were conducted at the  
124 Functional Ecology Laboratory, University of Neuchâtel, Switzerland.

### 125 **2.4. Grouping of chemical markers**

126 Based on the observed temporal patterns of soil chemical variables we defined three  
127 categories of markers:

128 (1) Early peak markers (EPM) show significantly higher concentrations in the soil beneath  
129 cadavers when compared to the controls at a certain point relatively early in the  
130 decomposition process (until the end of greatest cadaver mass loss and the end of the main  
131 leakage of cadaveric fluids).

132 (2) Late peak markers (LPM) show significantly higher concentrations in the soil beneath  
133 cadavers when compared to the controls at a certain point relatively late in the  
134 decomposition process i.e. not before the dry and remains stage.

135 (3) Late decrease markers (LDM) show significantly lower concentrations in the soil beneath  
136 cadavers when compared to the controls at a certain point relatively late in the  
137 decomposition process i.e. not before the dry and remains stage.

138 To be assigned to one of the categories a chemical marker had to be significantly different  
139 from both control treatments (control and fake) in at least one cadaver treatment (ground or  
140 hanging). In the case where peaks or decreases are followed by a relatively fast  
141 decrease/increase and levels discontinue being significantly higher or lower than the  
142 controls, markers are named EPM, LPM, LDM without any addition. In the case where peaks  
143 or decreases continue to be significantly higher/lower than the controls over a certain period  
144 of time either (+) EL (elevated levels) or (-) RL (reduced levels) will be added. If possible, the  
145 duration of EL or RL should be defined. Depending on their pattern, chemical markers may  
146 be attributed to one or more groups (or none if they show no pattern).

## 147 **2.5. Data analyses**

148 The duration of each decomposition stage was tested according to treatment (t-test adjusted  
149 according to Holm) to determine whether the length of the decomposition stages differed  
150 between hanging and ground pigs.

151 To test the significance of difference between treatments at each sampling day and overall,  
152 we used analysis of variance (ANOVA) and Tukey post hoc analysis (TukeyHSD). We  
153 assessed the significance over time using one- way ANOVA with repeated measure and post  
154 hoc multiple comparison of means (Tukey contrasts) with Bonferroni adjusted p-value. To  
155 follow the parametric assumptions of a normal distribution, variables were transformed (log  
156 10 or square root) before the analyses.

157 We explored the relationships between temporal changes in soil chemical variables and  
158 treatments using redundancy analysis (RDA) on previously transformed and standardised



159 variables. Day and treatment were used as explanatory variables and the fraction of variance  
160 explained by these variables quantified and their significance tested by Monte-Carlo  
161 permutation.

162 All statistical analyses were performed with R statistical software (version 3.1.0) (R Core  
163 Team, 2016) (R Core Team, 2016), and packages vegan, version 2.4.1 (Oksanen et al.,  
164 2016), nlme, version 3.1-128 (Pinheiro et al., 2016), multcomp, version 1.4-6 (Hothorn et al.,  
165 2008) and lme4, version 1.1-12 (Bates et al., 2015).

166

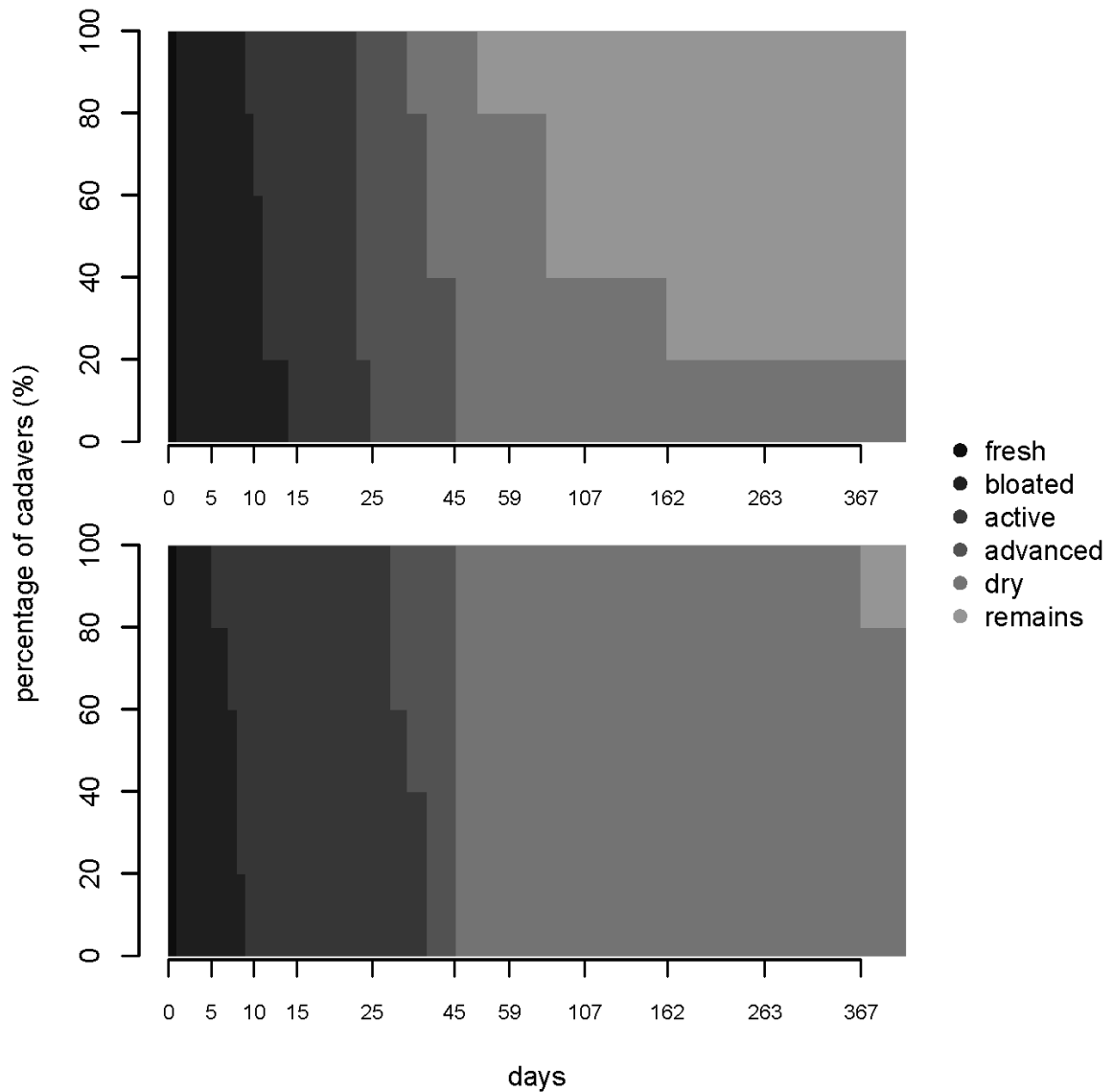
### 167 **3. Results**

#### 168 **3.1. Decomposition stages and mass loss**

169 At the end of the experiment (D367) four of the ground cadavers and one of the hanging  
170 cadavers had reached the remains stage, while one of the ground and four of the hanging  
171 pigs were still in the dry stage (Fig. 1). The bloated stage lasted on average twice as long for  
172 the ground cadavers as for the hanging cadavers (i.e. eight vs. four days;  $p < 0.05$ , t-test,  
173 adjusted p-value according to Holm). However, the active decay stage was significantly  
174 longer in the hanging cadavers ( $p < 0.01$ , t-test, adjusted p-value according to Holm) (Fig. 1).

175 Cadaver mass loss followed a sigmoidal pattern with the greatest mass loss before  
176 D59. At this point all cadavers had gone through the advanced decay stage with only bones  
177 and dry skin left. The mass loss from D59 onwards was more or less constant until the end of  
178 the experiment (Fig. 2).

