

1 **Assessing the relationship between perfluoroalkyl substances, thyroid hormones and**
2 **binding proteins in pregnant women; a longitudinal mixed effects approach**

3

4 Vivian Berg^{abc}

5 Therese Haugdahl Nøst^{abc}

6 Solrunn Hansen^c

7 Astrid Elverland^a

8 Anna-Sofía Veyhe^c

9 Rolf Jorde^d

10 Jon Øyvind Odland^c

11 Torkjel Manning Sandanger^{bc}

12

13 ^aDepartment of Laboratory Medicine, Diagnostic Clinic, University Hospital of Northern
14 Norway, Sykehusveien 38, NO-9038 Tromsø, Norway;

15 ^bDepartment of Environmental Chemistry, NILU- Norwegian Institute of Air Research, Fram
16 Centre, Hjalmar Johansens gate 14, NO-9296 Tromsø, Norway;

17 ^cDepartment of Community Medicine, Faculty of Health Sciences, University of Tromsø-The
18 Arctic University of Norway, Hansine Hansens veg 18, NO-9019 Tromsø, Norway;

19 ^dEndocrine Research Group, Institute of Clinical Medicine, University of Tromsø-The Arctic
20 University of Norway, Hansine Hansens veg 18, NO-9019 Tromsø, Norway

21

22 **Corresponding author:** Vivian Berg, Department of Laboratory Medicine, Diagnostic Clinic,
23 University Hospital of Northern Norway, Sykehusveien 38, NO-9038 Tromsø, Norway.
24 Tel.(+47)77750393. vivian.berg@uit.no

25

26 **Abstract**

27 The mechanisms involved in thyroid homeostasis are complex, and perfluoroalkyl substances
28 (PFASs) have been indicated to interfere at several levels in this endocrine system. Disruption
29 of the maternal thyroid homeostasis during early pregnancy is of particular concern, where
30 subclinical changes in maternal thyroid hormones (THs) may affect embryonic and foetal
31 development.

32 The present study investigated associations between THs, thyroid binding proteins (TH-BPs)
33 and PFAS concentrations in pregnant women from Northern Norway.

34 Women participating in The Northern Norway Mother-and-Child contaminant Cohort Study
35 (MISA) donated a blood sample at three visits related to their pregnancy and postpartum period
36 (during the second trimester, 3 days and 6 weeks after delivery) in the period 2007-2009.
37 Participants were assigned to quartiles according to PFAS concentrations during the second
38 trimester and mixed effects linear models were used to investigate potential associations
39 between PFASs and repeated measurements of THs, TH-BPs, thyroxin binding capacity and
40 thyroid peroxidase antibodies (anti-TPO).

41 Women within the highest perfluorooctane sulfonate (PFOS) quartile had 24% higher mean
42 concentrations of thyroid stimulating hormone (TSH) compared to the first quartile at all
43 sampling points. Women within the highest quartiles of perfluorodecanoate (PFDA) had 4%
44 lower mean concentrations of triiodothyronine (T3) and women within the highest quartile of
45 perfluoroundecanoate (PFUnDA) had 3% lower mean concentrations of free triiodothyronine
46 (FT3). Further, the difference in concentrations and the changes between three time points were
47 the same for the PFAS quartiles. Thyroxin binding capacity was associated with all the THs
48 and TH-BPs, and was selected as a holistic adjustment for individual changes in TH
49 homeostasis during pregnancy. Finally, adjusting for maternal iodine status did not influence
50 the model predictions.

51 Findings in the present study suggest modifications of TH homeostasis by PFASs in a
52 background exposed maternal population. The variation in levels of THs between PFAS
53 quartiles were within normal reference ranges and may not be of clinical significance in the
54 pregnant woman. However, subtle individual changes in maternal THs may have significant
55 consequences for foetal health.

56

57 **Abbreviations:**

58 Anti-TPO, Anti-thyroid peroxidase antibodies; HTP, Hypothalamic pituitary; LOD, Limit of
59 detection; MISA, The Northern Norway Mother-and-Child contaminant Cohort Study; PFASs,
60 Poly- and perfluoroalkyl substances; PFDA, Perfluorodecanoate; PFDoDA,
61 Perfluorododecanoate; PFHpS, Perfluoroheptane sulfonate; PFHxS, Pefluorohexane sulfonate;
62 PFNA, Perfluorononanoate; PFOA, Perfluorooctanoate; PFOS, Perfluorooctane sulfonate;
63 PFUnDA, Perfluoroundecanoate; PLS, Partial least square; T3, Triiodothyronine; FT3, Free
64 triiodothyronine; T4, Thyroxin; FT4, Free thyroxin; T-Uptake, Thyroxin binding capacity;
65 TBG, Thyroid binding globulin; TH, Thyroid hormone; TH-BP, Thyroid hormone binding
66 protein; TSH, Thyroid stimulating hormone; TTR, Transthyretin; UHPLC-MS/MS, Ultrahigh
67 pressure liquid chromatography triple–quadrupole mass-spectrometry.

68

69 **Keywords: Thyroid hormones; perfluoroalkyl substances; pregnant women; thyroxin**
70 **binding capacity; endocrine disruption**

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85 **1. Introduction**

86 Thyroid hormones (THs) like thyroid stimulating hormone (TSH), thyroxin (T4) and
87 triiodothyronine (T3), are involved in numerous physiological processes e.g. regulation of
88 metabolism, bone remodelling, cardiac function and mental status in the adult. For the embryo
89 and foetus, THs are crucial in all developmental stages. The onset of foetal thyroid function is
90 at approximately 20 weeks gestation, and thus prior to this, maternal T4 is the sole source of
91 TH to the developing foetal brain (Morreale De et al., 2004). In adults, THs are produced in the
92 thyroid gland and transported to peripheral target tissues aided by thyroid hormone binding
93 proteins (TH-BPs) e.g. thyroid binding globulin (TBG), transthyretin (TTR), and albumin. The
94 thyroid function is regulated by negative feedback mechanisms, in which TSH stimulates the
95 thyroid to synthesize T4 which is further converted to T3. TSH is in turn regulated by the
96 hypothalamus as well as by the levels of circulating T3 and T4. In healthy individuals, serum
97 levels of THs are maintained relatively stable with individuals having his or her specific set
98 point (Feldt-Rasmussen et al., 1980).

99

100 During the first two trimesters of pregnancy, marked changes are seen in the maternal
101 hypothalamic pituitary (HTP) thyroid axis to increase the availability of THs. In short, these
102 changes lead to a two- to three-fold increase in TBG production and a subsequent decrease in
103 levels of free thyroxin (FT4) and free triiodothyronine (FT3) followed by an increased
104 production of T3 and T4. The increase in T3 and T4 is less than the increase in TBG, resulting
105 in a decreased T4/TBG ratio, creating a state of relative hypothyroxinemia. Hence, these
106 adaptations mimic hyperthyroidism, but thyroid function per se does not change during
107 pregnancy. There is uncertainty regarding reference ranges for thyroid tests during pregnancy
108 as pregnancy-induced changes in thyroid physiology affects laboratory interpretation and
109 presently no universally accepted reference ranges exist (Fitzpatrick and Russell, 2010).
110 Changes in individual TH levels throughout pregnancy varies by gestational age, number of
111 foetuses and study population, but generally, the woman achieves a new steady state in HTP
112 function at the end of 2nd trimester which is maintained until delivery. After delivery, the
113 alterations in thyroid processes are gradually reversed over 4-6 weeks (Blackburn, 2013).

