



The importance of willow thickets for ptarmigan and hares in shrub tundra: the more the better?

Journal:	Oecologia
Manuscript ID:	Draft
Manuscript Type:	original research paper
Date Submitted by the Author:	n/a
Complete List of Authors:	Ehrich, Dorothee; University of Tromsø, Department of Arctic and Marine Biology Henden, John; University of Tromsø, Department of Arctic and Marine Biology Ims, Rolf; University of Tromsø, Department of Arctic and Marine Biology Killengreen, Siw; University of Tromsø, Department of Arctic and Marine Biology Lecomte, Nicolas; University of Tromsø, Department of Arctic and Marine Biology Pokrovsky, Ivan; University of Tromsø, Department of Arctic and Marine Biology; Russian Academy of Sciences, A.N. Severtsov Institute of Ecology and Evolution Skogstad, Gunnhild; University of Tromsø, Department of Arctic and Marine Biology Sokolov, Alexander; Ural Division Russian Academy of Sciences, Ecological Research Station of the Institute of Plant & Animal Ecology Sokolov, Vasily; Ural Division Russian Academy of Sciences, Institute of Plant & Animal Ecology Yoccoz, Nigel; University of Tromsø, Department of Arctic and Marine Biology
Keywords:	habitat use, habitat fragmentation, occupancy, availability, large scale

SCHOLARONE™ Manuscripts

The importance of willow thickets for ptarmigan and hares in shrub tundra: the more the better?

Dorothée Ehrich¹, John-André Henden¹, Rolf Anker Ims¹, Siw Turid Killengren¹, Nicolas Lecomte¹, Ivan G. Pokrovsky^{1,2}, Gunnhild Skogstad¹, Alexander A. Sokolov³, Vasily A. Sokolov⁴, Nigel Gilles Yoccoz¹

Corresponding author: Dorothée Ehrich, Department of Arctic and Marine Biology, University of Tromsø, NO-9037 Tromsø, Norway. Email: dorothee.ehrich@uit.no, tel. +47 77646272, fax +47 77646020.

¹ Department of Arctic and Marine Biology, University of Tromsø, NO-9037 Tromsø, Norway.

² A.N. Severtsov Institute of Ecology and Evolution, Russian Academy of Sciences, 33 Leninskij prospekt, Moscow, 119071, Russia.

³ Ecological Research Station of the Institute of Plant & Animal Ecology, Ural Division Russian Academy of Sciences, 21 Zelyonaya Gorka, Labytnangi, Yamalo-Nenetski district, 629400, Russia

⁴ Institute of Plant & Animal Ecology, Ural Division Russian Academy of Sciences, 202 - 8 Marta street, Ekaterinburg, 620144, Russia

Abstract

1

2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

In patchy habitats, the relationship between animal abundance and cover of a preferred habitat may change with the availability of that habitat, resulting in a functional response in habitat use. Here we investigate the relationship of two specialized herbivores, willow ptarmigan (Lagopus lagopus) and mountain hare (Lepus timidus), to willows (Salix sp) in three regions of the shrub tundra zone - northern Norway, northern European Russia and western Siberia. Shrub tundra is a naturally patchy habitat where willow thickets represent a major structural element and are important for herbivores both as food and shelter. Using faeces counts in a hierarchical spatial design and several measures of willow thicket configuration we document a functional response in the use of willow thickets by ptarmigan, but not by hare. For hares, whose range extends into forested regions, occurrence increased overall with willow cover. For willow ptarmigan, occurrence showed a strong positive relationship to willow cover and a negative relationship to thicket fragmentation in the region with lowest willow cover at landscape scale, where willow growth may be limited by reindeer browsing. In regions with higher cover, on the contrary, such relationships were not observed. Differences in predator communities among the regions may contribute to the observed pattern, enhancing the need for cover where willow thickets are scarce. Such region-specific relationships reflecting regional specificities of the ecosystem highlight the importance of large-scale investigations to understand the relationships of habitat availability and use, which is a critical issue considering that habitat availability changes quickly with climate change and human impact.

1920

Key words: habitat use, habitat fragmentation, occupancy, availability, large scale

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

Introduction

The availability of suitable habitats determines the distribution of animals at different scales (Johnson 1980; Mayor et al. 2009; Orians and Wittenberger 1991). As animals select their habitat in function of their needs for all activities required for successful reproduction and survival, the optimal habitat is in fact often composed of a mixture of patches of several habitat types (Orians and Wittenberger 1991). Different habitats may for example be optimal for foraging and for shelter or breeding, resulting in trade-off situations (e.g. Mysterud et al. 1999). Habitat and landscape selection can in such cases vary in space in relation to changes in availability of important landscape elements (Fortin et al. 2008). A positive relationship between animal abundance and cover of a particular habitat may for instance be restricted to a certain range of cover values and flatten out or even decrease at higher values, meaning that the preference for a given type of habitat may change with its availability. Such a relationship has been defined as a functional response in habitat use by Mysterud and Ims (1998). Understanding how animal abundance and habitat area are related is an important question in ecology, in particular as the availability of habitats changes quickly with climate change and human impact. Shrubs provide important ecological functions in many open habitats, and biodiversity as well as productivity are often enhanced where shrubs are present (Ripple and Beschta 2005). Willow thickets are a characteristic component of shrub tundra vegetation (Chernov and Matveyeva 1997; Walker et al. 2005) and represent a good example of a patchy habitat. Willows (Salix sp.) usually grow along rivers or on slopes and are often the tallest plants in the tundra landscape, which is otherwise characterized by low-statured plants (Pajunen 2009; Pajunen et al. 2010). The thickets represent a major structural element and

are highly productive habitats compared to the surrounding open tundra vegetation. They

46	play an important role in ecosystem functioning, providing food, shelter and/or breeding
47	sites for numerous species of insects, birds and mammals (den Herder et al. 2004, 2008;
48	Ims et al. 2007; Henden et al. 2010). Under global change, the growth of shrubs is
49	increasing in the tundra, and shrubs are expanding northwards (Sturm et al. 2001; Tape et
50	al. 2006; Wookey et al. 2009). At the same time in some parts of the Arctic willow growth
51	is reduced and thickets are fragmented due to intense browsing mostly by reindeer
52	(Rangifer tarandus) (Den Herder et al. 2004, 2008; Forbes et al. 2009; Kitti et al. 2009).
53	These two opposing processes lead to varying areal extent of willows in climatically
54	comparable regions (Pajunen et al. 2010).
55	
56	Two important herbivores, the willow ptarmigan (Lagopus lagopus, hereafter ptarmigan)
57	and the mountain hare (Lepus timidus, hereafter hare) can be considered as willow
58	specialists in Arctic environments. Ptarmigan depend strongly on willow shrubs, which
59	constitute their most important food resource (West and Meng 1966; Andreev 1988; Elson
60	et al. 2007; Hakkarainen et al. 2007) and provide cover in an otherwise barren landscape,
61	in particular in winter (Estaf'ev and Mineev1984; Tape et al. 2010). Recently, Henden et
62	al. (submitted) documented increased occurrence of ptarmigan in patches with higher cover
63	of willow thickets in north-eastern Norway. At the same time, they reported a negative
64	effect of increased fragmentation of willow thickets. Hares are a wide-spread species in the
65	tundra and boreal forest of Eurasia (Kolosov et al. 1965; Newey et al. 2007). In the erect
66	shrub tundra in the northern part of their range willow thickets and their direct
67	surroundings are the optimal habitat for hare (Labutin 1988; Shtro 2006). Willow shrubs
68	are the hares' main food plant in winter and spring (Pavlinin 1997; Newey et al. 2007),
69	whereas they constitute a sheltered habitat for reproduction in summer (Labutin 1988).
70	

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

The relationship of these two specialized herbivores to willow thickets, whose extent varies strongly between regions in the shrub tundra, represent a good model system to investigate whether habitat use changes in relation to availability. The aim of our study is to determine whether regional abundance of ptarmigan and hare and the intensity of use (i.e. selection) of willow habitats, changes with the availability of this habitat, which is liable to change in response to impacts of climate and the abundance of browsing ungulates. Building on the study of Henden et al. (submitted) on ptarmigan in Finnmark, northernmost Norway, we applied the same method – counts of fecal pellets on permanent plots in replicate riparian landscapes – in two comparable tundra regions in the Russian Arctic. The three regions form a gradient in the amount of willow thickets. Specifically we asked whether there was a general positive relationship between the amount of willows at a large scale (landscape scale) and the occurrence of ptarmigan and hare, and whether these two herbivores consistently preferred habitats at the edge of willow thickets, also in regions with more willows. In a second step we investigated whether the positive effect of willow cover and the negative effect of fragmentation at a smaller local scale on the presence of ptarmigan reported by Henden et al. (submitted) were also observed in regions with more willow thickets, and whether hare reacted in the same way to differences in the configuration of willow thickets.