179



180 **Figure 1.** Duration of decomposition stages, and percentage of cadavers representing a  
181 given decomposition stage in the ground (top) and hanging pig (bottom) cadaver treatments  
182 over time at the Bois-du-Clos spruce forest experimental site (Neuchâtel, Switzerland).  
183 Decomposition stages are shown in different shades of grey.

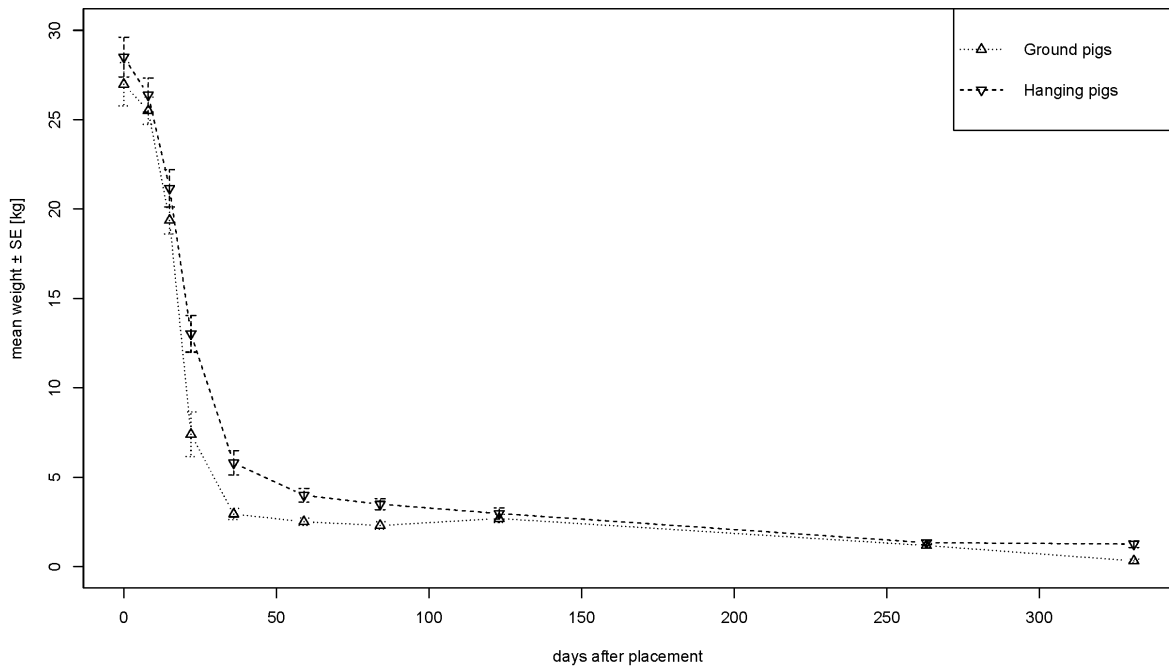
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cadaver weight loss over time



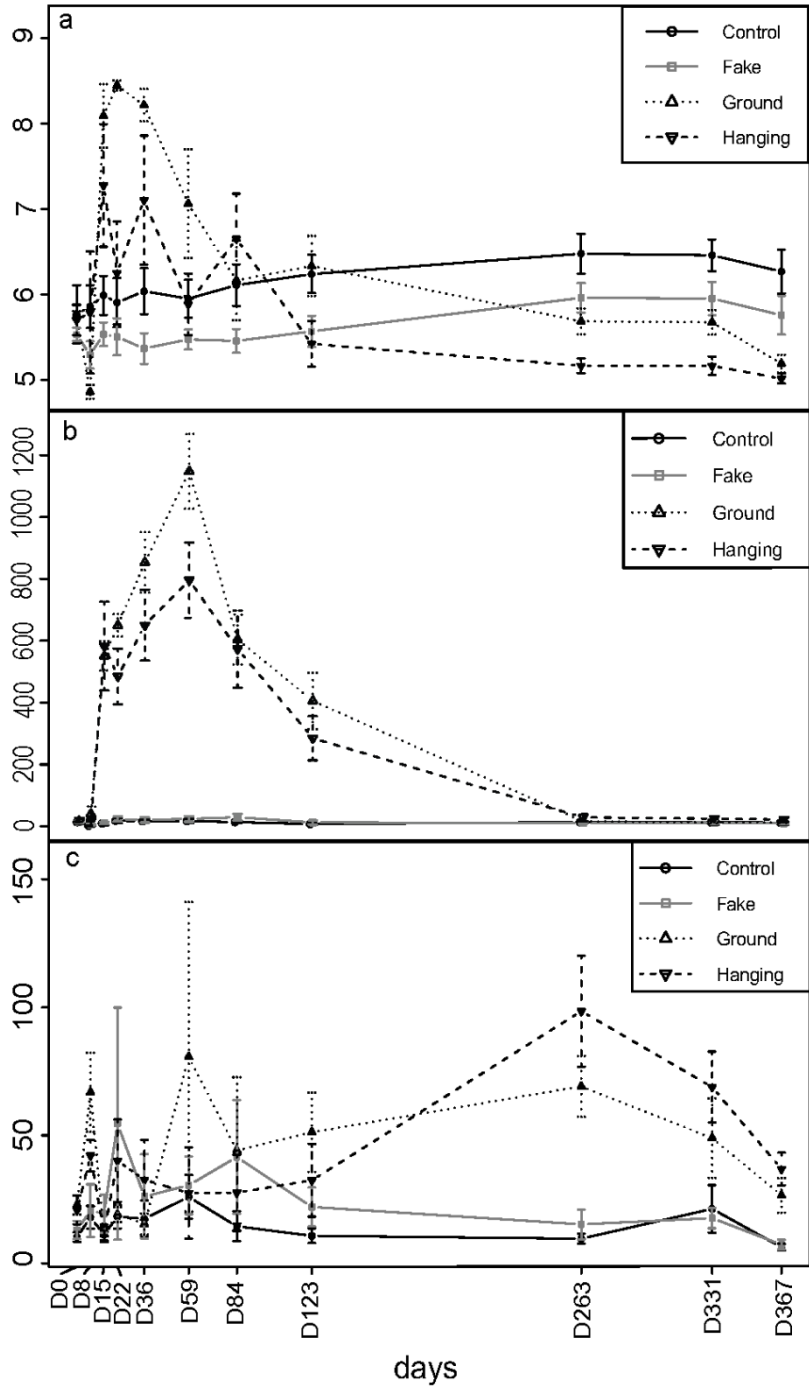
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189 **Figure 2.** Average cadaver weight loss  $\pm$  SE [kg] in the ground and hanging pig cadaver  
190 treatments over time at the Bois-du-Clos spruce forest experimental site (Neuchâtel,  
191 Switzerland).

### 192 3.2. Soil pH

193 Soil pH beneath the control and fake pigs fluctuated over the one-year period ranging from  
194 5.05 to 7.02 (controls) and 4.71 to 6.50 (fake pigs) (Table 2). In contrast, pH beneath the  
195 ground cadavers increased by 4.13 units (ranging from 4.63 to 8.76) and was significantly  
196 different in comparison to the control and fake pig samples from days 15 to 36 ( $p < 0.05$ ,  
197 ANOVA, TukeyHSD) (Table 2, Fig. 3a). Additionally, it was significantly higher to the hanging  
198 cadavers samples on D22 ( $p < 0.05$ , ANOVA, TukeyHSD). This increase was followed by a  
199 decrease reaching significantly lower pH values as compared to the control from D263 to  
200 D367 ( $p < 0.05$ , ANOVA, TukeyHSD) (Fig. 3a). In comparison, the increase in pH beneath the  
201 hanging cadavers (ranging from 4.68 to 8.70) at the beginning of the experiment was  
202 weaker, but the decrease towards the end of the experiment was also significant (D263-

203 D367) when compared to the control and fake pig treatment ( $p < 0.05$ , ANOVA, TukeyHSD)  
204 (Table 2, Fig. 3a).



206 **Figure 3.** Average  $\pm$  SE for pH (a), Ammonium ( $\text{NH}_4^+$ ) content [ $\mu\text{g g}^{-1}$ ] (b) and Nitrate ( $\text{NO}_3^-$ )  
207 [ $\mu\text{g g}^{-1}$ ] (c) in the control, fake pig, ground pig and hanging pig treatments over time at the  
208 Bois-du-Clos spruce forest experimental site (Neuchâtel, Switzerland).