114

115 Endocrine systems like the thyroid are susceptible to disruption by naturally-occurring and
116 man-made compounds, possibly by affecting the hormone homeostasis through carrier proteins
117 and receptors. One group of potential endocrine disrupting chemicals are poly- and
118 perfluoroalkyl substances (PFASs). PFASs are persistent substances that have been directly
119 emitted to the environment, intentionally or as by-products, during their production and use
120 (Prevedouros et al., 2006). Diet is currently suspected to be the major on-going exposure
121 pathway of PFASs for humans (Fromme et al., 2009; Haug et al., 2011a; Vestergren and
122 Cousins, 2009). In addition, these chemicals are passed to humans through air, house dust,
123 drinking water and water based beverages (Eschauzier et al., 2013; Haug et al., 2011a; Haug et
124 al., 2011b; Ullah et al., 2011).

125

126 Scientific and public concern regarding PFASs, are their potential to perturb maternal hormonal
127 homeostasis and subsequently affect pregnancy outcome by increasing the risk of spontaneous
128 abortions, placental disruptions, foetal distress, malformations, prematurity, decreased birth
129 weight, and hypertension (Boas et al., 2012; Morreale De et al., 2000; Morreale De et al., 2004;
130 Stahl et al., 2011). Disruption of the maternal thyroid homeostasis during early pregnancy is of
131 particular concern, where subclinical changes in maternal THs may affect embryonic and foetal
132 development (Boas et al., 2012). Compared to the wide population reference ranges for THs,
133 the range of variation within each individual are narrower. Hence, subtle changes in the
134 individual set point of thyroid homeostasis may have significant effects, especially if occurring
135 during critical developmental periods (Feldt-Rasmussen et al., 1980).

136

137 T3 and T4 are the only biological molecules which are halogenated (iodine). Similarly, PFASs
138 are halogenated (fluorine) with active sites that resembles those of T3 and T4 (Preau et al.,
139 2014). When assessing effects of PFASs on TH homeostasis, the relevant mechanisms of
140 disruption are; i) disturbance of the overall activity of the thyroid gland by interference with the
141 TH receptors, ii) stimulation or inhibition of enzyme functions which mediates iodine uptake
142 of the thyroid gland in the synthesis of T3 and T4, and iii) competitive displacement of THs on
143 their binding proteins (Boas et al., 2012). Disruption of the thyroid function is often investigated
144 in regards to hypothyroidism with the reporting of TSH concentrations. TSH levels can reflect
145 mild thyroid functional impairment even when T4 and T3 concentrations are within normal
146 ranges but hypothyroxinemia can still occur with normal TSH and T3 concentrations. Hence,

147 in the absence of assessment of the overall thyroid function; the clinical importance of
148 individual TH levels is unclear (Braverman and Utiger, 1986). Therefore, the present study aims
149 to investigate the overall thyroid function in relation to PFAS concentrations by investigating
150 associations between all the THs (TSH, T3, T4, FT3, FT4), thyroxin binding capacity, anti-
151 thyroid peroxidase antibodies (anti-TPO), thyroid hormone binding proteins (TH-BPs) (TBG,
152 TTR and albumin) at three time points; 2nd trimester of pregnancy, 3 days and 6 weeks after
153 delivery and PFAS concentrations in women from Northern Norway.

154

155 **2. Materials and methods**

156 *2.1 Study participants and collection of blood samples*

157 The selected subjects in the present study represent the 391 women who completed the Northern
158 Norway Mother-and-Child Contaminant Cohort Study (MISA) which consists of 515 enrolled
159 pregnant women, recruited from June 2007 to October 2009 (recruitment period; 867 days). All
160 participants answered a detailed questionnaire about diet and lifestyle at enrolment, and donated
161 a blood sample at three visits/time points related to their pregnancy (around gestational week
162 18, 3 days and 6 weeks after delivery). Detailed information about the study group
163 characteristics, ethical approvals, the food frequency questionnaire (FFQ) and the blood
164 collection procedures have been reported elsewhere (Hansen et al., 2010; Veyhe et al., 2012).

165

166 *2.2 Chemical analyses*

167 *2.2.1 PFAS analyses*

168 Blood samples donated at median gestational week 18 (ranging 10–34) were analysed for a
169 variety of PFASs. A total of 26 PFASs were initially screened for in a sub-group of 50 serum
170 samples. Compounds detected above the limit of detection (LOD) in more than 20% of the
171 samples were further quantified in the remaining serum samples (N = 391). Detailed
172 information about the compounds, sample preparation, extraction method, analytical method,
173 reagents and instrumentation has been reported elsewhere (Berg et al., 2014; Hanssen et al.,
174 2013). Briefly, PFASs were determined in serum samples using sonication-facilitated liquid–
175 liquid extraction, activated ENVI-carb clean-up (Powley et al., 2005) and analysed by ultrahigh
176 pressure liquid chromatography triple–quadrupole mass–spectrometry (UHPLC-MS/MS).

177

178 Quantification of the compounds was performed by the internal standard addition method with
179 isotope-labeled PFASs (Hanssen et al., 2013). Further details regarding quality control have
180 been reported elsewhere (Berg et al., 2014)

181

182 *2.2.2 TH and TH-BP analyses*

183 Determination of TH, TH-BP, thyroxin binding capacity and anti-TPO concentrations in non-
184 fasting serum samples from three visits (second trimester, 3 days and 6 weeks after delivery)
185 were performed by laboratory staff at the University Hospital of Northern Norway, Department
186 of Laboratory Medicine. The analyses are routine analyses used in the clinic for diagnostic
187 purposes except for T3, T4 and thyroxin binding capacity. Details on the different methods,
188 instrumentation, analytical variation and reference ranges are provided in Table S1 in the
189 Supplemental Material. The laboratory is certified according to ISO 151810 (Norwegian
190 accreditation, 2014) and all reagents, calibrators and equipment were CE-approved. Quality
191 controls are run at three different concentrations every day and additionally the laboratory
192 participates in the LabQuality external quality assessment program (Labquality Finland, 2014).

193

194 *2.3 Statistical analyses*

195 Statistical analyses were performed using SPSS statistic software, version 22 (IBM SPSS Inc.
196 Chicago, IL, USA). Statistical significance was defined as $p < 0.05$. All PFAS, TH and TH-BP
197 results were log-transformed in the statistical analyses. For PFASs, only compounds with
198 detection frequencies above 80% were evaluated in statistical models where concentrations
199 below LODs were replaced by $\text{LOD}/\sqrt{2}$ (Anda et al., 2007). Partial least square (PLS)
200 regressions were used for data reduction and screening for important variables. Mixed effects
201 linear models were used to investigate potential associations between PFASs and three repeated
202 measurement of THs, where the pregnant women were assigned to quartiles according to PFAS
203 concentrations in 2nd trimester. Separate models were built for five dependent variables; TSH,
204 T3, T4, FT3 and FT4. PFAS quartiles and TH-BPs were included as fixed factors and
205 covariates, respectively. A quadratic development over time was included as fixed factor in all
206 the models. The variance of the fixed factors and the distribution of significant covariates were

207 homogeneous across PFAS quartiles. Diagnostic plots of the residuals and potential influential
208 points were evaluated.