89

90

Material and methods

91 Study areas

- 92 The study was carried out in three regions within the southern arctic shrub tundra zone
- 93 (Walker et al. 2005): Finnmark in north-eastern Norway (Fig. 1; 70.4°N, 29°E), the
- Nenetsky Ridge in Nenetsky Autonomous Okrug, Russia (68.3°N, 53.3°E) and southern
- 95 Yamal, Russia (68.2°N, 69.1°E).

96

The study area in Finnmark has been described in detail in Henden et al. (2010) and Killengreen et al. (2007). The climate in this area is characterized by relatively mild winters due to the influence of the North Atlantic current and permafrost occurs only very scattered (Virtanen et al. 1999). Mean January temperature is -12.2°C and mean July temperature is 12.3°C. Mean annual precipitation is 455 mm, of which approximately 50% falls during the snow free period (meteostation Rustefjelbma, Norwegian Meteorological Institute, www.met.no). The landscape is mountainous with elevations up to 500 m asl. and sparse vegetation above 400 m. The mountain slopes are dominated by heaths mainly composed of dwarf shrubs, whereas the valleys are more productive and willow thickets surrounded by meadows grow on the riparian plains. The thicket communities in this region resemble the forb-rich types described by Pajunen et al. 2010, but differ from them in exact species composition. The coast near tundra on Varanger Peninsula in the eastern part of the area is classified as erect dwarf shrub tundra (Walker et al. 2005), but there is large intra-zonal variation due to topography and a variety of substrate types (Virtanen et al. 1999). The western part of the area at Ifjordfjellet lies in sub-arctic alpine tundra with similar main vegetation characteristics (Killengreen et al. 2007).

113

114

115

116

117

118

119

120

121

122

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

Nenetsky Ridge is situated in the buffer zone of the State Nature Reserve Nenetsky. The climate in this area is noticeably affected by the Arctic Ocean (van Erden 2000). Mean January temperature is -18.9°C and mean July temperature 13.3°C (World Meteorological Organisation: Zelenoye), with daily temperatures ranging from -40 to 30°C. Mean annual precipitation is 450 mm per year, of which 65% falls as rain during the frost free season (van Erden 2000). The Pechora floodplains and surrounding areas are dominated by glacial landforms on sediment ground with poor drainage. The region is at the western limit of continuous permafrost, with an active layer depth between 30 and 80 cm (van Erden 2000). Our study area is situated on the eastern slope of Nenetsky Ridge, which consists of a

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

nemorosa type.

tundra plateau with gentle slopes reaching up to 140 m asl. The study area includes three rivers with relatively deep (up to 70 m) and narrow (ca 300 m) valleys. Two of the valleys do not have a flood plain at their bottom, and the river flows directly between the two slopes. The river valleys are dominated by willow thickets (mainly S. glauca and S. phylicifolia) interspersed with lush meadow vegetation characterized by high species diversity and plants of high productivity (Skogstad 2009). The thicket communities in the valleys belong to the forb-rich types of Pajunen (2010), resembling the Salix-Trollius-Geranium type and the Salix-Comarum palustre-Filipendula ulmaria type. Some willow thickets grow also on the plateau, interspersed with tundra vegetation. The study area in Yamal is situated in the southern part of the Peninsula, close to the confluence of the Payutayaha and Erkutayaha rivers. The climate in southern Yamal is more continental than in the two other study areas. Mean temperature is -25.7°C in January and 8.6°C in July (Shiatov and Mazepa 1995, World Meteorological Organisation: Yarono). Mean annual precipitation is 350 mm per year, of which 70% fall as rain in the frost free season (Shiatov and Mazepa 1995). Most of the Yamal Peninsula consists of sandy and clayey marine, alluvial and lacustrine sediments (Walker et al. 2009; Pajunen 2009 and literature cited therein), and permafrost is continuous in the region (Pavlov and Moskalenko 2002). Our study area is characterized by flat tundra interspersed with hills (up to 40 m high) with sometimes steep slopes, and sandy cliffs along rivers. The tundra is subdivided by a dense network of rivers and lakes, and many low laying areas are flooded in spring. The area lies at the border between erect dwarf-shrub tundra and low-shrub tundra (Walker et al. 2005). Willow thickets are sometimes interspersed with Alnus and form communities which are close to the S. glauca-Carex aquatilis type (Pajunen et al. 2010). Some of the thicket communities can also be classified as Salix lanata-myosotis

Study design

Our study followed a hierarchical design with several nested levels. At the largest scale we compared the three study regions (Fig. 1). The three sub-regions in Finnmark will here be treated collectively as Henden et al. (submitted) did not find ptarmigan response to willow thicket variables to differ among them. The two regions in Russia comprised one study area each. In each study region sampling plots were arranged in units (two to five; see Fig. 1a), usually valleys. Within units, study plots were selected along willow thickets growing along the river as well as in the adjacent tundra (Fig. 1b-d). The selection of units and plots within units was made to cover the existing variation in willow thicket area and fragmentation within the unit. Units were separated by at least 2km. Willow thicket plots (W) and tundra plots within each unit were, as far as possible, arranged as pairs or triplets (Fig. 1b, c). Plots in tundra vegetation were thus chosen in proximity of W plots, however at least 30 m from the edge of meadows or thickets. The nearest neighbour distance between plots in the same habitat was on average 513 m (min = 129, max = 2359 m) and the distance between plots belonging to the same pair/triplet was on average 151 m (min = 36 m, max = 420 m).

The vegetation on W plots, a productive meadow dominated by herbaceous dicotyledons and grasses, placed with one side along a willow thicket (Fig. 1d; Henden et al. 2010), was chosen to be as homogenous as possible within and among the different study regions and to represent the most productive parts of the ecosystem. We chose willow thickets growing on riparian plains or valley/hill slopes and which were at least 0.5 m high. Thickets growing on rocks, mires, or which were flooded were excluded (cf Henden et al. 2010). Tundra plots, on the contrary, differed among the regions. In Finnmark, tundra plots were chosen to represent the dwarf shrub heath that dominates the tundra landscapes in northern

Fennoscandia (Virtanen et al. 1999; Moen 1998). These heaths are mainly composed by evergreen (*Empetrum nigrum hermaphroditum*) and deciduous dwarf shrubs (*Vaccinium* spp., *Betula nana*; Ravolainen et al. 2010). In Nenetsky, tundra plots were chosen in two of the most common vegetation types: Shrub tundra plots (S) were characterized by *B. nana* and ericoid shrubs (*Vaccinium* spp, *Rhododendron tomentosum*), interspersed with sedges (*Carex* spp) and *Rubus chamaemorus*. Hummock tundra plots (H) were dominated by cottongrass tussocks (*Eriophorum* spp) interspersed with dwarf shrubs and *R. chamaemorus* (Skogstad 2009). In Yamal, tundra plots were also chosen in two vegetation types which dominated in the landscape: dry tundra plots (D) were characterized by ericoid dwarf shrubs, mainly *R. tomentosum* but also *Vaccinium* spp, *B. nana* and *Eriophorum* spp, whereas on moist tundra plots (M) thick layers of *Shagnum* moss dominated together with *Carex* spp and *Eriphorum* spp tussocks, interspersed with *R. chamaemorus* and *B. nana*. Most tundra plots were situated on slopes or in the upland tundra, except the moist tundra plots in Yamal, which were placed in the lower flat tundra. Because of the configuration of the landscape, most plots were not grouped as triplets in Yamal.