209

### 210 **3.3. Ammonium ( $\text{NH}_4^+$ )**

211 Ammonium content in the soil of the control and fake pig samples ranged from 0.92 to 50.57  
212  $\mu\text{g g}^{-1}$  in the control and 1.0 to 62.51  $\mu\text{g g}^{-1}$  in the fake pig samples (Table 2). There was a  
213 massive and significant increase in Ammonium content in the ground (ranging from 1.98 to  
214 1561.78  $\mu\text{g g}^{-1}$ ) and hanging pig samples (ranging from 0.64 to 1124.71  $\mu\text{g g}^{-1}$ ) from D15 to  
215 D123 with a peak on D59 in contrast to both controls ( $p < 0.0001$ , ANOVA, TukeyHSD) (Table  
216 2, Fig. 3b). Ammonium content returned to basal levels towards the end of the experiment  
217 with no significant differences between treatments on D263, D331 and D367 ( $p > 0.05$ ,  
218 ANOVA, TukeyHSD) (Fig.3b). Overall ammonium content differed significantly between  
219 cadaver treatments and controls ( $p < 0.0001$ , ANOVA, TukeyHSD) but not between hanging  
220 and ground cadavers or between fake pigs and control ( $p > 0.5$ , ANOVA, TukeyHSD).

### 221 **3.4. Nitrate ( $\text{NO}_3^-$ )**

222 Soil nitrate content ranged from 3.12 to 57.26  $\mu\text{g g}^{-1}$  in the control samples, from 3.36 to  
223 235.89  $\mu\text{g g}^{-1}$  in the fake pig samples and from 3.7 to 321.97  $\mu\text{g g}^{-1}$  in the ground and 3.67 to  
224 164.35  $\mu\text{g g}^{-1}$  in the hanging pig samples (Table 2).

225 Although fluctuations were observed, no significant differences were recorded  
226 between the treatments until D263 (Fig. 3c). Ground cadavers samples were significantly  
227 different from both controls on D263 and D367 ( $p < 0.01$ , ANOVA, TukeyHSD) and hanging  
228 cadavers samples accordingly on D263, D331 and D367 ( $p < 0.05$ , ANOVA, TukeyHSD) (Fig.  
229 3c). Overall nitrate content differed significantly between cadaver treatments and controls  
230 ( $p < 0.01$ , ANOVA, TukeyHSD) but not between hanging and ground cadavers or between  
231 fake pigs and control ( $p > 0.4$ , ANOVA, TukeyHSD).

### 232 **3.5. Nitrogen (N)**

233 Total nitrogen content ranged from 0.45 to 1.95 % in the control, 0.31 to 1.55 % in the fake,  
234 0.58 to 1.81 % in the ground and 0.57 to 2.78 % in the hanging cadavers treatment (Table 2).  
235 In the soil samples from beneath the ground and hanging cadavers nitrogen content  
236 increased at the beginning of the experiment and was significantly higher as compared to  
237 both controls on D15 and D22 ( $p < 0.05$ , ANOVA, TukeyHSD) (Fig. 4a). Nitrogen content in  
238 the cadaver samples stayed above the controls until D331, not significantly and without any  
239 clear pattern (Fig. 4a). Overall nitrate content differed significantly between cadaver  
240 treatments and controls ( $p < 0.0001$ ), hanging and ground cadavers or between fake pigs and  
241 control ( $p > 0.6$ , ANOVA, TukeyHSD).

242

### 243 **3.6. Bioavailable Phosphorous ( $P_{\text{bio}}$ )**

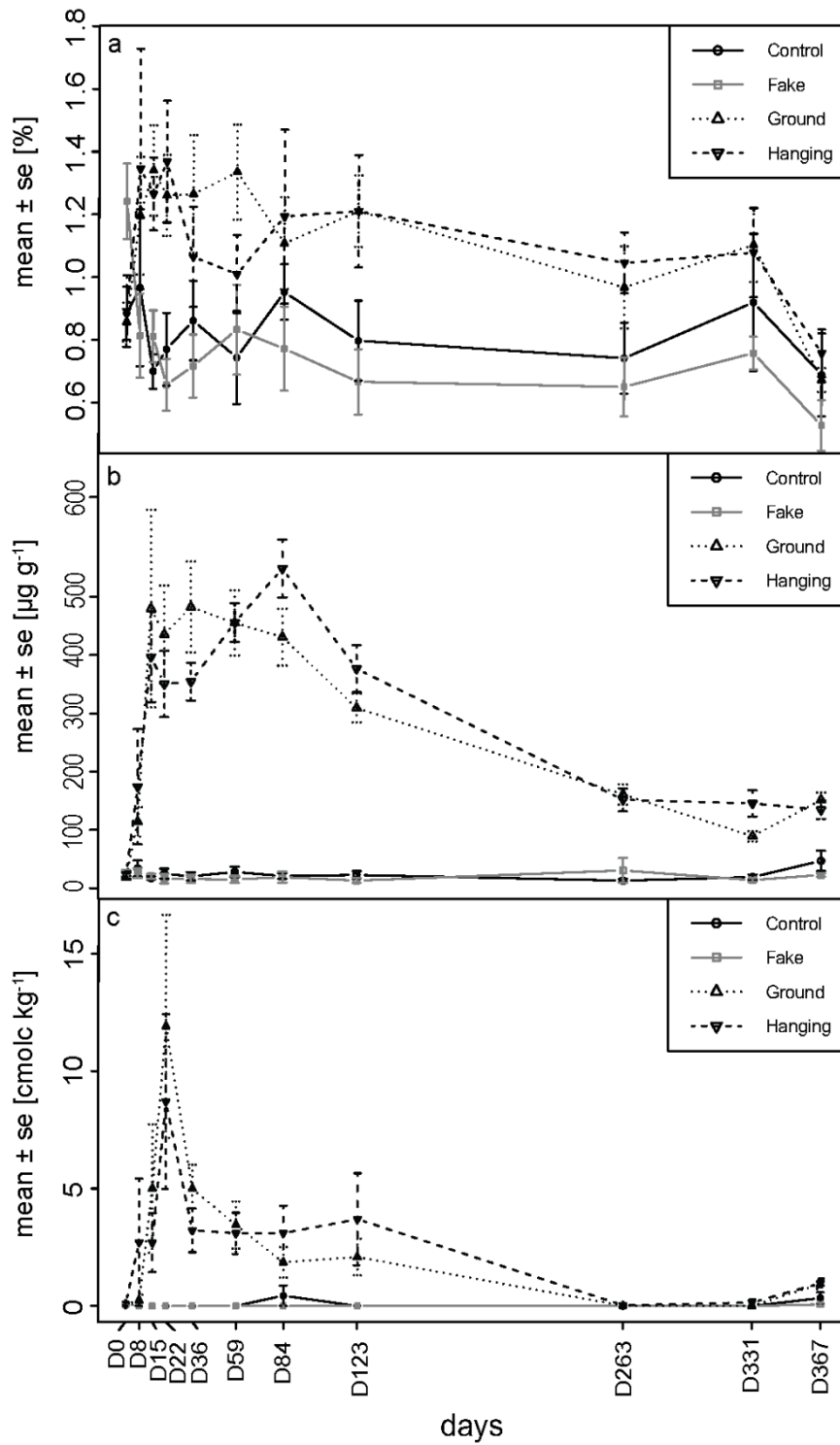
244 Bioavailable phosphorous content in soil ranged from 4.64 to 110.86  $\mu\text{g g}^{-1}$  in the control and  
245 from 0.56 to 114.41  $\mu\text{g g}^{-1}$  in the fake pig samples and varied slightly over the course of the  
246 experiment. In the ground and hanging pig samples it ranged from 10.96 to 1105.3  $\mu\text{g g}^{-1}$  and  
247 13.77 to 724.42  $\mu\text{g g}^{-1}$  respectively (Table 2).