209

210 **3. Results:**

211 *3.1 Population characteristics and PFAS concentrations*

212 The median age was 32 and the majority of the participants were nulliparous or primiparous
213 (parity varied from 0-4), for the MISA study population. Further details on demographic
214 characteristics are briefly presented in the Supplemental Material, Table S2 and described in
215 detail elsewhere (Veyhe et al., 2012). Seven PFASs; pefluorohexane sulfonate (PFHxS),
216 perfluoroheptane sulfonate (PFHpS), perfluorooctane sulfonate (PFOS), perfluorooctanoate
217 (PFOA), perfluorononanoate (PFNA), perfluorodecanoate (PFDA) and perfluoroundecanoate
218 (PFUnDA) were detected in more than 80% of blood samples and were included in the
219 statistical analyses. PFOS (median of 8.03 ng/mL) was the dominating compound followed by
220 PFOA (1.53 ng/mL), PFNA (0.56 ng/mL), PFHxS (0.44 ng/mL), PFUnDA (0.26 ng/mL),
221 PFDA (0.23 ng/mL) and PFHpS (0.10 ng/mL). Spearman correlation coefficients showed high
222 to moderate correlations between PFOS and PFOA ($r=0.65$), PFHxS, ($r=0.63$), PFHpS ($r=0.68$),
223 PFNA ($r=0.60$), PFDA ($r=0.57$) and PFUnDA ($r=0.45$). The correlation between PFDA and
224 PFUnDA was higher than between PFOS and any of these two compounds. PFAS
225 concentrations and their predictors are described in detail in a previous publication (Berg et al.,
226 2014).

227

228 *3.2 Concentrations of THs and TH-BPs*

229 Concentrations and specific study group reference ranges of THs, TH-BPs and thyroxin binding
230 capacity are provided in Table 1. Sixteen women were excluded from the study due to self-
231 reported thyroid related disease and/or use of medications. Further, 22, 15 and 15 women had
232 thyroid peroxidase antibodies above 34 IU/L at visit 1, 2 and 3 respectively, and were
233 categorized as anti-TPO positive according to the reference range applied by the manufacturer.
234 The anti-TPO positive women were included in all analyses, tables and figures as they did not
235 alter the variance in TH concentrations compared to anti-TPO negative women. The specific

236 study group reference range (2.5–97.5th percentile) for the THs were within the normal
237 reference ranges for the respective hormones.

238

Table 1

Maternal concentrations^a of THs, TH-BPs, thyroxin binding capacity and anti-TPO, and study group specific reference ranges at three repeated measurements

| Compound | Visit 1: 2nd trimester, N= 375 | | | | Visit 2: 3 days postpartum, N=372 | | | | Visit 3: 6 weeks postpartum, N= 374 | | | |
|--|--------------------------------|----------|------|------------------------------|-----------------------------------|----------|------|------------------------------|-------------------------------------|----------|------|------------------------------|
| | Median (Range) | AM | SD | Reference range ^b | Median (Range) | AM | SD | Reference range ^b | Median (Range) | AM | SD | Reference range ^b |
| TSH (mIU/L) | 1.55 (0.06-10.2) | 1.76 | 1.04 | 0.44-4.48 | 2.37 (0.15-9.51) | 2.56 | 1.20 | 0.98-5.39 | 1.39 (0.06-6.54) | 1.55 | 0.80 | 0.47-3.38 |
| T3 (nmol/L) | 2.71 (1.47-4.75) | 2.75 | 0.46 | 1.97-3.73 | 2.75 (1.32-4.66) | 2.80 | 0.50 | 1.81-3.88 | 1.70 (1.15-2.53) | 1.72 | 0.22 | 1.29-2.22 |
| T4 (nmol/L) | 145 (92.00-215) | 146 | 21.1 | 111-190 | 144 (77.0-232) | 145 | 26.3 | 97.1-204 | 97.0 (63.0-153) | 98.5 | 14.3 | 72-130 |
| FT3 (pmol/L) | 4.59 (2.99-7.08) | 4.62 | 0.53 | 3.66-5.79 | 4.50 (2.72-6.77) | 4.52 | 0.57 | 3.50-5.73 | 4.63 (3.14-6.48) | 4.66 | 0.45 | 3.80-5.77 |
| FT4 (pmol/L) | 13.0 (9.00-20.0) | 13.4 | 1.62 | 10.0-17.0 | 13.0 (8.00-19.0) | 13.0 | 1.87 | 10.0-17.0 | 14.0 (10.0-25.0) | 14.3 | 1.74 | 12.0-18.0 |
| TBG (mg/L) | 36.7 (23.2-69.6) | 37.2 | 6.74 | 26.2-53.3 | 37.0 (19.9-56.7) | 37.3 | 6.35 | 25.2-51.4 | 17.9 (11.7-39.1) | 18.4 | 3.44 | 12.9-26.54 |
| TTR (g/L) | 0.19 (0.09-0.27) | 0.19 | 0.03 | 0.15-0.25 | 0.19 (0.09-0.42) | 0.19 | 0.04 | 0.11-0.29 | 0.22 (0.13-0.38) | 0.22 | 0.03 | 0.16-0.29 |
| Albumin (g/L) | 40.0 (33.9-47.4) | 40.2 | 2.42 | 36.0-46.0 | 34.9 (24.3-46.4) | 34.8 | 3.15 | 28.0-40.0 | 46.5 (40.9-53.6) | 46.6 | 2.27 | 42.0-51.0 |
| Thyroxin binding capacity (TBI)^c | 1.26 (0.84-1.50) | 1.26 | 0.09 | 1.07-1.43 | 1.29 (0.20-1.50) | 1.28 | 0.10 | 1.10-1.42 | 1.01 (0.50-1.18) | 1.00 | 0.07 | 0.81-1.11 |
| | Positive > 34 IU/L | % | | | Positive > 34 IU/L | % | | | Positive > 34 IU/L | % | | |
| Anti-TPO (IU/mL) | 22 | 6 | | | 15 | 4 | | | 15 | 4 | | |

^aAnti-TPO positive women are included in medians. The same 15 women were anti-TPO positive at all three visits

^bDefined as the 2.5 percentile (lower range) and 97.5 percentile (upper range) for this population

^cThyroxin binding index, the measure unit for thyroxin binding capacity

241 *3.3 THs and associations to PFAS concentrations*

242 After adjusting for significant covariates such as parity, age, thyroxin binding capacity and
243 BMI, and regardless of including anti-TPO positive women or not, TSH was positively
244 associated with PFOS; T3 was negatively associated with PFDA; and FT3 was negatively
245 associated with PFUnDA (Table 2). The significant covariates were included in the respective
246 models (Table 2), but many more were evaluated (e.g. iodine sufficiency and gestational week)
247 and are listed in Table S2 in the Supplemental material. Thyroxin binding capacity was
248 significantly associated with all the thyroid hormones and the individual binding proteins, and
249 was selected as a holistic adjustment for individual changes in thyroid hormone homeostasis
250 during pregnancy and postpartum periods. Several PFASs (Results not presented) were
251 inversely associated with T4 and FT4 after adjusting for age and BMI, but after including
252 thyroxin binding capacity the associations were no longer significant. Further, adjusting for
253 estimated daily dietary intakes of iodine ($\mu\text{g}/\text{day}$) and concentrations of iodine in urine ($\mu\text{g}/\text{L}$
254 for 212 participants), did not influence the model predictions (results not presented) when
255 grouped into iodine sufficient, mildly deficient and deficient, according to guidelines from the
256 World Health Organisation (World Health Organization et al., 2014).