Feces counts and willow thicket variables

Ptarmigan and hares produce conspicuous fecal pellets, which can be used as index of abundance and habitat use (hare: Krebs et al. 2001; ptarmigan: Evans et al. 2007; Ims et al. 2007). In Varanger, the willow ptarmigan is sympatric with the rock ptarmigan (*Lagopus muta*). However, the rock ptarmigan is using mainly other habitat types, at higher altitudes than considered in this study. Fecal pellets were counted in eight permanently marked small quadrates of 0.5 x 0.5 m arranged around a 15 x 15 m study plot (Fig. 1d). Counts were performed twice per year, shortly after snow melt in spring (spring) and in the second part of august/beginning of September (fall), from 2005 to 2009 in Finnmark (a few plots were excluded from the counts in 2009 because of a change in the monitoring protocol)

and from 2007 to 2009 in Russia. After counting, faeces were removed from the plots. As
faeces had not been removed previous to spring 2007 in Nenetsky and Yamal, the counts
from spring 2007 may represent cumulative use over more than one winter. This is,
however, unlikely to bias the results on relative habitat use.
The areal extent and degree of fragmentation of willow thickets were derived from aerial
photographs (Finnmark) as described in Henden et al. (2010), or from Quickbird satellite
images with a resolution of 0.6 m (Russian regions; DigitalGlobe TM 2001). For the Russian
regions, the outlines of the thickets were digitized in ArcGIS (ESRI TM). Thickets were
considered distinct when they were separated by an open area of at least 2 m, as such an
opening could be identified with reasonable confidence on the pictures. We quantified the
areal extent of willow thicket as percent willow cover in squares of 2 x 2 km (C-land) and
200 x 200 m (C-loc) using the software FRAGSTATS (McGarigal et al. 2002). Squares
were centred on each study plot (except for C-land in Finnmark where the measurement
was centred on each unit). Thicket fragmentation was quantified as patch density (number
of patches per 4 ha – PD) and edge density (meters of edge per 4 ha – ED) measured in
squares of 200 x 200 m centred on each plot. An increase in both of these measures reflects
increased fragmentation or shredding (cf Meffe and Carroll 1994) of willow thickets.
Choice of spatial scale is important in habitat selection studies (e.g. Mayor et al. 2009;
Henden et al. 2010). However, in lack of specific data on the scaling of area use of

individual hares and ptarmigan in the study regions the focal scales were chosen arbitrarily

based on the spatial constraints of the study design; the local scale was the largest possible

avoiding overlapping willow configuration measurements, whereas the landscape scale

corresponded roughly with the size of the study units.

227

228

229

230

231

The vertical structure of thickets at the edge of the plots was described by willow height (W-height) and density (W-density). These were measured at four points situated at 1 m inside the edge of the willow thicket along the side of the plot. Density was determined as the number of times a willow bush (leaf or branch) touched a vertical pole placed at the measuring point (point intercepts). The mean of the four measurements was taken as the value for each plot.

232

233

234

235

236

237

238

239

240

241

242

243

244

245

246

247

248

249

250

Statistical analysis

The data were analysed at two hierarchical levels for each species. First, in order to compare the effect of C-land in the three regions, the analysis was carried out at the level of the unit. Estimates of C-land originated thus from spatially non-overlapping squares for each replicate, minimizing spatial autocorrelation (Eigenbrod et al. 2010). The number of small quadrats where faeces were present was summed over all study plots belonging to the same habitat type within each unit, season and year and used as binary response variable (number of small quadrats with presence versus number of small quadrats with absence per habitat/unit/season/year, referred to as occurrence in the following). Generalized linear mixed effects model (GLMM) with a logit link and a binomial distribution were used for the analysis. Fixed effects were C-land (for the Russian sites an average value was used for each unit), habitat, region, season and year (as factors). We used only the years with observations in all three regions (2007-2009) for the statistical analysis. In order to be applicable in all regions, habitat was coded as W versus tundra plots, thus pooling the different tundra types (T, H, S, D, M). Unit identity was used as random effect to account for repeated measurements in the same plots. C-land was standardized by scaling it with mean = 0 and standard deviation = 0.5 to make effect estimates comparable with a the two level factor habitat (Gelman and Hill 2007). The

preference of ptarmigan and hare for the different tundra types in the Russian regions was analysed separately using Chi-square tests.

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

271

251

252

Second, we considered the effect of the configuration of willow thickets directly surrounding each W plot on habitat use by herbivores. Here we summed the number of small quadrats where faeces were present among the eight small quadrats arranged around each plot, and used it as a binary response variable in GLMM as above. C-loc quantified willow cover at this scale. We used PD as measure of thicket fragmentation. On the satellite picture from Yamal it was not always easy to trace edges precisely, and we considered thus PD a more robust indicator of willow fragmentation than ED in this case. Some of the willow configuration variables were strongly correlated, such as W-height and W-density in Yamal (r = 0.77; Electronic Supplementary Material (ESM) Table S1). We nevertheless included them into the analysis, as Smith et al. (2009) showed that in studies of habitat fragmentation it is best to include all variables despite possible correlations. Removing some of the variables can indeed lead to biased estimates given suppressor relationships between some variables (Smith et al. 2009). In addition to these four willow configuration variables, region, year and season were included as fixed effects, and plot identity was used as random effect. For all analyses, the best model was selected among eight candidate models comprising an additive model and models with interactions of willow and habitat variables with region, year or season. All willow variables were kept in all candidate models (Smith et al. 2009). In addition, an interaction of season with year was considered.

273

274

275

276

272

Statistical analyses were carried out in R version 2.9.2 (R Core development Team 2010).

GLMMs were fitted using the Laplace approximation as implemented in the lme4 package
(Bates et al. 2008). Log-Likelihood ratio tests were used to compare the candidate models

and a model was considered superior to the next simpler model when P < 0.05. Selected models were checked for constant variance of the residuals, presence of outliers and approximate normality of the random effects. A few (1 to 8) outliers were detected in the four analyses. However, as removing them did not alter the results qualitatively and only modified estimates slightly, all data were retained in the analysis.

Results

Regional patterns of willow thicket configuration

The extent, fragmentation, as well as the vertical structure of willow thickets differed considerably between the regions (Table 1). C-land was lowest in Finnmark, much higher in Nenetsky, and intermediate in Yamal. At the local scale, considering willow cover in the vicinity of W plots, the contrasts were not as strong, and C-loc was on average lowest in Yamal. The vertical structure of the willow thickets also exhibited a contrasting pattern. Willow thickets were lower in Yamal than in the two other regions, whereas thicket density was highest in Nenetsky (Table 1). Altogether the clearest contrasts in willow thicket configuration variables were between Nenetsky and the two other regions and variation among plots was smallest in Yamal (ESM Fig. S1).

Regional patterns of herbivore abundance

There were considerable differences in occurrence of ptarmigan and hare between the regions (Fig. 2). Overall, occurrence was highest in Nenetsky. Whereas ptarmigan occurred at similar levels in eastern Finnmark and in Yamal, hare was almost absent from Finnmark (Fig. 2). Therefore this region was excluded from the statistical analysis of hare occurrence. There was a clear seasonal effect for both species, with fewer faeces found in the fall (Fig. 2). Such a difference may partly be due to the difference in the length of the seasons (about two month in summer compared to the rest of the year). As the fall

observation was missing in 2007 and 2009 in Yamal, only spring was analysed for this region. Some variation in occurrence was also observed between years, but given the short duration of the observations from Russia we will not make inferences about multi-annual population dynamics (see Henden et al. submitted for ptarmigan in Finnmark).