248 In the early phase of decomposition (D15), there was a massive and significant  
249 increase in phosphorous content in both cadaver samples with a first peak on D15 and a  
250 second peak on D36 (ground cadavers) and D84 (hanging cadavers) ( $p < 0.0001$ , ANOVA,  
251 TukeyHSD; Fig. 4b). Although phosphorous decreased again after the second peaks, the  
252 content stayed significantly higher until the end of the experiment (D367) ( $p < 0.01$ , ANOVA,  
253 TukeyHSD; Fig. 4b). Overall phosphorous content differed significantly between cadaver  
254 treatments and controls ( $p < 0.0001$ , ANOVA, TukeyHSD) but not between hanging and  
255 ground pigs or between fake and control ( $p > 0.5$ , ANOVA, TukeyHSD).

### 256 **3.7. Potassium ( $K^+$ ) (exchangeable cation)**

257 Potassium concentrations in soil ranged from 0 to 2.2  $\text{cmol}_c \text{ kg}^{-1}$  in the control, 0 to 0.34  
258  $\text{cmol}_c \text{ kg}^{-1}$  in the fake pigs, 0 to 30.76  $\text{cmol}_c \text{ kg}^{-1}$  in the ground cadavers and 0 to 22.93  $\text{cmol}_c$   
259  $\text{kg}^{-1}$  in the hanging cadavers treatment (Table 2). Potassium content in the control and fake  
260 pig samples did not change over the course of the experiment (Fig. 4c). However, it  
261 increased in the ground and hanging cadavers samples at the beginning of the experiment  
262 and was significantly different from both controls from D36 until D59 ( $p < 0.05$  ANOVA,  
263 TukeyHSD). Overall potassium content was significantly different between cadaver  
264 treatments and controls ( $p < 0.001$ , ANOVA, TukeyHSD) but not between hanging and ground  
265 cadavers or between fake pigs and control ( $p > 0.9$ , ANOVA, TukeyHSD).





267 **Figure 4.** Average  $\pm$  SE for total Nitrogen (N) concentration [%] (a), bioavailable  
 268 Phosphorous ( $P_{bio}$ ) content [ $\mu\text{g g}^{-1}$ ] (b) and Potassium ( $K^+$ ) content [ $\text{cmol}_c \text{kg}^{-1}$ ] (c) in the

269 control, fake pig, ground pig and hanging pig treatments over time at the Bois-du-Clos spruce  
270 forest experimental site (Neuchâtel, Switzerland).

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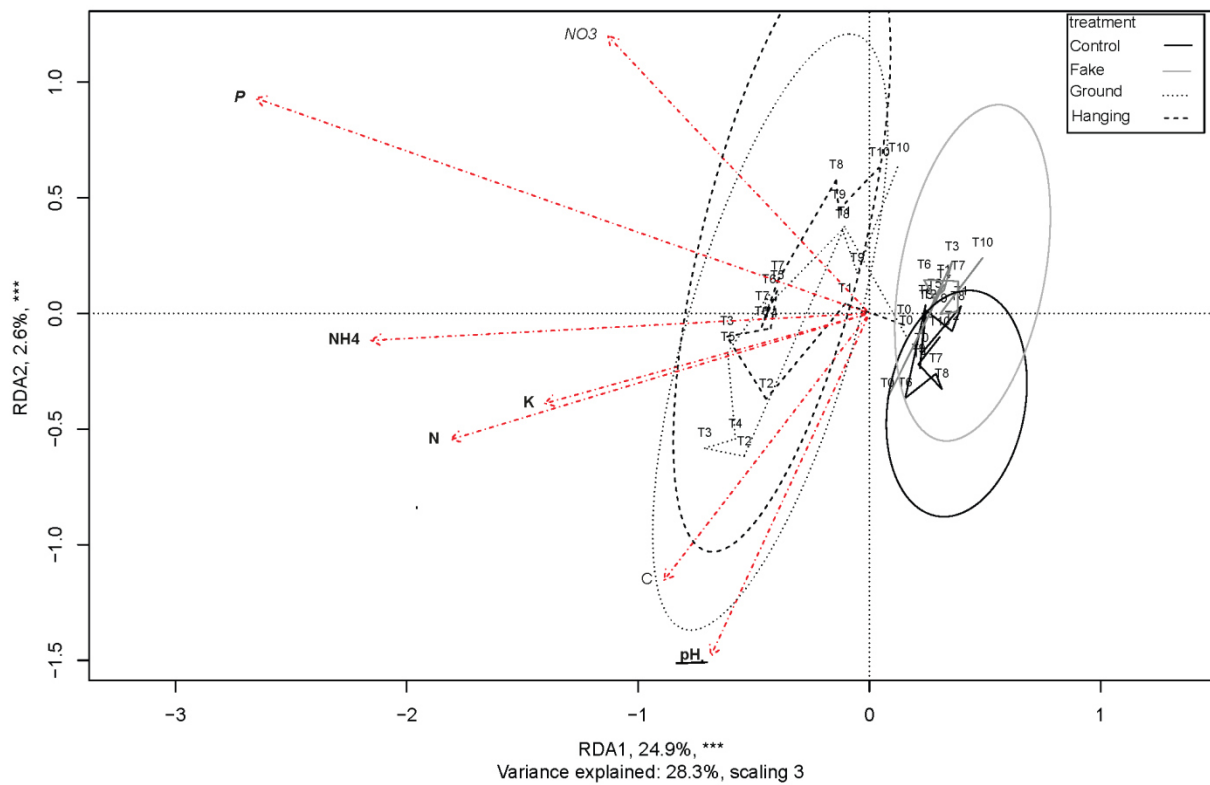
### 272 **3.8. Carbon (C)**

273 Soil carbon content ranged from 8.51 to 36.54 % in the control, 5.8 to 35.31 % in the fake,  
274 9.01 to 31.97 % in the ground and 8.78 to 36.68 % in the hanging pig cadavers treatment  
275 (Table 2). No significant differences between the four sets of samples were observed on any  
276 of the sampling days ( $p > 0.05$ , ANOVA, TukeyHSD; Fig. S2).

### 277 **3.9. Redundancy analysis (RDA)**

278 The RDA of the chemical variables (response variables) in function of time (sampling days)  
279 and treatments (explanatory variables) showed a clear difference between the cadaver  
280 treatments and the controls (axis 1) as well as temporal changes (axes 1 and 2) (Fig. 5). The  
281 ground and hanging cadavers' samples diverged from the control samples from T1 onwards  
282 (Fig. 5). Variables most strongly correlated with axis 1 and thus best explaining the difference  
283 between cadaver and control samples were P,  $\text{NH}_4^+$ , total N and  $\text{K}^+$ . Starting from D263-  
284 D367, cadaver-impacted samples started to converge back towards the control and fake pig  
285 samples. However, by T10 they clearly remained different from the control and fake pigs,  
286 owing mainly to higher nitrate concentrations.