257

258 Women in the highest PFOS quartile had higher mean concentration of TSH at all three time
259 point compared to women in the first quartile (Figure 1 A and B). Similar, women in the highest
260 PFDA and PFUnDA quartiles had lower T3 and FT3 concentrations, respectively (Figure 1 C,
261 D and E, F). Further, the proportion of women with a depleted supply of T4/FT4 and T3/FT3
262 (subclinical hypothyroidism), characterized by elevated TSH concentrations (>3.6 mIU/L) but
263 with normal FT4 and FT3 concentrations (Fitzpatrick and Russell, 2010), increased for each
264 PFOS quartile (Q1: n=12, Q2: n=16, Q3: n=24, Q4: n=30). Women in the extreme quartiles of
265 PFHxS and PFOA had higher concentrations of TSH compared to the lowest quartiles, but when
266 including PFOS concentration as a covariate, the associations were no longer significant (results
267 not presented). PFDA was negatively associated with FT3, but not after adjusting for PFUnDA
268 (results not presented). Finally, the difference as well as the relative change in thyroid hormone
269 concentrations over time was constant between PFAS quartiles (parallel growth curves, Figure
270 1 B, D and E).

271

Table 2Mixed-effects model estimated mean^a differences in thyroid hormone concentrations over time between PFAS quartiles

| Fixed factor | | Model ^b | | |
|-----------------------------------|------|-----------------------------------|--------------------------|----------|
| | | Model 1 ^d : TSH mIU/L | | |
| | | \hat{Y} | 95 % Confidence Interval | <i>p</i> |
| PFOS (ng/mL)^c | | Reference | | |
| Quartile 1: 0.3-5.7 | N=94 | Reference | | |
| Quartile 2: 5.8-8.0 | N=90 | 0.18 | 0.06, 0.31 | 0.11 |
| Quartile 3: 8.1-11.0 | N=95 | 0.26 | 0.13, 0.40 | 0.03 |
| Quartile 4: 11.1-35.9 | N=96 | 0.35 | 0.21, 0.50 | 0.00 |
| | | Model 2 ^f : T3 nmol/L | | |
| | | \hat{Y} | 95 % Confidence Interval | <i>p</i> |
| PFDA (ng/mL)^e | | Reference | | |
| Quartile 1: 0.05-0.17 | N=93 | Reference | | |
| Quartile 2: 0.17-0.23 | N=94 | -0.04 | -0.08, 0.04 | 0.46 |
| Quartile 3: 0.23-0.31 | N=94 | -0.05 | -0.08, 0.00 | 0.52 |
| Quartile 4: 0.31-2.34 | N=94 | -0.1 | -0.14,-0.06 | 0.03 |
| | | Model 3 ^f : FT3 pmol/L | | |
| | | \hat{Y} | 95 % Confidence Interval | <i>p</i> |
| PFUnDA (ng/mL)^g | | Reference | | |
| Quartile 1: LOD-0.15 | N=92 | Reference | | |
| Quartile 2: 0.16-0.25 | N=94 | -0.08 | -0.15, -0.00 | 0.14 |
| Quartile 3: 0.26-0.37 | N=94 | -0.09 | -0.16, -0.01 | 0.23 |
| Quartile 4: 0.4-0.96 | N=95 | -0.18 | -0.25, -0.12 | 0.00 |

^aMean differences in TH concentrations (\hat{Y}) are backtransformed from log-estimates of fixed effect variables^bModels are based on three measurements of THs per subject and included a subject-specific random intercept^cCoefficients express change for TSH concentrations across PFOS quartiles, with quartile 1 as the reference group^dParity and thyroxin binding capacity were included as covariates (fixed effects variables) in the model^eCoefficients express change for T3 concentrations across PFDA quartiles, with quartile 1 as the reference group^fAge, BMI and thyroxin binding capacity were included as covariates (fixed effects variables) in the model^gCoefficients express change for FT3 concentrations across PFUnDA quartiles, with quartile 1 as the reference group

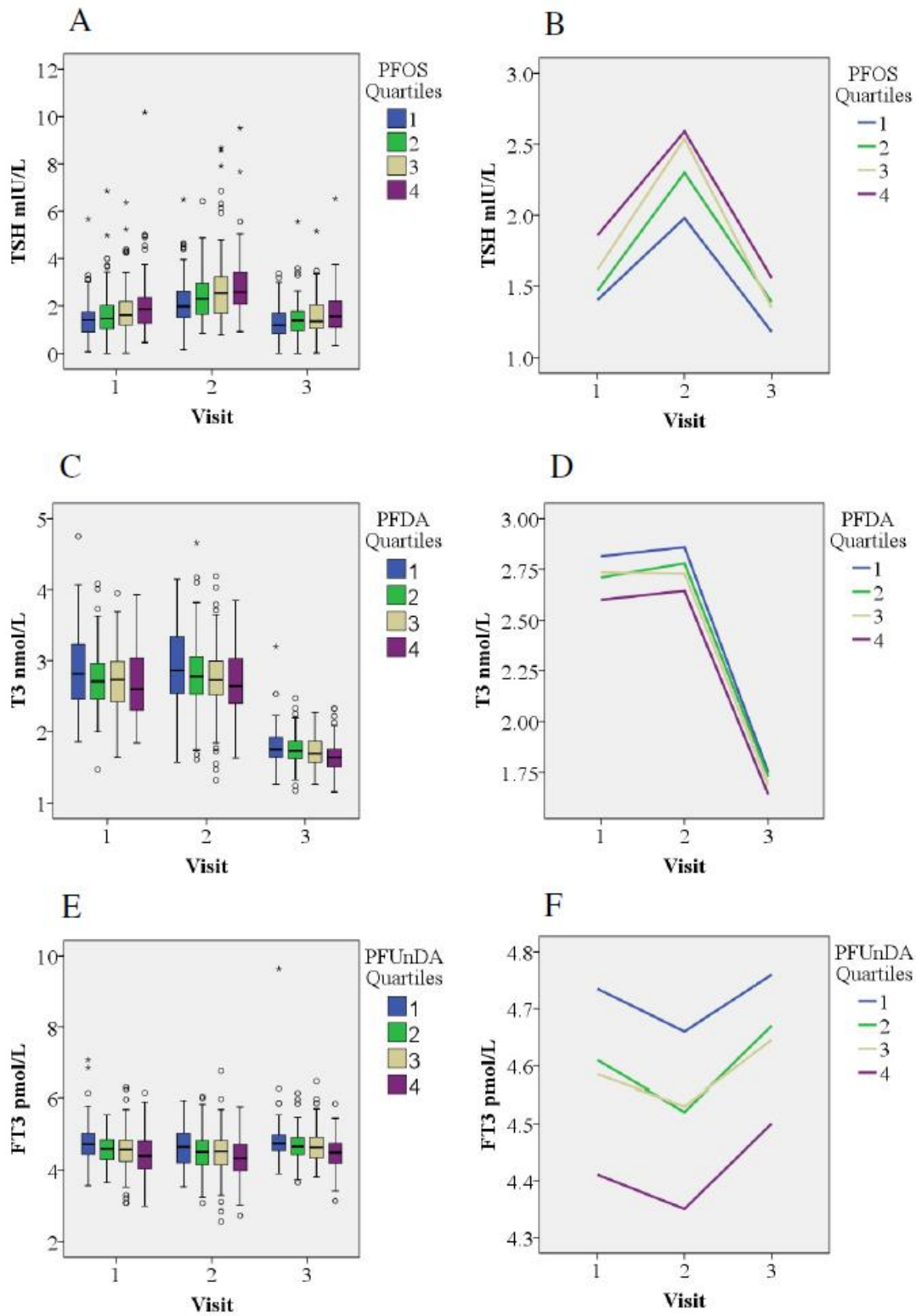
272

273

274

275

276



277

278 Figure 1. TSH, T3 and FT3 concentrations for the PFAS quartiles at three sampling points,
 279 presented as boxplot (A, C and E) and as median concentrations (B, D and F).