Landscape scale habitat use

For ptarmigan occurrence at the large scale, the best model included an interaction between region and the two habitat variables C-land and habitat, as well as between year and season. Willow cover at the scale of units had no significant effect on the overall occurrence of ptarmigan in the Russian regions. In Finnmark on the contrary where willow cover was on average lowest (Table 1), occurrence increased with C-land (logit estimate = 6.29, standard error (SE) = 2.06; ESM Fig. S2). Considering habitat, in Yamal occurrence was nearly twice as low on tundra plots as on W plots (logit estimate for T plots with W plots as reference = -0.56, SE = 0.26, odds ratio = 0.57). In Finnmark, the preference of ptarmigan for W plots was even stronger (logit estimate for T plots = -1.83, SE = 0.30, odds ratio = 0.16), whereas in Nenetsky, where willow cover was on average highest, ptarmigan clearly preferred tundra plots (logit estimate for T plots = 0.96, SE = 0.17, odds ratio = 2.62). In addition, occurrence was significantly lower in fall than in spring, an effect which was strongest in 2008 (see ESM Table S2 for complete model output). Chisquare tests showed that in Nenetsky overall ptarmigan occurrence did not differ between the two tundra types ($X^2 = 0.11$, df = 2, P = 0.74), whereas in Yamal ptarmigan clearly avoided M plots ($X^2 = 17.86$, df = 2, P < 0.001)

323

324

325

326

327

322

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

The best model for hare occurrence in the Russian regions at large scale included interactions of the two habitat variables with year and the interaction between year and season. There was a general positive effect of willow cover on occurrence per unit (logit estimate = 1.66, SE = 0.31; ESM Fig. S2). There was no consistent difference between the

two habitat types over the years, but more hare faeces were found on W plots in 2009 in both regions (see ESM Table S3 for complete model output). As for ptarmigan, occurrence was lower in fall than in spring. The difference between the seasons was smaller in 2007 than in subsequent years (ESM Table S3). Considering tundra types, hares clearly avoided M plots in Yamal ($X^2 = 35.64$, df = 2, P < 0.001), whereas in Nenetsky they avoided S plots ($X^2 = 13.41$, df = 2, P < 0.001). These habitat preferences were consistent over seasons (Fig. 2).

Local scale habitat use

Considering only willow plots and their direct surroundings (200 x 200 m), the best model for ptarmigan occurrence included interactions of the willow configuration variables with region. For Finnmark, we observed a positive effect of C-loc and a negative effect of thicket fragmentation, as reported by Henden et al. (submitted). In addition, there was a negative effect of W-height, which was not significant in the previous analysis. The estimates of the effect of W-height were, however, not very different between the two analyses, which included a different set of years and explanatory variables. The variables used were strongly correlated, reflecting the same pattern of willow configuration, but the exact choice of variables to include can modify the estimates of the other effects (Smith et al. 2009). In Nenetsky on the contrary there was no effect of any of the willow configuration variables on the occurrence of ptarmigan and the contrasts in slope with Finnmark as reference level were significant (Fig. 3; ESM Table S4). In Yamal, the effects of willow configuration were not different from those observed in Finnmark.

For hare, the best model at the local scale included interactions of the willow configuration variables with season. As fall counts were missing in Yamal, these interactions could be estimated only for Nenetsky. In Nenetsky, PD had a negative effect on hare occurrence in

spring (logit estimate = -1.37, SE = 0.37), but not in fall (Fig. 4). There was also a significant contrast in the effect of W-height, which was slightly negative in spring but positive in fall (Fig. 4, ESM Table S5). Considering only spring counts from both regions produced consistent results and revealed a similar negative effect of PD, indicating that in winter hares prefer less fragmented willow thickets.

Discussion

354

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

370

371

372

373

374

375

376

377

For ptarmigan, our analyses revealed that the importance of willow thickets for regionspecific abundance and habitat use decreased with increasing willow cover both at the scale of units and at the local scale of W plots. In each region, the effects were surprisingly consistent over years and seasons. In eastern Finnmark, where willow thickets occupy only a very small proportion of the landscape, are restricted to rather narrow riparian plains and may be additionally fragmented by intense reindeer browsing (Henden et al. 2010), there was a significant positive relationship between willow cover at the large scale and the occurrence of ptarmigan. This result was in clear contrast to the Russian regions, where willow cover was higher but did not relate to ptarmigan occurrence. At the same time the preference of ptarmigan for W plots was highest in eastern Finnmark, whereas in Nenetsky, where willow cover was highest and thickets occur also on the plateau between the valleys, ptarmigan preferred tundra plots. In Yamal willow cover was intermediate and ptarmigan preferred W plots, but not as strongly as in eastern Finnmark. At the local scale of W plots and their direct surroundings, our results show that the positive effect of willow cover and negative effect of fragmentation reported by Henden et al. (submitted) for Finnmark was also observed in Yamal. In Nenetsky on the contrary, where W plots were not the preferred habitat, ptarmigan did not select plots with relatively higher willow cover and less fragmented thickets. This may be explained by the general abundance of willows

in Nenetsky, but also by the fact that the range of C-loc and PD did not include equally low values in Nenetsky as in the other regions (Table 1).

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

398

399

400

401

402

403

378

379

Altogether we thus documented a functional response in habitat choice (Mysterud and Ims 1998) for ptarmigan with respect to willow thickets – an important resource both as food and as shelter. Ptarmigan, which are characteristic medium sized herbivores of the typically patchy shrub tundra, show increasing preference for willow thicket edge habitats when the amount of willow thickets on the regional level decreases. Furthermore, within regions with low amount of willows (such as in eastern Finnmark) ptarmigan prefer local areas or landscape sections with a maximum amount of less fragmented willow thickets. Indeed, the contrasting results between Finnmark and the two Russian sites indicate that willow thickets are a strongly limiting resource for ptarmigan in Finnmark. Whether the willows are most important in terms of forage or protective cover is unknown. However, cover may be particularly important in presence of specialized avian predators such as gyr falcons (Falco rusticolus, specialized on ptarmigan; Nyström et al. 2005) and golden eagle (Aquila chrysaetos, specialized on both ptarmigan and hares; Johnsen et al. 2007; Nyström et al. 2006). Both of these raptors are quite common year round residents in Finnmark, while they are not breeding in the two Russian regions. In the Russian regions the main avian predators of ptarmigan are rough-legged buzzard (Buteo lagopus) and peregrine falcon (Falco peregrinus; Osmolovskaya 1948), which are both absent in winter. In addition, snowy owls (Bubo scandiaca) are present in all three regions in winter. The fact that ptarmigan in Finnmark equally strongly selected willow thicket edges in summer and in winter, although willow twigs are most foraged in winter (Tape et al. 2010), underline the importance of willow thickets as cover for ptarmigan in this region. The preference for willow habitats in Yamal in winter was not as strong as in Finnmark and was in fact due to avoidance of the flat and Sphagnum dominated M plots. D plots and W plots were used

equally in this region, as could be expected if ptarmigan experienced lower predation pressure and thus less need for cover. The preference for tundra habitats in Nenetsky was also consistent with a reduced need for cover, and may be additionally explained by topography and landscape characteristics. Willow thickets in the deep valleys in Nenetsky are likely to be covered by thick layers of snow in winter making them less accessible for foraging than thickets on the plateau. In summer, the lush and high meadow vegetation on W plots (Skogstad 2009) may be little suitable for ptarmigan.

In Finnmark ptarmigan preferred lower willows, but this was not the case in Nenetsky. Yamal, where willow thickets were on average lowest, was not significantly different from Finnmark in this respect, although the parameter estimate was similar to that from Nenetsky (Fig. 3), not indicating any preference for low willows. The different effect of willow height is likely to be due to different willow architecture. In Finnmark W-height and W-density were not correlated, and some of the higher shrubs had little lower branches where ptarmigan could feed. In the Russian regions willow shrubs were more dense and likely to offer equal feeding opportunities at different heights.

As overall level of occurrence of ptarmigan was highest in Nenetsky and lower in eastern Finnmark, an alternative explanation the region-specific use of tundra habitat could be density-dependent habitat selection; i.e. that use of tundra habitats increase when the regional abundance increase due to competition for optimal habitats (e.g. Fretwell 1972). However, habitat use did not differ between the seasons although the level and mode of competition between ptarmigan is expected shift between seasons as they are territorial in summer but gather in flocks in winter (Storch 2007). Thus, we consider this explanation unlikely.