287



288 **Figure 5.** Redundancy analysis (RDA) ordination diagram showing the response of soil  
 289 chemistry according to treatment (control, fake pig, ground pig and hanging pig treatments)  
 290 and time in a spruce forest at the Bois-du-Clos experimental site (Neuchâtel, Switzerland).  
 291 The lines (solid black: control; solid grey: fake; dotted: ground pig; dashed: hanging pig) join  
 292 the centroids of the five replicates from each sampling day. T0 to T10 represent the mean  
 293 coordinates of the 5 replicates per treatment (numbers indicating the time since death in  
 294 days). For better readability D0 was represented by T0, D8 by T1, D15 by T2 and so forth  
 295 until D367 by T10. Arrows represent the chemical i.e. explanatory variables i.e. NO<sub>3</sub>, P, NH<sub>4</sub>,  
 296 K, N, C, and pH. Ellipses show the SD of the mean position of every treatment (solid black:  
 297 control; solid grey: fake; dotted: ground pig; dashed: hanging pig). The grouping of the  
 298 chemical markers is indicated by different font styles: EPM (bold), LPM (italic), EPM+ LDM  
 299 (bold/underlined), and EPM+LPM (bold/italic).

300

301

### 302 **3.10. Grouping according to EPM, LPM and LDM**

303 Seven chemical soil markers (pH,  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ , N, C, P,  $\text{K}^+$ ) were investigated in all treatments  
304 and at all time points. The turning point from early ( $\leq$  D59) to late markers ( $>$  D59 -  $\leq$   
305 D367) in our study is two months after the cadavers were placed, which is after the greatest  
306 mass loss (Fig. 2) and the end of the main pulse of cadaveric fluids into the soil (after  
307 advanced decay) (Fig. 1). Based on significant differences between controls and cadaver  
308 treatments, chemical markers were grouped into three categories: early peak markers  
309 (EPM), late peak markers (LPM) and late decrease markers (LDM) (Table 3, Figure 5). As  
310 some chemical markers could be attributed to more than one category, in this analysis five  
311 groups of markers could be identified:

312 EPM: Nitrogen and potassium

313 LPM: Nitrate

314 EPM and LDM: pH

315 EPM with continuing elevated levels (ELs): Ammonium

316 EPM and LPM followed by EL: Phosphorous

317 No category: Carbon

318 No (+) RL (reduced levels) could be assigned.

### 319 **4. Discussion**

320 In both cadaver treatments mass loss followed a sigmoidal pattern in line with the classical  
321 pattern of breakdown of cadaver tissue and release of fluids taking place at the beginning of  
322 the decomposition process (Carter et al., 2007; Spicka et al., 2011). The longer active decay  
323 stage in the hanging cadavers was due to a lower insect activity (especially beetles) on the  
324 hanging cadavers (unpublished data) and the continuous dripping and loss of maggot  
325 masses from the hanging cadavers. However, overall in this study soil chemistry between  
326 ground and hanging cadavers did not reveal significant differences.

327 At the beginning of the experiment ( $>$ D15) soil pH,  $\text{NH}_4^+$ , N, P and  $\text{K}^+$  (EPMs)  
328 increased in at least one of the two cadaver treatments. On D15 all cadavers were in the

329 active decay stage, skin was ruptured and cadaveric fluids were released into the soil. The  
330 observed pattern is in line with the documented release of C-, N- and P-based products into  
331 the soil due to proteins, lipids and carbohydrates degradation from vertebrate cadavers  
332 (Stokes et al., 2009b).

333 During these processes an increase of soil pH in our study was observed beneath the  
334 ground cadavers as compared to the controls. In previous studies, soil pH has been shown  
335 to either decrease and increase beneath human and other mammal remains (Aitkenhead-  
336 Peterson et al., 2012; Benninger et al., 2008). In our study the increase of pH is probably due  
337 to an accumulation of ammonium- ions that follow the same pattern as shown by Benninger  
338 et al. (2008). Therefore, pH and  $\text{NH}_4^+$  can be regarded as EPMs. It is suggested that during  
339 and after the release of cadaveric fluids the soil beneath cadavers becomes more and more  
340 anoxic for a while, which would explain why  $\text{NH}_4^+$  ions were not further nitrified (Aitkenhead-  
341 Peterson et al., 2015).

342 Although pH beneath the hanging cadavers was also elevated at the beginning, it did  
343 not reach the significant values from the ground pig treatment. The dripping of the fluids and  
344 maggot masses probably did not cause a complete temporary shift to anoxia and did not  
345 cover the area beneath the cadaver completely. This would have allowed some nitrification to  
346 take place. The significant decrease of pH towards the end of the experiment in both cadaver  
347 treatments is in line with the decline of  $\text{NH}_4^+$  after >2 months and an increase of  $\text{NO}_3^-$ . This  
348 groups pH additionally into LDMs and  $\text{NO}_3^-$  into LPMs. It suggests a return of aerobic  
349 conditions allowing aerobic nitrification after an initial lag phase (Aitkenhead-Peterson et al.,  
350 2015; Stokes et al., 2013). This follows a pattern shown by Meyer et al. (2013) for  $\text{NH}_4^+$  and  
351  $\text{NO}_3^-$ , who suggested that ammonification is the dominant process up to advanced decay and  
352 nitrification after advanced decay. Significantly elevated  $\text{NO}_3^-$  was described after one and  
353 three years beneath decomposing pig cadavers (Anderson et al., 2013).

354 In our study, total N (EPM) increased two and three weeks after the beginning of the  
355 experiment in the cadaver treatments. Similar findings were observed by Benninger et al.  
356 (2008) showing an increase of total N in the first 14 days of the decomposition trial and

357 smaller peaks between days 21 and 42, and could be either the influx of organic or inorganic  
358 nitrogen forms. This is not surprising as a cadaver is a rich source for N for instance  $26\text{g kg}^{-1}$   
359 N concentration is reported for pigs (Benninger et al., 2008). The main N from cadavers  
360 derives from the breakdown of proteins, this process does not occur at a uniform rate and the  
361 degradation products can be released over a longer time- span including more  
362 decomposition stages (Macdonald et al., 2014). It might not be straightforward to group N  
363 into EPMS alone because other studies have shown that total N was significantly higher after  
364 one year beneath decomposing pigs (Anderson et al., 2013; Parmenter and MacMahon,  
365 2009). Here more data will be necessary.

366         Although carbon accounts for 20% of the mass of cadavers (Carter et al., 2007) no  
367 significant changes were observed in the soil beneath the cadavers, which is in line with  
368 other studies (Anderson et al., 2013; Benninger et al., 2008; Meyer et al., 2013). One reason  
369 for this might be that the intense pulse of C input caused an increase in micro-organisms that  
370 utilize carbon and then release  $\text{CO}_2$  into the atmosphere via respiration. Nevertheless,  
371 results are conflicting and some studies describe significant increases in total carbon  
372 beneath decomposing cadavers (Macdonald et al., 2014).

373         The input of P from cadavers, where P is stored in proteins, coenzymes, sugar  
374 phosphates and phospholipids (Dent et al., 2004), may translate into a large increase in soil  
375 as available P (Perrault and Forbes, 2016). In our study, bioavailable P peaked at the  
376 beginning of the experiment (EPM) but also on day 84 (LPM) and showed significantly  
377 elevated levels until the end of the experiment ((+) EL in the cadaver treatments when  
378 compared to the controls). Therefore, it cannot be assigned to just one category. Our results  
379 are in line with previous studies: The presence of a double peak was also noted by  
380 Benninger et al. (2008) and Perrault and Forbes (2016). Additionally, MacDonald et al.  
381 (2014) described a significant and lasting increase in plant available P relative to the control  
382 12 and 24 weeks after carcass addition and extractable P concentrations were described to  
383 be higher at carcass-impacted sites than in the surrounding soil one and three years post-

384 mortem (Towne, 2000). Phosphorous concentration seems to be a good indicator for locating  
385 the decomposition of remains (Perrault and Forbes, 2016).