280

281 **4. Discussion:**

282 *4.1 Associations between TH and PFAS concentrations*

283 The observed associations in our study suggest modifications of the thyroid homeostasis by
284 PFASs in a background exposed maternal population. The results demonstrate higher TSH
285 concentrations with higher PFOS concentrations in pregnant women. Women within the highest
286 PFOS quartile had 24% higher mean TSH concentrations compared to the first quartile at all
287 sampling points. These observations are in accordance with those in another pregnant
288 population in Norway (Wang et al., 2013), where PFOS concentrations were positively
289 associated to TSH concentrations. Furthermore, the proportion of woman being classified with
290 subclinical hypothyroidism at visit 1 (2nd trimester) were higher with increasing PFOS
291 concentrations in our study. Higher PFHxS and PFOA concentrations were associated to higher
292 TSH concentrations, although not significantly when adjusting for PFOS concentrations. In
293 comparison, PFHxS was positively associated to TSH levels but not to PFOS in a pregnant
294 population from the Taiwan Maternal and Infant Cohort Study (Wang et al., 2014). The results
295 indicate that PFOS can interfere with the production and elimination of T3 and T4, where
296 elevated levels of TSH in women with high concentrations of PFOS could be the adjusted
297 homeostasis state due to a reduction in T3 and T4. TSHs ability to maintain equilibrium in the
298 TH homeostasis, may further explain that the variance in levels of T3 and T4 between PFOS
299 quartiles were within the normal reference ranges in our study. In accordance with
300 investigations carried out in rats, a single dose of PFOS caused a reduction in T3 and T4 levels
301 (Chang et al., 2008).

302

303 In the present study PFDA concentrations were inversely associated to T3, and the women
304 within the highest quartiles had 4% lower concentrations compared to women in the first
305 quartile at all sampling points. Similar, PFUnDA concentrations were inversely associated to
306 FT3 where women within the highest quartile had 3% lower concentrations. To the best of our
307 knowledge, inverse associations between longer chained PFASs and maternal serum T3 and
308 FT3 levels have not been reported previously, whereas Wang et al. (2014) reported a significant
309 association between PFDA and cord blood T3 levels. Presently, there are no reports on the
310 mechanisms by which PFDA and PFUnDA might modify thyroid hormones in humans or
311 animals, but in a study conducted on a rat pituitary tumour cell line expressing intracellular
312 thyroid receptor (TR), PFDA and PFUnDA significantly decreased the T3-induced cell

313 proliferation (Long et al., 2013). The authors therefore believe that PFDA and PUnDA might
314 compete with T3 binding to TR. Considering the increasing temporal trends of longer chained
315 PFASs in humans (Nost et al., 2014) their endocrine disrupting potencies are a growing concern
316 and studies on thyroid disrupting effects of these longer chained PFASs are warranted.

317

318 PFASs have been indicated to interfere at several levels of the thyroid homeostasis. We
319 observed that PFASs were associated to individual TH set points but not to the relative change
320 in TH levels across sampling period. The quartile differences in TH levels were consistent at
321 all the time points with no interaction between time and quartiles, thus PFASs exert its effect
322 in a constant matter. Further, we did not observe any associations between thyroxin binding
323 capacity and PFOS, as was reported in the C8 health project (Knox et al., 2011) suggesting
324 competitive displacement of T4 by PFOS on TH-BPs. This could be explained by elevated TH-
325 BP levels in pregnancy and a subsequent excess in thyroxin binding sites, which likely renders
326 it difficult to detect any displacement of T4 by PFASs on TH-BPs in this study.

327

328 Studies performed on pregnant women present conflicting results in regards to thyroid
329 disrupting potencies of PFASs (Chan et al., 2011; Wang et al., 2013; Wang et al., 2014; Webster
330 et al., 2014), whereas similar studies in animals demonstrate consistent physiological effects
331 including disruption of thyroid hormone homeostasis (Fuentes et al., 2006; Lau et al., 2003;
332 Thibodeaux et al., 2003). Different exposure histories in different populations may influence
333 the associations between PFASs and THs as well as variation in lifestyle predictors such as
334 parity, breastfeeding, birth year and study period that affect PFAS concentrations (Berg et al.,
335 2014). Hence, measurement of PFASs at one time point may not reflect the historic exposure
336 to PFASs and potential early effects on thyroid function. Further, the human exposure scenario
337 with lifelong exposure to a mixture of chemicals in low doses, the large physiological variation
338 in TH levels between individuals and complicated pathways, render human effect studies more
339 difficult to perform. Finally, TSH, T3 and T4 are tightly regulated within a given individual,
340 where the expected inter-individual variations may camouflage differences associated with
341 exposure.

342

343 *4.2 Study design and strengths*

344 To the best of our knowledge, this is the first study investigating the effect of PFASs on ten
345 thyroid hormone parameters, and we have also included three repeated measurements. Due to
346 the complex thyroid system, assessment of potential thyroid impairment cannot be interpreted
347 from individual TH levels only. Furthermore, T3 and T4 levels per se are not adequate hormone
348 indicators in pregnant women (Glinoeer and Spencer, 2010) due to the alterations in TH levels,
349 blood composition and volume. To accompany this complexity we have adjusted for the
350 increase in thyroxin binding capacity (reflects elevated levels of all the TH-BPs) as a proxy for
351 the pregnancy related alterations in blood THs and TH-BPs in statistical models. We have also
352 reported concentrations of FT3 and FT4 in addition to T3 and T4, as the evaluation of these are
353 generally preferred in pregnant women due to that the increase in TH-BPs may mask an actual
354 decrease in levels of T4 and T3. We have reported TSH levels as well because TSH reflects the
355 thyroid status more directly, and can reveal abnormalities in T3 and T4 levels despite apparently
356 normal levels of these THs (Glinoeer and Spencer, 2010). Still, in a situation with a low supply
357 of T4, normal T3 levels might prevent an increase in circulating TSH and consequently,
358 hypothyroxinemia will not be detected if only TSH is measured (Braverman and Utiger, 1986).
359 Further, as the thyroid homeostasis is dynamic and a single measurement may not adequately
360 characterize the maternal thyroid function, we have reported repeated measurements of THs.

361

362 In the present study, PFHxS, PFDA and PFUnDA were inversely associated to FT4 and T4.
363 However, when adjusting for the levels of TH-BPs by including thyroxin binding capacity,
364 associations were no longer observed. Consequently, had we not corrected for the influence of
365 TH-BPs on TH levels, we would have reported significant associations between several PFASs
366 and FT4/T4. This might explain why our observations are not in accordance to the most recent
367 publication on PFASs and THs in a pregnant population (Wang et al., 2014), who reported
368 PFUnDA to be significant inversely associated with FT4 and T4. That study did not report on
369 adjustments for gestational week or elevated levels of TH-BPs. Further, due to differences in
370 PFAS concentrations and sampling points during pregnancy, the results from the two studies
371 may not be comparable. Hence, inconsistencies in analytes and covariates investigated may
372 account for some of the discrepancies in observed relationships between PFASs and THs in
373 different studies which complicate the conclusion on definite relations.