For hare, region-specific habitat selection could only be analysed in the two Russian regions where the contrast in willow cover was less. Nevertheless, our results showed that hare occurrence at the landscape scale was higher where willow cover was higher, and this effect was not different between the regions. This consistent positive effect of willow cover was in contrast to the functional response observed in ptarmigan and may be related to the distribution of hares, which extends far into the boreal and temperate zone (Kolosov et al. 1965). As an animal also living in forests, hares may be less dependent on the characteristic patchy structure of shrub tundra than ptarmigan. At the same time there was no clear preference of hares for a particular habitat. Occurrence was higher on W plots in 2009, but not in the two other years. At the local scale, hares preferred less fragmented thickets in winter, but this was not the case in summer. In winter, larger thickets may provide better foraging opportunities and protection. In summer on the contrary hares feed mostly in open habitats such as meadows (Labutin 1988; Pavlinin 1997), which are likely to be most accessible in a landscape with smaller willow patches.

In addition to differences in region-specific habitat selection, our data indicated differences in regional abundance of ptarmigan and hare. The overall occurrence of ptarmigan was highest in Nenetsky and lower in Yamal and eastern Finnmark, whereas the occurrence of hares was slightly higher in Nenetsky than in Yamal, and very low in Finnmark (Fig. 2). As sampling was stratified to include main habitat types in each region, overall occurrence reflects regional abundance. When discussing ptarmigan abundance, the multi-annual population dynamics of the species should be considered (Storch 2007). In Finnmark ptarmigan numbers have consistently decreased over the last years, a decline which can neither be explained directly by willow thicket degradation or by a predator mediated effect of small rodent dynamics (Henden et al. submitted). In Yamal, the years of our study were years of low ptarmigan abuncance (V. A. Sokolov, unpublished), whereas the

dynamics in Nenetsky are unknown. Nevertheless, overall regional abundance of both
species seemed positively correlated with the amount of willow thickets, suggesting that
ptarmigan and hare could be limited by the availability of thickets in Finnmark. Several
non-exclusive hypotheses may explain the observed differences in regional abundance.
Assuming that willow thicket growth in Finnmark is affected by intense reindeer browsing
(Den Herder et al. 2004, 2008; Kitti et al. 2009), a negative impact of reindeer numbers on
medium sized herbivores could be hypothesized (Ims et al. 2007). This interpretation,
involving a trophic bottom-up effect, should however be completed by considerations of
the predator community, which also differs between the regions. In addition to the
presence of avian predators during winter (see above), red foxes (Vulpes vulpes) – an
important predator for both hare and ptarmigan – are more abundant in Finnmark than in
the Russian regions, where arctic foxes (V. lagopus) dominate (Killengreen et al. 2007;
Arctic Predators project, unpublished). Raven (Corvus corax) and crow (C. cornix), two
generalist predators which have been shown to have a negative impact on rock ptarmigans
in Scottland (Watson and Moss 2004), are also considerably more abundant in Finnmark
than in the Russian regions (Killengreen 2010; Arctic Predators project, unpublished).
Total predation pressure is thus likely to be higher in Finnmark and may contribute to
lower abundance. Our data do, however, not allow us to present more than suggestive
correlations concerning regional abundance as many factors such as multi-annual
population dynamics (Storch 2007; Newey et al. 2007) or the influence of hunting, which
is likely to be stronger in Norway than in Russia, were not considered.

Conclusions

Investigating the importance of willow thickets for two medium sized herbivores in three different shrub tundra regions revealed clear differences in region specific abundance and habitat selection. We document a functional response in the use of willow habitats by

ptarmigan, but not for hare. Region specific relationships reflecting the regional particularities of the landscape and ecosystem highlight the importance of large scale investigations to understand the relationships of habitat availability and use, as for many other questions in ecology. Under climate change willow shrubs are likely to expand in the arctic tundra, a process which may be limited by browsing of large herbivores (Post and Pedersen 2008). Understanding how medium sized herbivores may react to changes in willow cover and thicket configuration will add an important element to predictions of how the arctic tundra ecosystem may change in the near future. Willow thickets are, however, only one component of the ecosystem influencing abundance and habitat selection of ptarmigan and hare. A complete understanding of the changes in the position of these herbivores in the tundra ecosystem will require the integration of other factors, such as predation (Lima and Dill 1990).

Acknowledgements

We are grateful to Eeva Soininen, Ingrid Jensvoll, Anna Rodnikova, Victor Sidorov, Olga Kulikova and many others for great contributions to the field work. This study was financed by the Research Council of Norway through the projects "EcoFinn" and "IPY-Arctic Predators".

199	neielelices
500	Andreev A (1988) The 10 Year Cycle of the Willow Grouse of Lower Kolyma. Oecologia 76:261-267
501	Chernov YI, Matveeva NV (1997) Arctic ecosystems in Russia. In: Wielgolaski FE (ed) Ecosystems of the
502	World. Volume 3: Polar and Alpine Tundra. Elsevier, Oxford, pp 361-507
503	Bates D, Maechler M, Dai B (2008) lme4: Linear mixed-effects models using S4 classes [Computer
504	software]. Available at http://cran.r-project.org/
505	Den Herder M, Virtanen R, Roininen H (2004) Effects of reindeer browsing on tundra willow and its
506	associated insect herbivores. J Appl Ecol 41:870-879
507	Den Herder M, Virtanen R, Roininen H (2008) Reindeer herbivory reduces willow growth and grouse forage
508	in a forest-tundra ecotone. Basic Appl Ecol 9:324-331
509	Eigenbrod F, Hecnar SJ, Fahrig L (2010) Sub-optimal design has major impacts on landscape-scale
510	inference. Biol Conserv doi:10.1016/j.biocon.2010.09.007
511	Elson LT, Schwab FE, Simon NPP (2007) Winter food habits of Lagopus (Willow ptarmigan) as a
512	mechanism to explain winter sexual segregation. Northeastern Nat 14:89-98
513	Estaf'ev AA, Mineev YN (1984) Seasonal distribution of willow ptarmigan in tundra of European north-east
514	of USSR. In: Animals – components of the ecosystem of the European North and Ural. University of
515	Syktyvkar, Russia (in Russian).
516	Evans SA, Mougeot F, Redpath SM, Leckie F (2007) Alternative methods for estimating density in an upland
517	game bird: the red grouse Lagopus lagopus scotieus. Wildl Biol 13:130-139
518	Forbes BC, Stammler F, Kumpula T, Meschtyb N, Pajunen A, Kaarlejarvi E (2009) High resilience in the
519	Yamal-Nenets social-ecological system, West Siberian Arctic, Russia. Proc Natl Acad Sci USA 106:22041-
520	22048
521	Fortin D, Courtois R, Etcheverry P, Dussault C, Gingras A (2008) Winter selection of landscapes by
522	woodland caribou: behavioural response to geographical gradients in habitat attributes. J Appl Ecol 45:1392-
523	1400
524	Fretwell SD (1972) Populations in a Seasonal Environment. Princeton University Press, Princeton, USA
525	Gelman A, Hill J (2007) Data analysis using regression and multilevel/hierarchical models. Cambridge
526	University Press, Cambridge, UK.
527	Hakkarainen H, Virtanen R, Honkanen JO, Roininen H (2007) Willow bud and shoot foraging by ptarmigan
528	in relation to snow level in NW Finnish Lapland, Polar Biol 30:619-624