386 Potassium was also grouped into the EPMs. Assuming that 100 g of pig body tissue  
387 contain approximately 280 mg K (Spray and Widdowson, 1950) being released into the soil  
388 relatively early in the decomposition process when tissues are broken down. Elevated K  
389 levels were also reported by Aitkenhead-Peterson et al. (2012) and Stokes et al. (2009a;  
390 2013) beneath decomposing cadavers and buried skeletal muscle tissues respectively.

## 391 **5. Conclusion**

392 The results from this and other studies indicate that it might be possible to categorize soil  
393 chemical markers according to their response pattern to decomposition products over time.  
394 As this is the first attempt to group cadaver-impacted soil chemical markers, we correlated  
395 the changes to decomposition stages and weight loss of the cadavers. A grouping into  
396 defined markers can be highly useful when the changes in soil chemistry are related to  
397 changes in the composition of soil organism communities. When applied in a forensic context  
398 a marker that shows clear and high peaks and/or decreases for a short period of time might  
399 be more useful than a marker that has elevated levels over a longer time-span to estimate  
400 the PMI. Chemical markers may thus be a useful addition to the forensic research toolkit  
401 when investigating homicides or other unclear death cases.

## 402 403 **Conflict of interest**

404 No conflict of interest declared.

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## 411 References

- 412 1. Aitkenhead-Peterson, J.A., Alexander, M.B., Bytheway, J.A., Carter, D.O., Wescott,  
413 D.J., 2015. Applications of soil chemistry in forensic entomology, in: Forensic  
414 entomology: international dimensions and frontiers. CRC Press, Boca Raton, pp.  
415 283–296.
- 416 2. Aitkenhead-Peterson, J.A., Owings, C.G., Alexander, M.B., Larison, N., Bytheway,  
417 J.A., 2012. Mapping the lateral extent of human cadaver decomposition with soil  
418 chemistry. *Forensic Sci. Int.* 216, 127–134. doi:10.1016/j.forsciint.2011.09.007
- 419 3. Anderson, B., Meyer, J., Carter, D.O., 2013. Dynamics of ninhydrin-reactive nitrogen  
420 and pH in gravesoil during the extended postmortem interval. *J. Forensic Sci.*, 1348-  
421 1352.
- 422 4. Barton, P. S., McIntyre, S.M., Evans, J., Bump, J. K., Cunningham, S. A., Manning, A.  
423 D., 2016. Substantial long-term effects of carcass addition on soil and plants in a  
424 grassy eucalypt woodland. *Ecosphere* 7(10):e01537. 10.1002/ecs2.153758, 1348–  
425 1352. doi:10.1111/1556-4029.12230
- 426 5. Barton, P.S., Cunningham, S.A., Lindenmayer, D.B., Manning, A.D., 2013. The role of  
427 carrion in maintaining biodiversity and ecological processes in terrestrial ecosystems.  
428 *Oecologia* 171, 761–772. doi:10.1007/s00442-012-2460-3
- 429 6. Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects  
430 models using lme4. *J. Stat. Softw.* 67. doi:10.18637/jss.v067.i01
- 431 7. Benninger, L.A., Carter, D.O., Forbes, S.L., 2008. The biochemical alteration of soil  
432 beneath a decomposing carcass. *Forensic Sci. Int.* 180, 70–75.  
433 doi:10.1016/j.forsciint.2008.07.001
- 434 8. Campobasso, C.P., Di Vella, G., Introna, F., 2001. Factors affecting decomposition  
435 and Diptera colonization. *Forensic Sci. Int.* 120, 18–27. doi:10.1016/S0379-  
436 0738(01)00411-X



- 437 9. Carter, D.O., Tibbett, M., 2008. Cadaver decomposition and soil: processes, in:  
438 Tibbett, M., Carter, D.O. (Eds.), *Soil Analysis in Forensic Taphonomy*. CRC Press,  
439 Boca Raton, pp. 29–51.
- 440 10. Carter, D. O., Tibbett, M., 2006. Microbial decomposition of skeletal muscle tissue  
441 (*Ovis aries*) in a sandy loam soil at different temperatures. *Soil Biol. Biochem.* 38,  
442 1139–1145. doi:10.1016/j.soilbio.2005.09.014
- 443 11. Carter, D.O., Yellowlees, D., Tibbett, M., 2010. Moisture can be the dominant  
444 environmental parameter governing cadaver decomposition in soil. *Forensic Sci. Int.*  
445 200, 60–66. doi:10.1016/j.forsciint.2010.03.031
- 446 12. Carter, D.O., Yellowlees, D., Tibbett, M., 2008. Temperature affects microbial  
447 decomposition of cadavers (*Rattus rattus*) in contrasting soils. *Appl. Soil Ecol.* 40,  
448 129–137. doi:10.1016/j.apsoil.2008.03.010
- 449 13. Carter, D.O., Yellowlees, D., Tibbett, M., 2007. Cadaver decomposition in terrestrial  
450 ecosystems. *Naturwissenschaften* 94, 12–24. doi:10.1007/s00114-006-0159-1
- 451 14. Clark, M.A., Worrell, M.B., Pless, J.E., 1997. Postmortem changes in soft tissues, in:  
452 Haglund, W.D., Sorg, M.H. (Eds.), *Forensic taphonomy: The postmortem fate of*  
453 *human remains*. CRC Press, Boca Raton, pp. 151–160.
- 454 15. Cobough, K.L., Schaeffer, S.M., DeBruyn, J.M., 2015. Functional and structural  
455 succession of soil microbial communities below decomposing human cadavers. *PLoS*  
456 *ONE* 10(6), e0130201. doi:10.1371/journal.pone.0130201
- 457 16. Dent, B.B., Forbes, S.L., Stuart, B.H., 2004. Review of human decomposition  
458 processes in soil. *Environ. Geol.* 45, 576–585. doi:10.1007/s00254-003-0913-z
- 459 17. DeVault, T.L., Rhodes, Jr., O.E., Shivik, J.A., 2003. Scavenging by vertebrates:  
460 behavioral, ecological, and evolutionary perspectives on an important energy transfer  
461 pathway in terrestrial ecosystems. *Oikos* 102, 225–234. doi:10.1034/j.1600-  
462 0706.2003.12378.x