374

375 *4.3 Clinical relevance*

376 Concentrations of all the THs varied within normal reference ranges and as the thyroid system
377 is tightly regulated, small changes in TH concentrations will likely be adjusted for through the
378 negative mechanisms which naturally secure TH equilibrium. Consequently, the indicated
379 PFAS induced changes in TH concentrations, may not have caused clinical effects in the
380 mother, thus the physiological importance of the observations is not established. Several studies
381 have described ranges for thyroid hormone levels during pregnancy but have demonstrated
382 variation by gestational age, number of foetuses, population studied, laboratory, and testing
383 method (Fitzpatrick and Russell, 2010). In the present study, increasing levels of TSH according
384 to increased PFOS concentrations may indicate low individual levels of T4 in these women.
385 The specific reference ranges for T4 and FT4 in this study group were within the mid to high
386 end of the normal reference ranges (Table 1 and Table S1), while the specific reference ranges
387 for PFOS quartile four (results not presented) were within the lower end. Considering that levels
388 of THs in pregnant women should be 40-100% higher than in non-pregnant (Blackburn, 2013),
389 the subsequent distribution should have been towards the higher end of the reference range for
390 the respective T4 and FT4 levels to be “normal”. Hence, the clinical relevance of individual
391 levels might be masked in non-pregnant population reference ranges.

392

393 Independent of the mode of action of PFASs on the maternal thyroid function and the
394 subsequent clinical relevance, disruption of maternal TH homeostasis in any degree would only
395 increase the difficulties encountered by the newborn in meeting postnatal hormone
396 requirements (Morreale De et al., 2000). The foetus relies on maternal THs throughout gestation
397 and a normal supply of maternal T4 has an important protective role also after midgestation.
398 This is underlined by reports of poor developmental outcome e.g. impaired mental development
399 and growth in babies faced with a premature interruption of the maternal supply of THs
400 (Morreale De et al., 2000). Although the indicated PFAS induced changes in TH concentrations
401 were within the reference ranges in the present study, small changes in THs may affect foetal
402 development, especially if occurring during critical periods. Therefore, concerns have been
403 raised regarding the effect of mild maternal thyroid hormone deficiency on foetal
404 neurodevelopment. Decreases in childhood intellectual performance can occur even when a
405 pregnant woman’s hypothyroidism is subclinical (mild and asymptomatic) where marginally
406 low T4 levels in the pregnant woman cause reduction in cognitive functions of the offspring
407 (Berbel et al., 2009; Haddow et al., 1999; Pop et al., 2003).. Hence, subtle changes in THs may
408 have significant consequences for foetal health and consistent evidence confirms that disrupted

409 maternal thyroid homeostasis negatively affects newborn development (Morreale De et al.,
410 2000).

411

412 *4.4 Thyroid function and iodine*

413 The two principal causes of maternal hypothyroidism are iodine deficiency and exposure to
414 xenobiotic thyroid disruptors. The negative health effects during pregnancy of thyroid
415 hormones that have been associated to PFAS exposures, can resemble those related to iodine
416 deficiency (eg. decreased maternal FT4 and increased maternal TSH, increased risk of
417 prematurity, spontaneous abortion, and neurodevelopmental impairment) (Morreale De et al.,
418 2000; Morreale De et al., 2004; Stahl et al., 2011). Hence, the importance of assessing iodine
419 status when investigating associations between thyroid disruptors and THs is obvious. In the
420 present study, maternal iodine status did not influence the observed associations between
421 PFASs and THs and was not included as a covariate in the final models. Still, iodine status may
422 affect the variance in TH concentrations, as the degree of iodine sufficiency or deficiency
423 affects individual TH set points and changes in concentrations throughout the pregnancy
424 (Blackburn, 2013; Morreale De et al., 2004). For example, iodine deficient women may be more
425 susceptible to TH disruption. This could not be observed in our cohort as the variation in iodine
426 status was low. This is also important for the general population and emphasizes that iodine
427 status may interact with PFAS effects on thyroid homeostasis. Still, the majority of studies are
428 not considering iodine status in regards to thyroid disruptors.

429

430 *4.5 Mixture effects*

431 Although the statistical analyses investigate the relationship between one or several dependent
432 and independent variables at the same time, contaminants do not occur isolated in the human
433 blood circulation. This means that the actual causality between physiological processes and the
434 impairment by contaminants is complicated by the complex correlation of exposures. Also,
435 there might be interactions and dose dependencies that we are not able to capture. This is evident
436 in the observation of PFOA, PFHxS and PFDA being significantly associated to THs in
437 individual models, and where the associations were no longer significant when adjusting for
438 dominant components such as PFOS. Due to the strong correlation between the contaminants
439 and their joint explanation of the outcome, it was not possible to isolate the variance in THs

440 explained by PFOA, PFHxS or PFDA alone. A possible solution to this is to summarize
441 contaminants based on similarities e.g. chemical properties, mode of action and emission
442 patterns, and report mixture effects, but this is beyond the scope of this publication. For further
443 progression in the research on contaminants and human health effects, it is important to assess
444 multiple exposure scenarios. Therefore, in subsequent studies we aim to include several
445 contaminants and thyroid related outcomes in the children, to assess associations of thyroid
446 function with combined exposures.

447

448 **5. Conclusions**

449 Women with the highest concentrations of PFOS, PFDA and PFUnDA had consistently higher
450 TSH and lower T3 and FT3 concentrations, respectively, at all sampling points compared to
451 women with the lowest concentrations. The difference in concentrations and the changes
452 between three time points were the same for the PFAS quartiles. Despite significant associations
453 between PFASs and THs, TH levels were within normal reference ranges and may not be of
454 clinical significance in the pregnant woman. However, the foetus relies on maternal THs
455 throughout gestation and a normal supply of maternal T4 has an important protective role also
456 after midgestation, hence, subtle individual changes in thyroid hormones may have significant
457 consequences for foetal health.

458

459 **Acknowledgements**

460 The project was financially supported by the Northern Norway Regional Health Authority, the
461 EU project ArcRisk (www.arcrisk.eu) and The Fram Centre Flagship research programme
462 Hazardous substances. The authors wish to thank the participating mothers and the service
463 provided by the Medical Birth Registry of Norway (MBRN). We gratefully acknowledge the
464 collaboration with the colleagues at Department of Laboratory Medicine, UNN and NILU-
465 Norwegian Institute of Air Research with special thanks to Lisbeth Hansen, Tom Sollid, Sandra
466 Huber and Elbjørg Sofie Heimstad for valuable input and advice. Special appreciation is
467 extended to Bente Augdal, UIT, for her contribution to the project. The authors declare that
468 they have no competing financial interests.