529	Henden JA, Ims RA, Yoccoz NG, Sørensen R, Killengreen ST (2010) Population dynamics of tundra voles in
530	relation to configuration of willow thickets in southern arctic tundra. Polar Biol DOI 10.1007/s00300-010-
531	0908-7
532	Ims RA, Yoccoz NG, Brathen KA, Fauchald P, Tveraa T, Hausner V (2007) Can reindeer overabundance
533	cause a trophic cascade? Ecosystems 10:607-622
534	Johnsen TV, Systad GH, Jacobsen KO, Nygard T, Bustnes JO (2007) The occurrence of reindeer calves in
535	the diet of nesting Golden Eagles in Finnmark, northern Norway. Ornis Fenn 84:112-118
536	Johnson DH (1980) The comparison of usage and availability measurements for evaluating resource
537	preference. Ecology 61:65-71
538	Killengreen ST (2010) Sources of spatial variation in food web structure in low Arctic tundra. PhD
539	dissertation, University of Tromsø, Tromsø, Norway
540	Killengreen ST, Ims RA, Yoccoz NG, Brathen KA, Henden JA, Schott T (2007) Structural characteristics of
541	a low Arctic tundra ecosystem and the retreat of the Arctic fox. Biol Conserv 135:459-472
542	Kitti H, Forbes BC, Oksanen J (2009) Long- and short-term effects of reindeer grazing on tundra wetland
543	vegetation. Polar Biol 32:253-261
544	Kolosov AM, Lavrov NP, Naumov SP (1965) Mountain hare – Lepus timidus L. In: Biology of hunted
545	animals of the USSR. Vyshaya Shkola, Moskva, pp 328-335 (in Russian)
546	Krebs CJ, Boonstra R, Nams V, O'Donoghue M, Hodges KE, Boutin S (2001) Estimating snowshoe hare
547	population density from pellet plots: a further evaluation. Can J Zool 79:1-4
548	Labutin YV (1988) Particularities of the spatial distribution and behaviour of mountain hares (<i>Lepus timidus</i>)
549	in Yakutia as adaptations of the species to northern conditions. Russ J Ecol 2:40-44 (in Russian)
550	Mayor SJ, Schneider DC, Schaefer JA, Mahoney SP (2009) Habitat selection at multiple scales. Ecoscience
551	16:238-247
552	Lima SL, Dill LM (1990) Behavioral decisions made under the risk of predation – a review and prospectus.
553	Can J Zool 68:619-640
554	McGarigal K, Cushman SA, Neel MC, Ene E (2002) FRAGSTATS: Spatial Pattern Analysis Program for
555	Categorical Maps. Computer software program produced by the authors at the University of Massachusetts,
556	Amherst, USA
557	Meffe GK, Carroll CR (1994) Principles of Conservation Biology. Sinauer Associates, Inc. Sunderland,
558	Massachusetts, USA
559	Moen A (1998) Nasjonalatlas for Norge. Vegetasjon. Statens kartverk, Hønefoss, Norway

560	Mysterud A, Ims RA (1998) Functional responses in habitat use: Availability influences relative use in trade-
561	off situations. Ecology 79:1435-1441
562	Mysterud A, Larsen PK, Ims RA, Ostbye E (1999) Habitat selection by roe deer and sheep: does habitat
563	ranking reflect resource availability? Can J Zool 77:776-783
564	Newey S, Willebrand T, Haydon DT, Dahl F, Aebischer NJ, Smith AA, Thirgood SJ (2007) Do mountain
565	hare populations cycle? Oikos 116:1547-1557
566	Nystrom J, Ekenstedt J, Angerbjorn A, Thulin L, Hellstrom P, Dalen L (2006) Golden Eagles on the Swedish
567	mountain tundra - diet and breeding success in relation to prey fluctuations. Ornis Fenn 83:145-152
568	Nystrom J, Ekenstedt J, Engstrom J, Angerbjorn A (2005) Gyr Falcons, ptarmigan and microtine rodents in
569	northern Sweden. Ibis 147:587-597
570	Orians GH, Wittenberger JF (1991) Spatial and temporal scales in habitat selection. Am Nat 137:29-49
571	Osmolovskaya VI (1948) Ecology of raptors on the Yamal Peninsula. Trudy Instituta Geografii, 41:5-77 (in
572	Russian)
573	Pajunen AM (2009) Environmental and Biotic Determinants of Growth and Height of Arctic Willow Shrubs
574	along a Latitudinal Gradient. Arct Antarct Alp Res 41:478-485
575	Pajunen AM, Kaarlejarvi EM, Forbes BC, Virtanen R (2010) Compositional differentiation, vegetation-
576	environment relationships and classification of willow-characterised vegetation in the western Eurasian
577	Arctic. J Veg Sci 21:107-119
578	Pavlinin VV (1997) Particularities of the ecology of mountain hare (<i>Lepus timidus</i>) on Yamal. In: Materialy
579	po istorii i sovremennomu sostoyaniyu fauni severa Zapadnoy Sibiri. pp31-42.(In Russian)
580	Pavlov AV, Moskalenko NG (2002) The thermal regime of soils in the north of Western Siberia. Permafr
581	Periglac Process 13:43-51
582	Post E, Pedersen C (2008) Opposing plant community responses to warming with and without herbivores.
583	Proc Natl Acad Sci USA 105:12353-12358
584	R Development Core Team (2010) R: A Language and Environment for Statistical Computing. R Foundation
585	for Statistical Computing, Vienna, Austria
586	Ravolainen VT, Yoccoz NG, Brathen KA, Ims RA, Iversen M, Gonzalez VT. 2010. Additive Partitioning of
587	Diversity Reveals No Scale-dependent Impacts of Large Ungulates on the Structure of Tundra Plant
588	Communities. Ecosystems 13:157-170
589	Ripple WJ, Beschta RL (2005) Refugia from browsing as reference sites for restoration planning. West N
590	Am Nat 65:269-273

621

Scotland. Biol Conserv 116:267-275

591 Shiatov SG, Mazepa VS (1995) Climate. In: Dobrinskiy LN (ed) The Nature of Yamal. Nauka, Ekaterinburg, 592 pp 32-67 (in Russian) 593 Shtro VG (2006) Notes about the behaviour of mountain hares in the tundra of Yamal. Nauchniy Vestnik 594 YNAO 1(38):173-174 (in Russian) 595 Skogstad G (2009) Does Habitat Fragmentation of Willow Thickets have Spill-Over Effects on Tundra 596 Vegetation? Master Thesis, University of Tromsø, Tromsø, Norway 597 Smith AC, Koper N, Francis CM, Fahrig L (2009) Confronting collinearity: comparing methods for 598 disentangling the effects of habitat loss and fragmentation. Landsc Ecol 24:1271-1285 599 Storch I (2007) Grouse. Status and Conservation Action Plan 2006-2010. IUCN Species Survival 600 Commission. 601 Sturm M, Racine C, Tape K (2001) Climate change - Increasing shrub abundance in the Arctic. Nature 602 411:546-547 603 Tape K, Sturm M, Racine C (2006) The evidence for shrub expansion in Northern Alaska and the Pan-Arctic. 604 Glob Change Biol 12:686-702 605 Tape KD, Lord R, Marshall HP, Ruess RW (2010) Snow-mediated ptarmigan browsing and shrub expansion 606 in arctic Alaska. Ecoscience 17:186-193 607 Van Erden MRE (2000) Pechora Delta: Structure and dynamics of the Pechora Delta ecosystems (1995-608 1999). Leylstad, Naryan-Mar, Syktyvkar 609 Virtanen R, Oksanen L, Razzhivin VY (1999) Topographical and regional patterns of tundra heath vegetation 610 from northern Fennoscandia to the Taimyr peninsula. Acta Bot Fenn 167:29-83 611 Walker DA, Leibman MO, Epstein HE, Forbes BC, Bhatt US, Raynolds MK, Comiso JC, Gubarkov AA, 612 Khomutov AV, Jia GJ, Kaarlejarvi E, Kaplan JO, Kumpula T, Kuss P, Matyshak G, Moskalenko NG, 613 Orekhov P, Romanovsky VE, Ukraintseva NG, Yu Q (2009) Spatial and temporal patterns of greenness on 614 the Yamal Peninsula, Russia: interactions of ecological and social factors affecting the Arctic normalized 615 difference vegetation index. Environ Res Lett 4 616 Walker DA, Raynolds MK, Daniels FJA, Einarsson E, Elvebakk A, Gould WA, Katenin AE, Kholod SS, 617 Markon CJ, Melnikov ES, Moskalenko NG, Talbot SS, Yurtsev BA (2005) The Circumpolar Arctic 618 vegetation map. J Veg Sci 16:267-282 619 Watson A, Moss R (2004) Impacts of ski-development on ptarmigan (*Lagopus mutus*) at Cairn Gorm,

West GC, Meng MS (1966) Nutrition of willow ptarmigan in northern Alaska. Auk 83:603-615

- Wookey PA et al. (2009) Ecosystem feedbacks and cascade processes: understanding their role in the
- responses of Arctic and alpine ecosystems to environmental change. Glob Change Biol 15:1153-1172



Table 1. Willow thicket configuration variables presented as means and ranges (in brackets) for the three study regions: number of plots (n), percent willow cover estimated on plots of 2x2 km (C-land), percent willow cover estimated on plots of 200x200 m (C-loc), patch density (PD) and edge density (ED), both measured on plots of 200x200 m.