- 463 18. Forbes, S.I., 2008. Decomposition chemistry in a burial environment, in: Tibbett, M.,  
464 Carter, D.O. (Eds.), Soil analysis in forensic taphonomy: chemical and biological  
465 effects of buried human remains. CRC Press, Boca Raton, pp. 202–223.
- 466 19. Gill-King, H., 1997. Chemical and ultrastructural aspects of decomposition, in:  
467 Haglund, W.D., Sorg, M.H. (Eds.), Forensic taphonomy: The postmortem fate of  
468 human remains. CRC Press, Boca Raton, pp. 93–104.
- 469 20. Goff, L.M., 2009. Early post-mortem changes and stages of decomposition in  
470 exposed cadavers. *Exp. Appl. Acarol.* 49, 21–36. doi:10.1007/s10493-009-9284-9
- 471 21. Haglund, W.D., Sorg, M.H. (Eds.), 1997. Introduction to forensic taphonomy, in:  
472 Haglund, W.D., Sorg, M.H. (Eds.), Forensic taphonomy: The postmortem fate of  
473 human remains. CRC Press, Boca Raton, pp.1-9.
- 474 22. Haslam, T.C.F., Tibbett, M., 2009. Soils of contrasting pH affect the decomposition of  
475 buried mammalian (*Ovis aries*) skeletal muscle tissue. *J. Forensic Sci.* 54, 900–904.  
476 doi:10.1111/j.1556-4029.2009.01070.x
- 477 23. Hopkins, D.W., Wiltshire, P.E.J., Turner, B.D., 2000. Microbial characteristics of soils  
478 from graves: an investigation at the interface of soil microbiology and forensic  
479 science. *Appl. Soil Ecol.* 14, 283–288. doi:10.1016/S0929-1393(00)00063-9
- 480 24. Hothorn, T., Bretz, F., Westfall, P., 2008. Simultaneous inference in general  
481 parametric models. *Biometrical J.* 50, 346--363.
- 482 25. Macdonald, B.C.T., Farrell, M., Tuomi, S., Barton, P.S., Cunningham, S.A., Manning,  
483 A.D., 2014. Carrion decomposition causes large and lasting effects on soil amino acid  
484 and peptide flux. *Soil Biol. Biochem.* 69, 132–140. doi:10.1016/j.soilbio.2013.10.042
- 485 26. Matuszewski, S., Konwerski, S., Frątczak, K., Szafałowicz, M., 2014. Effect of body  
486 mass and clothing on decomposition of pig carcasses. *Int. J. Legal Med.* 128, 1039–  
487 1048. doi:10.1007/s00414-014-0965-5

- 488 27. Melis, C., Selva, N., Teurlings, I., Skarpe, C., Linnell, J.D.C., Andersen, R., 2007. Soil  
489 and vegetation nutrient response to bison carcasses in Białowieża Primeval Forest,  
490 Poland. *Ecol. Res.* 22, 807–813. doi:10.1007/s11284-006-0321-4
- 491 28. Metcalf, J. L. et al., 2016. Microbial community assembly and metabolic function  
492 during mammalian corpse decomposition. *Science* 351, 158–162.
- 493 29. Meyer, J., Anderson, B., Carter, D.O., 2013. Seasonal variation of carcass  
494 decomposition and gravesoil chemistry in a cold (Dfa) climate. *J. Forensic Sci.* 58,  
495 1175–1182. doi:10.1111/1556-4029.12169
- 496 30. Oksanen, J., Guillaume Blanchet, F., Friendly, M., Kindt, R., Legendre, P., McGlenn,  
497 D., Minchin, P.R., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H.,  
498 Szoecs, E., Wagner, H., 2016. Package 'vegan': Community Ecology Package. R  
499 package version 2.4-1.
- 500 31. Olsen, S.R., Cole, C.V., Watanabe, F.S., Dean, L.A., 1954. Estimation of available  
501 phosphorus in soils by extraction with sodium bicarbonate. United States Department  
502 of Agriculture, Circular No .939, 1-19.
- 503 32. Pansu, M., Gautheyrou, J., 2006. Handbook of soil analysis: mineralogical, organic  
504 and inorganic methods. Springer, Berlin, New York.
- 505 33. Parmenter, R.R., MacMahon, J.A., 2009. Carrion decomposition and nutrient cycling  
506 in a semiarid shrub—steppe ecosystem. *Ecol. Monogr.* 79, 637–661.
- 507 34. Payne, J.A., 1965. A summer carrion study of the baby pig *Sus scrofa* Linnaeus.  
508 *Ecology* 46, 592–602. doi:10.2307/1934999
- 509 35. Perrault, K.A., Forbes, S.L., 2016. Elemental analysis of soil and vegetation  
510 surrounding decomposing human analogues. *Can. Soc. Forensic Sci. J.* 49, 138–151.  
511 doi:10.1080/00085030.2016.1184840
- 512 36. Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., R Core Team, 2016. Package 'nlme':  
513 Linear and Nonlinear Mixed Effects Models\_. R package version 3.1-128,  
514 <http://CRAN.R-project.org/package=nlme>.

- 515 37. R Core Team, 2016. R: A language and environment for statistical computing. R  
516 Foundation for Statistical Computing, Vienna, Austria.
- 517 38. Scheiner, J.D., 2005. Spéciation du carbone, de l'azote et du phosphore de  
518 différentes boues de stations d'épuration au cours de leurs incubations contrôlées  
519 dans deux types de sol, thèse pour obtenir le diplôme de docteur de l'2005. Institut  
520 National Polytechnique de Toulouse, Toulouse.
- 521 39. Spicka, A., Johnson, R., Bushing, J., Higley, L.G., Carter, D.O., 2011. Carcass mass  
522 can influence rate of decomposition and release of ninhydrin-reactive nitrogen into  
523 gravesoil. *Forensic Sci. Int.* 209, 80–85. doi:10.1016/j.forsciint.2011.01.002
- 524 40. Spray, C.M., Widdowson, E.M., 1950. The effect of growth and development on the  
525 composition of mammals. *Brit. J. Nutr.* 4, 332–353. doi:10.1079/BJN19500058
- 526 41. Stokes, K.L., Forbes, S.L., Tibbett, M., 2013. Human versus animal: contrasting  
527 decomposition dynamics of mammalian analogues in experimental taphonomy. *J.*  
528 *Forensic Sci.* 58, 583–591. doi:10.1111/1556-4029.12115
- 529 42. Stokes, K. L., Forbes, S.L., Tibbett, M., 2009a. Freezing skeletal muscle tissue does  
530 not affect its decomposition in soil: evidence from temporal changes in tissue mass,  
531 microbial activity and soil chemistry based on excised samples. *Forensic Sci. Int.* 183,  
532 6–13. doi:10.1016/j.forsciint.2008.08.013
- 533 43. Stokes, K.L., Forbes, S.L., Benninger, L.A., Carter, D.O., Tibbett, M., 2009b.  
534 Decomposition studies using animal models in contrasting environments: Evidence  
535 from temporal changes in soil chemistry and microbial activity, in: Ritz, K., Dawson,  
536 L., Miller, D. (Eds.), *Criminal and Environmental Soil Forensics*. Springer, pp. 357–  
537 377.
- 538 44. Szelecz, I., Fournier, B., Seppey, C., Amendt, J., Mitchell, E., 2014. Can soil testate  
539 amoebae be used for estimating the time since death? A field experiment in a  
540 deciduous forest. *Forensic Sci. Int.* 236, 90–98. doi:10.1016/j.forsciint.2013.12.030
- 541 45. Szelecz, I., Sorge, F., Seppey, C.V.W., Mulot, M., Steel, H., Neilson, R., Griffiths,  
542 B.S., Amendt, J., Mitchell, E.A.D., 2016. Effects of decomposing cadavers on soil

543 nematode communities over a one-year period. *Soil Biol. Biochem.* 103, 405–416.  
544 doi:10.1016/j.soilbio.2016.09.011

545 46. Towne, E.G., 2000. Prairie vegetation and soil nutrient responses to ungulate  
546 carcasses. *Oecologia* 122, 232–239. doi:10.1007/PL00008851

547 47. Tumer, A.R., Karacaoglu, E., Namli, A., Keten, A., Farasat, S., Akcan, R., Sert, O.,  
548 Odabaşı, A.B., 2013. Effects of different types of soil on decomposition: An  
549 experimental study. *Legal Med.* 15, 149–156. doi:10.1016/j.legalmed.2012.11.003