469

470
471
472
473
474
475
476
477
478
479
480
481
482
483
484
485
486
487
488
489

Reference List

491

492 Anda EE, Nieboer E, Dudarev AA, Sandanger TM, and Odland JO. Intra- and intercompartmental
493 associations between levels of organochlorines in maternal plasma, cord plasma and breast

- 494 milk, and lead and cadmium in whole blood, for indigenous peoples of Chukotka,
495 Russia. *Journal of environmental monitoring* 8-6-2007;9:884-93.
- 496 Berbel P, Mestre JL, Santamaria A, Palazon I, Franco A, Graells M, Gonzalez-Torga A, and de
497 Escobar GM. Delayed neurobehavioral development in children born to pregnant women with
498 mild hypothyroxinemia during the first month of gestation: the importance of early iodine
499 supplementation. *Thyroid* 2009;19:511-9.
- 500 Berg V, Nost TH, Huber S, Rylander C, Hansen S, Veyhe AS, Fuskevåg OM, Odland JO, and
501 Sandanger TM. Maternal serum concentrations of per- and polyfluoroalkyl substances and
502 their predictors in years with reduced production and use. *Environ. Int* 6-5-2014;69C:58-66.
- 503 Blackburn, S. T. *Pituitary, Adrenal, and Thyroid function*. Elsevier 2013; Fourth: 627-652.
- 504 Boas M, Feldt-Rasmussen U, and Main KM. Thyroid effects of endocrine disrupting
505 chemicals. *Mol. Cell Endocrinol.* 22-5-2012;355:240-8.
- 506 Braverman LE and Utiger RD. *The thyroid: A fundamental and clinical text*. 9. Lippincott; 1986
- 507 Chan E, Burstyn I, Cherry N, Bamforth F, and Martin JW. Perfluorinated acids and hypothyroxinemia
508 in pregnant women. *Environ. Res.* 2011;111:559-64.
- 509 Chang SC, Thibodeaux JR, Eastvold ML, Ehresman DJ, Bjork JA, Froehlich JW, Lau C, Singh RJ,
510 Wallace KB, and Butenhoff JL. Thyroid hormone status and pituitary function in adult rats
511 given oral doses of perfluorooctanesulfonate (PFOS). *Toxicology.* 20-1-2008;243:330-9.
- 512 Eschauzier C, Hoppe M, Schlummer M, and de VP. Presence and sources of anthropogenic
513 perfluoroalkyl acids in high-consumption tap-water based beverages. *Chemosphere*
514 2013;90:36-41.
- 515 Feldt-Rasmussen U, Hyltoft PP, Blaabjerg O, and Horder M. Long-term variability in serum
516 thyroglobulin and thyroid related hormones in healthy subjects. *Acta Endocrinol. (Copenh)*
517 1980;95:328-34.
- 518 Fitzpatrick DL and Russell MA. Diagnosis and management of thyroid disease in
519 pregnancy. *Obstet. Gynecol. Clin. North Am.* 2010;37:173-93.
- 520 Fromme H, Tittlemier SA, Volkel W, Wilhelm M, and Twardella D. Perfluorinated compounds--
521 exposure assessment for the general population in Western countries. *Int. J. Hyg. Environ. Health*
522 2009;212:239-70.
- 523 Fuentes S, Colomina MT, Rodriguez J, Vicens P, and Domingo JL. Interactions in developmental
524 toxicology: concurrent exposure to perfluorooctane sulfonate (PFOS) and stress in pregnant
525 mice. *Toxicol. Lett.* 20-6-2006;164:81-9.
- 526 Glinoe D and Spencer CA. Serum TSH determinations in pregnancy: how, when and
527 why? *Nat. Rev. Endocrinol.* 2010;6:526-9.
- 528 Haddow JE, Palomaki GE, Allan WC, Williams JR, Knight GJ, Gagnon J, O'Heir CE, Mitchell ML,
529 Hermos RJ, Waisbren SE, Faix JD, and Klein RZ. Maternal thyroid deficiency during
530 pregnancy and subsequent neuropsychological development of the child. *N. Engl. J. Med.* 19-8-
531 1999;341:549-55.
- 532 Hansen S, Nieboer E, Odland JO, Wilsgaard T, Veyhe AS, and Sandanger TM. Levels of
533 organochlorines and lipids across pregnancy, delivery and postpartum periods in women from
534 Northern Norway. *J. Environ. Monit.* 2010;12:2128-37.

- 535 Hanssen L, Dudarev AA, Huber S, Odland JO, Nieboer E, and Sandanger TM. Partition of
536 perfluoroalkyl substances (PFASs) in whole blood and plasma, assessed in maternal and
537 umbilical cord samples from inhabitants of arctic Russia and Uzbekistan. *Sci.Total Environ.* 1-
538 3-2013;447:430-7.
- 539 Haug LS, Huber S, Becher G, and Thomsen C. Characterisation of human exposure pathways to
540 perfluorinated compounds--comparing exposure estimates with biomarkers of
541 exposure. *Environ.Int.* 2011a;37:687-93.
- 542 Haug LS, Huber S, Schlabach M, Becher G, and Thomsen C. Investigation on per- and polyfluorinated
543 compounds in paired samples of house dust and indoor air from Norwegian
544 homes. *Environ.Sci.Technol.* 1-10-2011b;45:7991-8.
- 545 Knox SS, Jackson T, Frisbee SJ, Javins B, and Ducatman AM. Perfluorocarbon exposure, gender and
546 thyroid function in the C8 Health Project. *J.Toxicol.Sci.* 2011;36:403-10.
- 547 Labquality Finland. External quality assessment for medical laboratories;2014. Available from:
548 <http://www.labquality.fi/eqa-eqas/eqa-eqas-program-scheme/external-quality-assessment/>.
- 549 Lau C, Thibodeaux JR, Hanson RG, Rogers JM, Grey BE, Stanton ME, Butenhoff JL, and Stevenson
550 LA. Exposure to perfluorooctane sulfonate during pregnancy in rat and mouse. II: postnatal
551 evaluation. *Toxicol.Sci.* 2003;74:382-92.
- 552 Long M, Ghisari M, and Bonefeld-Jorgensen EC. Effects of perfluoroalkyl acids on the function of the
553 thyroid hormone and the aryl hydrocarbon receptor. *Environ.Sci.Pollut.Res.Int.* 2013;20:8045-
554 56.
- 555 Morreale De EG, Obregon MJ, and Escobar del RF. Is neuropsychological development related to
556 maternal hypothyroidism or to maternal hypothyroxinemia? *J Clin.Endocrinol.Metab*
557 2000;85:3975-87.
- 558 Morreale De EG, Obregon MJ, and Escobar del RF. Role of thyroid hormone during early brain
559 development. *Eur.J Endocrinol.* 2004;151 Suppl 3:U25-U37.
- 560 Norwegian accreditation. Norwegian accreditation;2014. Available from:
561 <http://www.akkreditert.no/en/hva-er-akkreditering/hva-vi-akkrediterer/laboratorier/>.
- 562 Nost TH, Vestergren R, Berg V, Nieboer E, Odland JO, and Sandanger TM. Repeated measurements
563 of per- and polyfluoroalkyl substances (PFASs) from 1979 to 2007 in males from Northern
564 Norway: assessing time trends, compound correlations and relations to age/birth
565 cohort. *Environ.Int.* 2014;67:43-53.
- 566 Pop VJ, Brouwers EP, Vader HL, Vulmsa T, van Baar AL, and de Vijlder JJ. Maternal
567 hypothyroxinaemia during early pregnancy and subsequent child development: a 3-year
568 follow-up study. *Clin.Endocrinol.(Oxf)* 2003;59:282-8.
- 569 Powley CR, George SW, Ryan TW, and Buck RC. Matrix effect-free analytical methods for
570 determination of perfluorinated carboxylic acids in environmental matrixes. *Anal.Chem.* 1-10-
571 2005;77:6353-8.
- 572 Preau L, Fini JB, Morvan-Dubois G, and Demeneix B. Thyroid hormone signaling during early
573 neurogenesis and its significance as a vulnerable window for endocrine
574 disruption. *Biochim.Biophys.Acta* 27-6-2014;