	Finnmark	Nenetksy	Yamal
n	37	12	12
C-land (% area)	1.8 [0.6-3.5]	23.6 [17.9-30.5]	11.1 [1.2-22.7]
C-loc (% area)	19.8 [1.4-54.3]	35.0 [13.1-52.6]	12.3 [1.5-31.9]
PD (nb patches / 4 ha)	18.5 [1-87]	29.5 [10-69]	8.5 [3.0-16.1]
ED (m edge / 4 ha	1760 [367-4036]	2933 [1120-4908]	1021 [293-1904]
W-height (m)	1.61 [0.78-2.70]	1.70 [1.05-2.38]	0.82 [0.53-1.24]
W-density (nb of hits)	2.5 [0.3 – 5.5]	6.4 [2.3-9.5]	3.6 [2.0-6.5]

Figure legends

628

629

630

631

632

633

634

635

636

637

638

639

640

641

642

643

644

645

646

Fig. 1. Map of the study regions relative to the Bioclimatic subzones defined by Walker et al. (2005). In Finnmark, IF referes to Ifjordfjellet, VJ to Vestre Jakobselv and KO to Komag. The inserts show an overview of the study design: a) Three units in three small valleys in Nenetsky. b) Triplets of plots representing each of three habitat types were chosen. c) The contour of willow thickets was drawn on satellite images or aerial photographs in order to estimate their surface. d) Each plot comprised eight small quadrates where faeces were counted. Willow (W) plots were in meadows at the edge of willow thickets. Fig. 2. Occurrence of faeces of willow ptarmigan and mountain hare in the three study regions is plotted for each habitat type. W - meadow plots at the edge of willow thickets, T - tundra plots in eastern Finnmark, S shrubby tundra, H – hummock tundra, D – dry tundra and M – moist tundra (see main text for a description of the habitat types). Occurrence refers to the number of small quadrates surrounding a plot where faeces were recorded. For each year, spring and fall counts are shown. Fig. 3. Local scale: ptarmigan occurrence in willow plots in the three study regions in spring 2008 as a function of willow cover and patch density in the 4 ha surrounding each plot and as a function of willow height in the plots. Points show values predicted from the selected model, and lines show relationships given average values for the other predictor variables. Fig. 4. Hare occurrence in function of willow patch density in the 4 ha surrounding each plot at the edge of a willow thicket, and in function of willow height. Points show values predicted from the selected model, and lines show relationships given average values for the other predictor variables.

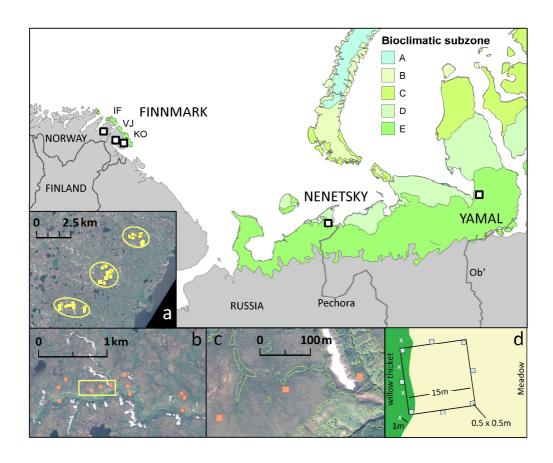


Figure 1

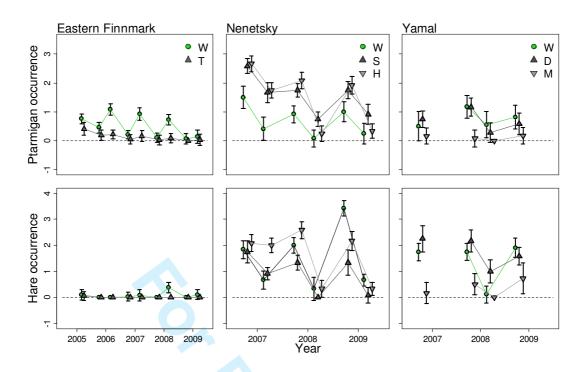


Figure 2

Page 31 of 39 Oecologia

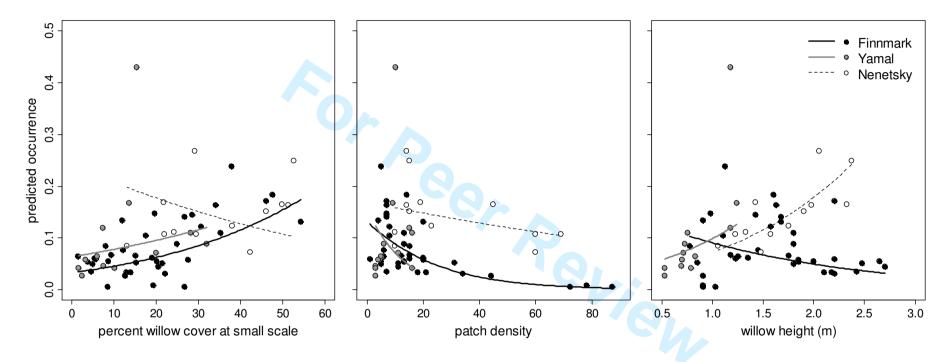


Figure 3

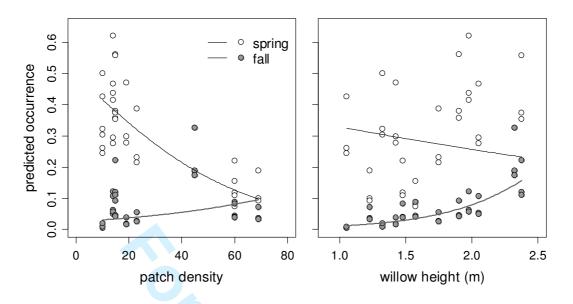


Figure 4

Online supplementary Information to the paper *The importance of willow thickets for ptarmigan and hares in shrub tundra: the more the better?* By D Ehrich, JA Henden, RA Ims, L Doronina, ST Killengreen, N Lecomte, IG Pokrovsky, G Skogstad, AA Sokolov, VA Sokolov, NG Yoccoz.

Table S1

Correlation coefficients among the willow thicket configuration variables: C-land is the percent cover of willow thickets estimated on squares of 2x2 km, C-loc is the percent cover estimated on squares of 200x200m, PD is patch density estimated as the number of patches per 4 ha and ED is the total length of thicket edge (m) per 4 ha. Correlations are shown for the total data set as well as for each region separately. Correlation coefficients of 0.5 or more are highlighted in bold.