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reference	cadavers	time span/year	sampling days	country	pH	Ammonium	Nitrate	Nitrogen	Phosphorous	Potassium	Carbon
Altkenhead-Peterson et al., 2012	2 human bodies	2009-2010	288 (corpse 1) and 248 (corpse 2) days after	Texas, USA	lower (p<0.001)	–	–	higher (p<0.001)	higher (p<0.001)	higher (p<0.001)	higher (p<0.001)
Anderson et al., 2013	3 (2005) + 3 pigs (2007)	2005-2010	1 and 3 years after	Nebraska, USA	lower (1 year; p<0.05)	–	higher (1 + 3 years, p<0.05)	higher (p<0.05) after 1 year	–	–	–
Barton et al., 2016	12 kangaroos	2010-2015	5 years after	Canberra, Australia	–	–	–	–	higher (p<0.015)	–	–
Benninger et al., 2008	5 pigs	100 days (2006)	weekly (first 6 weeks), monthly after	Ontario, Canada	higher (D14, D23, D43; p<0.05) lower (D30, D72, D100; p<0.05)	–	–	–	–	–	–
Carter et al., 2008	juvenile rats	28 days	7,14,21,28 days after	Queensland, Australia	higher (D7- D28; p<0.001)	–	–	–	–	–	–
Cobaugh et al., 2015	4 human bodies	summer, autumn, 2012	up to 198 days after	Tennessee, USA	–	–	–	higher (p<0.05)	–	–	higher (p<0.05)
Hopkins et al., 2000	3 pigs	1996-1998	430 days after	England	elevated levels†	elevated levels	–	elevated levels	–	–	elevated levels
Macdonald et al., 2014	18 kangaroos	2010	0, 12, 24 weeks after	Canberra, Australia	higher (week 12, 24; p<0.001)	higher (week 12, 24; p<0.001)	–	higher (week 12, 24; p<0.001)	higher (week 12, 24; p<0.001)	–	higher (week 12; p<0.001)
Mellis et al., 2007	6 bisons	1997-2004	summer 2004	Poland	higher (1 to 6 years; p<0.0001)	–	higher (1 year, p<0.001)	–	–	–	–
Metcalfe et al., 2016	120 mice†	71 days	0,3,6,9,14,29,44,70 days after	Colorado, USA	higher (p<0.05)‡	higher (p<0.05)‡	higher (p<0.05)‡	higher (p<0.05)‡	–	–	–
Meyer et al., 2013	6 pigs	winter, 2008-2010 summer, 2008-2010	0,15,30,60 days after	Nebraska, USA	higher (D60; p<0.001) higher (D15; p<0.05) lower (D60;p<0.001)	higher (D60; p<0.05) higher (D15-D60, p<0.001)	higher (D60; p<0.05) higher (D15 (p<0.05)- D60( p<0.001))	higher (D60; p<0.05) higher (D30 (p<0.05), D60 (p<0.001))	–	–	–
Parmenter&MacMahon, 2009	various vertebrates†	all seasons, 3 years	15, 27, 39 months	Wyoming, USA	–	–	–	higher (first and second year)†/‡	–	higher†/‡	–
Stokes et al., 2009a	skeletal muscle tissue (pork)	37 days	2,4,6,8,12,16,23,30,37 days after	WA, Australia	higher (from D2; p<0.001)	higher (from D2; p<0.001)‡	higher (from D16; p<0.001)‡	–	–	higher (from D2; p<0.001)	–
Stokes et al., 2013	skeletal muscle tissue (human, pork, beef, lamb)	37 days	2,4,6,8,12,16,23,30,37 days after	WA, Australia	higher (from D2)‡	higher (from D2-D16/23)‡	higher (from D8/D12)‡	–	–	higher (from D2)	–
Towne, 2000	bison, cattle, deer	5 years	yearly	Kansas, USA	lower (p<0.01)‡	–	–	higher (1, 2 years after; p<0.05)	higher (1-3 years after; p<0.05)	–	–

† no significance given  
‡ see reference for details

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**Table 1.** Overview of selected studies on vertebrate cadaver decomposition and its effects on defined chemical markers in soil. Unless indicated, only significant differences are shown for the cadaver impacted soils in comparison to controls ("days, weeks, months, years after" refers to time elapsed since the beginning of the experiment i.e. the placing of the cadavers).

Table 2.					
		control	fake pig	ground pig	hanging pig
pH	mean ± [SE]	6.1 ± [0.08]	5.58 ± [0.05]	6.5 ± [0.18]	5.95 ± [0.16]
	min	5.05	4.71	4.63	4.68
	max	7.02	6.5	8.76	8.7
NH <sub>4</sub> <sup>+</sup> [μg g <sup>-1</sup> ]	mean ± [SE]	12.57 ± [1.4]	16.04 ± [2.03]	391.88 ± [54.84]	316.7 ± [45.88]
	min	0.92	1	1.98	0.64
	max	50.57	62.51	1561.78	1124.71
NO <sub>3</sub> <sup>-</sup> [μg g <sup>-1</sup> ]	mean ± [SE]	14.82 ± [1.63]	24.52 ± [5.07]	41.42 ± [6.8]	39.87 ± [4.85]
	min	3.12	3.36	3.7	3.67
	max	57.26	235.89	321.97	164.35
N [%]	mean ± [SE]	0.82 ± [0.04]	0.77 ± [0.04]	1.12 ± [0.05]	1.11 ± [0.06]
	min	0.45	0.31	0.58	0.57
	max	1.95	1.55	1.81	2.78
C [%]	mean ± [SE]	16.51 ± [0.85]	15.53 ± [0.87]	17.95 ± [0.71]	17.62 ± [0.78]
	min	8.51	5.8	9.01	8.78
	max	36.54	35.31	31.97	36.68
P [μg g <sup>-1</sup> ]	mean ± [SE]	24.39 ± [2.64]	19.89 ± [2.5]	284.29 ± [29.58]	283.03 ± [25.11]
	min	4.64	0.56	10.96	13.77
	max	110.86	114.41	1105.3	724.42
K [cmolc kg <sup>-1</sup> ]	mean ± [SE]	0.08 ± [0.05]	0.01 ± [0.01]	2.78 ± [0.66]	2.59 ± [0.55]
	min	0	0	0	0
	max	2.2	0.34	30.76	22.93

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560 **Table 2.** Chemical components in the control, fake pig, ground pig and hanging pig  
561 treatments over the course of the experiment at the Bois-du-Clos spruce forest experimental  
562 site (Neuchâtel, Switzerland) showing mean and standard error (SE), minimum (min) and  
563 maximum value (max).

Table 3.												
	T0	T1	T2	T3	T4	T5	T6	T7	T8	T9	T10	
days	0	8	15	22	36	59	84	123	263	331	367	Figures
pH	–	–	<b>EPM</b>	<b>EPM</b>	<b>EPM</b>	–	–	–	<i>LDM</i>	<i>LDM</i>	<i>LDM</i>	Fig. 3a
NH4+	–	–	<b>EPM</b>	<b>EPM</b>	<b>EPM</b>	<b>EPM</b>	(+) EL	(+) EL	–	–	–	Fig. 3b
NO3-	–	–	–	–	–	–	–	–	<i>LPM</i>	<i>LPM</i>	<i>LPM</i>	Fig. 3c
N	–	–	<b>EPM</b>	<b>EPM</b>	–	–	–	–	–	–	–	Fig. 4a
P	–	–	<b>EPM</b>	<b>EPM</b>	<b>EPM</b>	<b>EPM</b>	<i>LPM</i>	(+) EL	(+) EL	(+) EL	(+) EL	Fig. 4b
K	–	–	–	–	<b>EPM</b>	<b>EPM</b>	–	–	–	–	–	Fig. 4c
C	–	–	–	–	–	–	–	–	–	–	–	Fig. S2

565 **Table 3.** Grouping of chemical components into EPM (early peak marker), LPM (late peak  
566 marker), LDM (late decrease marker), +EL (+ elevated levels). The grouping of the chemical  
567 markers is indicated by different font styles: EPM (bold), LPM (italic), EPM+ LDM  
568 (bold/underlined), and EPM+LPM (bold/italic).

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