- 575 Prevedouros K, Cousins IT, Buck RC, and Korzeniowski SH. Sources, fate and transport of
576 perfluorocarboxylates. *Environ.Sci.Technol.* 1-1-2006;40:32-44.
- 577 Stahl T, Mattern D, and Bruun H. Toxicology of perfluorinated compounds. *Environmental Sciences*
578 *Europe* 2011;23:1-52.
- 579 Thibodeaux JR, Hanson RG, Rogers JM, Grey BE, Barbee BD, Richards JH, Butenhoff JL, Stevenson
580 LA, and Lau C. Exposure to perfluorooctane sulfonate during pregnancy in rat and mouse. I:
581 maternal and prenatal evaluations. *Toxicol.Sci.* 2003;74:369-81.
- 582 Ullah S, Alsberg T, and Berger U. Simultaneous determination of perfluoroalkyl phosphonates,
583 carboxylates, and sulfonates in drinking water. *J Chromatogr.A* 16-9-2011;1218:6388-95.
- 584 Vestergren R and Cousins IT. Tracking the pathways of human exposure to
585 perfluorocarboxylates. *Environ.Sci.Technol.* 1-8-2009;43:5565-75.
- 586 Veyhe AS, Hansen S, Sandanger TM, Nieboer E, and Odland JO. The Northern Norway mother-and-
587 child contaminant cohort study: implementation, population characteristics and summary of
588 dietary findings. *Int J Circumpolar Health* 2012;71:18644.
- 589 Wang Y, Rogan WJ, Chen PC, Lien GW, Chen HY, Tseng YC, Longnecker MP, and Wang SL.
590 Association between maternal serum perfluoroalkyl substances during pregnancy and maternal
591 and cord thyroid hormones: Taiwan maternal and infant cohort study. *Environ.Health.Perspect.*
592 2014;122:529-34.
- 593 Wang Y, Starling AP, Haug LS, Eggesbo M, Becher G, Thomsen C, Travlos G, King D, Hoppin JA,
594 Rogan WJ, and Longnecker MP. Association between perfluoroalkyl substances and thyroid
595 stimulating hormone among pregnant women: a cross-sectional study. *Environ.Health*
596 2013;12:76.
- 597 Webster GM, Venners SA, Mattman A, and Martin JW. Associations between Perfluoroalkyl acids
598 (PFASs) and maternal thyroid hormones in early pregnancy: a population-based cohort
599 study. *Environ.Res.* 2014;133:338-47.
- 600 World Health Organization, UNICEF, ICCIDD. Assessment of iodine deficiency disorders and
601 monitoring their elimination;2014. Available from:
602 [http://www.who.int/nutrition/publications/micronutrients/iodine_deficiency/97892415958](http://www.who.int/nutrition/publications/micronutrients/iodine_deficiency/9789241595827/en/)
603 [27/en/](http://www.who.int/nutrition/publications/micronutrients/iodine_deficiency/9789241595827/en/).
604
605

606

SUPPLEMENTAL MATERIAL

607

608 **Assessing the relationship between perfluoroalkyl substances, thyroid hormones and**
609 **binding proteins in pregnant women; a longitudinal mixed effects approach**

610

611 Vivian Berg^{abc}

612 Therese Haugdahl Nøst^{abc}

613 Solrunn Hansen^c

614 Astrid Elverland^a

615 Anna-Sofía Veyhe^c

616 Rolf Jorde^d

617 Jon Øyvind Odland^c

618 Torkjel Manning Sandanger^{bc}

619

620 ^aDepartment of Laboratory Medicine, Diagnostic Clinic, University Hospital of Northern
621 Norway, Sykehusveien 38, NO-9038 Tromsø, Norway;

622 ^bDepartment of Environmental Chemistry, NILU- Norwegian Institute of Air Research, Fram
623 Centre, Hjalmar Johansens gate 14, NO-9296 Tromsø, Norway;

624 ^cDepartment of Community Medicine, Faculty of Health Sciences, University of Tromsø-The
625 Arctic University of Norway, Hansine Hansens veg 18, NO-9019 Tromsø, Norway;

626 ^dEndocrine Research Group, Institute of Clinical Medicine, University of Tromsø-The Arctic
627 University of Norway, Hansine Hansens veg 18, NO-9019 Tromsø, Norway

628

629

| | | |
|-----|--|----------|
| 630 | Contents: | |
| 631 | | |
| 632 | Supplemental material, Table S1 | 3 |
| 633 | Supplemental material, Table S2 | 3 |
| 634 | | |
| 635 | | |
| 636 | | |
| 637 | | |
| 638 | | |
| 639 | | |
| 640 | | |
| 641 | | |
| 642 | | |
| 643 | | |
| 644 | | |
| 645 | | |
| 646 | | |

Table S1

Analytical methods, instrumentation, analytical variation and reference ranges

| Analysis | Method/Instrument | Analytical variation (%) ^a | Reference range ^b |
|-----------------------|---|---------------------------------------|------------------------------|
| TSH ^c | ECLIA/Cobas 8000, e602 | 3.2 | 0.20-4.30 mIU/L |
| T3 | ECLIA/Cobas 8000, e602 | 1.9 | 1.3-3.1 nmol/L |
| T4 | ECLIA/Cobas 8000, e602 | 4.2 | 66-181 nmol/L |
| FT3 ^c | ECLIA/Cobas 8000, e602 | 5.1 | 2.8-7.1 pmol/L |
| FT4 ^c | ECLIA/Cobas 8000, e602 | 3.6 | 9-22 pmol/L |
| Anti-TPO ^c | ECLIA/Cobas 8000, e602 | 9.9 | < 34 IU/ml |
| T-Uptake | ECLIA/Cobas 8000, e602 | 4.8 | 0.8-1.3 TBI |
| TBG | Chemiluminescence enzyme IA/IMMULITE 2000 | 7.5 | 47-45 mg/L |
| TTR ^c | Immunoturbidimetry/Cobas 8000, c702 | 2.6 | 0.15-0.29 g/L |
| Albumin ^c | Colometric assay (bromocresol green)/Cobas 8000, c702 | 1.9 | 39.7-49.4 g/L |

^aThe analytical variation is calculated from all samples run in the diagnostic routine in the year 2013, except for T3, T4, TTR and

Thyroxin binding capacity (T-Uptake) where the analytical variation was calculated during the verification of the analyses for the present study

^bReference ranges (2.5-97.5th percentiles) are those recommended by the manufacturer (Roche) representing a healthy nonpregnant population, except for TBG which is for pregnant women (third trimester)^cThe analysis is accredited according to ISO standard 15189

647

648

Table S2

Covariates evaluated in mixed effects models

| Predictor | Median | Range |
|--|-------------------|-------|
| Age | 32 | 18-43 |
| Children/Parity | 1 | 0-4 |
| Gestational week at visit 1 | 18 | 10-34 |
| Sampling time visit 2 (Days after delivery) | 3 | 1-13 |
| Sampling time visit 3 (Weeks after delivery) | 7 | 3-24 |
| Prepregnancy BMI | 23 | 18-44 |
| BMI at visit 1 | 25 | 18-43 |
| BMI at visit 2 | 27 | 18-45 |
| BMI at visit 3 | 24 | 17-40 |
| Education: Years in school | 16 | 8-20 |
| Thyroxin binding capacity (T-uptake) | Table S1 | - |
| TBG | Table S1 | - |
| TTR | Table S1 | - |
| Albumin | Table S1 | - |
| Blood sampling season | Month of the year | - |
| Iodine sufficiency ^{ab} | Yes/No | - |
| Alcohol during pregnancy | Yes/No | - |
| Smoking | Yes/no | - |

^aParticipants were categorized as iodine sufficient according to a prepregnancy dietary intake of iodine above 150 µg/d (manuscript in preparation by Hansen et al.)^bParticipants were categorized as iodine sufficient according to concentrations of iodine in urine above 150 µg/L (manuscript in preparation by Hansen et al.)

649

650