Total dataset

	C-loc	C-land	PD	ED	W-height
C-land	0.42				
PD	0.17	0.14			
ED	0.64	0.38	0.77		
W-height	0.36	-0.02	-0.09	0.15	
W-hits	0.37	0.72	0.29	0.48	0.12

Eastern Finnmark

	C-loc	C-land	PD	ED	W-height
C-land	0.52				
PD	-0.02	-0.15			
ED	0.43	0.04	0.76		
W-height	0.09	0.03	-0.40	-0.31	
W-hits	-0.02	0.01	0.23	0.20	0.02

Nenetsky

renetsky						
	C-loc	C-land	PD	ED	W-height	
C-land	0.24					
PD	0.00	-0.19				
ED	0.46	0.18	0.79			
W-height	0.80	0.29	-0.19	0.21		
W-hits	0.38	0.16	0.18	0.42	0.40	

Yamal

	C-loc	C-land	PD	ED	W-height
C-land	0.57				
PD	0.44	0.23			
ED	0.93	0.50	0.67		
W-height	0.15	0.10	0.49	0.39	
W-hits	0.37	0.40	0.69	0.62	0.77

Oecologia Page 34 of 39

Table S2

Results of the selected general linear mixed effects model (GLMM) for ptarmigan occurrence at the large scale in function of willow cover (C-land), habitat, region, year and season. Significant effects (P < 0.05) are highlighted in bold. Estimates represent effects on the logit scale. Estimates of contrasts and interaction terms should be added to other contrasts for interpretation. Thus Year 2008 represents the difference between Year 2007 and Year 2008 and Habitat Tundra the difference between Habitat Willow and Habitat Tundra. The effect of C-land for occurrence in Tundra plots in Nenetsky, Spring 2007 is for example -2.01 + 1.02 - 0.56 - 0.54 – 0.24 + 1.52 = -0.81, indicating a small negative effect.

Random effect: unit, Standard deviation = 3.38×10^{-7} (Number of observations: 176; number of groups: 15)

Fixed effects:

Reference levels for factors were habitat Willow, region Yamal, year 2007and season Spring.

Effect	Estimate	Standard Error	Р
Intercept	-2.01	0.22	< 0.001
C-land	1.02	0.57	0.074
Habitat Tundra	-0.56	0.26	0.034
Region Nenetsky	-0.54	1.34	0.687
Region Finnmark	2.00	0.78	0.010
Year 2008	-0.22	0.14	0.123
Year 2009	-0.59	0.16	< 0.001
Season fall	-0.84	0.18	< 0.001
C-land x Region Nenetsky	-0.24	1.92	0.899
C-land x Region Finnmark	5.26	2.14	0.014
Habitat Tundra x region Nenetsky	1.52	0.31	< 0.001
Habitat Tundra x region Finnmark	-1.28	0.40	0.001
Year 2008 x season fall	-1.02	0.31	0.001
Year 2009 x season fall	-0.46	0.31	0.138

Table S3

Results of the selected GLMM for hare occurrence at the large scale in function of willow cover (C-land), habitat, region, year and season. Significant effects (P < 0.05) are highlighted in bold. Estimates represent effects on the logit scale. See legend to table 2 for interpretation of the estimates.

Random effect: unit, Standard deviation = 5.12×10^{-7} (Number of observations: 72; number of groups: 5)

Fixed effects:

Reference levels for factors were habitat Willow, region Yamal, year 2007and season Spring.

Effect	Estimate	Standard Error	Р
Intercept	-0.41	0.25	0.10
C-land	1.66	0.31	< 0.001
Habitat Tundra	0.09	0.19	0.617
Year 2008	0.28	0.24	0.240
Year 2009	0.81	0.23	< 0.001
Region Nenetsky	-1.30	0.27	< 0.001
Season fall	-0.54	0.21	0.011
C-land x Year 2008	0.08	0.29	0.778
C-land x Year 2009	0.40	0.30	0.186
Habitat Tundra x Year 2008	-0.31	0.28	0.266
Habitat Tundra x Year 2009	0.96	0.27	< 0.001
Year 2008 x season fall	1.92	0.44	< 0.001
Year 2009 x season fall	-1.67	0.38	< 0.001

Table S4

Results of the selected GLMM for ptarmigan occurrence at the local scale in function of willow cover (C-loc), patch density (PD), willow height (W-height), willow density (W-density), region, year and season. Significant effects (P < 0.05) are highlighted in bold. Estimates represent effects on the logit scale. See legend to table 2 for interpretation of the estimates.

Random effects: plot, Standard deviation = 0.149 (Number of observations: 311; number of groups: 61)

Fixed effects:

Reference levels for factors were region Finnmark, year 2007and season Spring.

Effect	Estimate	Standard Error	Р
Intercept	-2.29	0.22	< 0.001
C-loc	0.98	0.25	< 0.001
PD	-1.79	0.54	0.001
W-height	-0.71	0.29	0.015
W-density	0.41	0.46	0.373
Region Nenetsky	0.54	0.42	0.191
Region Yamal	0.13	1.29	0.919
Year 2008	-0.26	0.19	0.181
Year 2009	-0.78	0.23	0.001
Season Fall	-1.66	0.24	< 0.001
C-loc x Region Nenetsky	-1.55	0.66	0.019
C-loc x Region Yamal	-0.34	0.85	0.685
PD x Region Nenetsky	1.47	0.64	0.022
PD x Region Yamal	-1.66	2.58	0.519
W-height x Region Nenetsky	1.84	0.82	0.025
W-height x Region Yamal	1.99	1.54	0.120
W-density x Region Nenetsky	0.11	0.63	0.855
W-density x Region Yamal	2.92	1.20	0.057

Table S5

Results of the selected GLMM for hare occurrence at the local scale in Nenetsky in function of willow cover (C-loc), patch density (PD), willow height (W-height), willow density (W-density), year and season. Significant effects (P < 0.05) are highlighted in bold. Estimates represent effects on the logit scale. See legend to table 2 for interpretation of the estimates.

Random effects: plot, Standard deviation = 0.207 (Number of observations: 72; number of groups: 12)

Fixed effects:

Reference levels for factors were year 2007and season Spring.

Effect	Estimate	Standard Error	Р
Intercept	-1.18	0.23	< 0.001
C-loc	0.57	0.53	0.286
PD	-1.37	0.37	< 0.001
W-height	-0.29	0.58	0.619
W-density	0.23	0.36	0.519
Season Fall	-2.10	0.36	< 0.001
Year 2008	-0.9	0.28	0.759
Year 2009	0.74	0.99	0.009
C-loc x season Late	-0.34	0.68	0.728
PD x season Late	2.25	0.99	0.001
W-height x season Late	1.98	0.76	0.045
W-density x season Late	0.01	0.01	0.99

Figure S1

Principal components analysis of the willow thicket configuration variables: percent willow cover at large scale (2 x 2 km; C.land), percent willow cover at the local scale (200 x 200 m; C.loc), patch density (PD), edge density (ED), willow height (W.height) and willow density (W.density). On the left plot, arrows close to each other represent closely correlated variables. On the right plot all W plots are represented by a black dot and grouped in order to display the variation among the three regions.

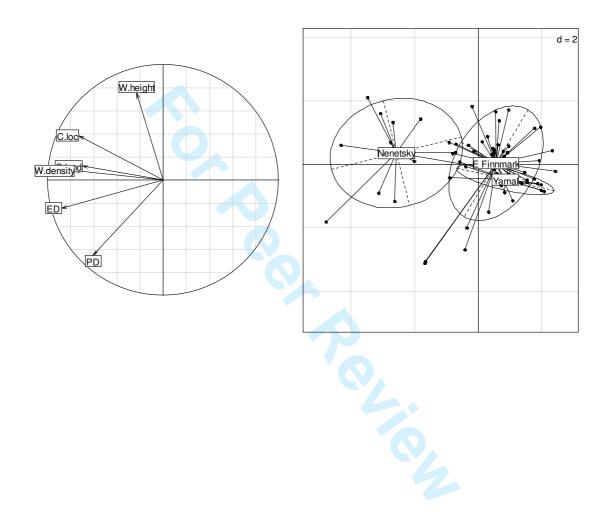


Figure S2

Effect of willow cover at the landscape scale (units) on the occurrence of ptarmigan and hare in spring. Circles and lines indicate willow plots whereas triangles and dashed lines show tundra plots. For ptarmigan, the slope of occurrence with willow cover was different in the three regions, which are plotted in black (Finnmark), grey (Yamal) and white (Nenetsky) respectively. For hare the best model included a non significant interaction of willow cover with year. White symbols and thin lines represent 2007, light grey symbols and line 2008 and dark grey symbols and lines 2009.

