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### 3 4 **The longer the better: the effect of substrate on sessile biota in Arctic kelp forests**

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6 Concise title: Sessile biota in Arctic kelp forest

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#### 29 30 **Abstract**

31 Kelps are ecosystem engineers and thus enhance biodiversity and subsidize food-webs in nearshore areas. Numerous  
32 studies describing diversity and abundance of biota associated with kelp have focused on sub-tropical and temperate  
33 waters while kelp forests at high latitudes, where kelp is predicted to expand in distribution, remain mostly  
34 unexplored. Kelp forests contribute significantly to regional biodiversity, and associated fauna and the kelp  
35 themselves play ecologically important roles as habitat and feeding areas. Here, we report patterns in diversity,  
36 abundance and seasonal dynamics of fouling organisms associated with different regions of *Saccharina latissima*  
37 and nearby substrates (including stones of the barren ground). The study was conducted in Kongsfjorden, a high  
38 Arctic fjord on the west coast of Spitsbergen; and samples were taken five times between September 2013 and  
39 January 2015. Trends in species richness of epifauna were similar for stones and holdfasts: higher in winter (due the  
40 presence of rare species), and lower in spring and autumn. Species richness and abundance demonstrated a clear  
41 tendency to increase in accordance with substrate stability. Stones housed the most diverse biota compared to living  
42 substrates. Holdfasts demonstrated similar patterns in species composition and abundance as stones due their close  
43 spatial arrangement and presence of demersal larvae in the most of fouling organisms. Similarly, assemblages on  
44 blades in prostrate kelp forests are influenced by the species inhabiting stones of the barren ground. Both biotic and  
45 abiotic factors, including habitat stability and proximity to source populations, contribute to these spatial and  
46 temporal patterns in faunal abundance and diversity. .

#### 47 48 **Keywords**

49 Arctic kelp forest; sessile biota; seasonal dynamics; species richness; abundance; recruitment

#### 50 51 **Introduction**

52 Kelp forests are widely distributed and highly productive marine coastal communities supporting diverse  
53 associated biota (Steneck et al. 2002). Being ecosystem engineers (Jones, et al. 1994), kelps serve as a food source,  
54 provide habitats for associated fauna and flora, and alter the hydrodynamic regime, sedimentation rates, and the  
55 ambient light climate below them. There is an extensive literature describing the influence of different  
56 environmental parameters on the flora and fauna associated with kelp forests, such as latitudinal gradients (e.g.,  
57 Sheppard et al. 1977), seasonality (e.g., Christie et al. 2003), depth (Smith 1996; Smith et al. 1996), wave exposure  
58 (e.g., Schultze et al. 1990; Lippert et al. 2001; Spurkland and Iken 2011), and water flow (e.g., Duggins et al. 1990).  
59 These numerous studies have focused on kelp forests in subtropical and temperate waters, and only few have  
60 investigated kelp beds in the Arctic (Rozycki and Gruszczynski 1986; Lippert et al. 2001; Carlsen et al. 2007;  
61 Wlodarska-Kowalczyk et al. 2009). Most of published studies concern motile epifauna, while sessile organisms in  
62 such assemblage are often neglected (see Wlodarska-Kowalczyk et al. 2009 for a discussion of this).

63 Kelp beds are complex three-dimensional structures including both living and non-living substrates of  
64 varying longevity: ephemeral (blades), short-lived (holdfasts, stipes, understory macroalgae and large sessile  
65 organisms such as some bryozoans and cirripedians), and stable (stones). It was shown for different species of  
66 Laminariales that the number of epifouling species is often lower on blades and stipes than on holdfasts (Norton  
67 1971; Withers et al. 1975; Schultze et al. 1990; Christie et al. 2003, 2009; Blight and Thompson 2008; Włodarska-  
68 Kowalczyk et al. 2009), nevertheless kelp beds were never addressed as entire structure – though every part of  
69 which has a specific fouling assemblage but they affect each other.

70 Sugar kelps (*Saccharina latissima*) form prostrate kelp forests that cover the bottom with their blades  
71 (Steneck et al. 2002). The meristemic growth site is located above the stipe, thus the older part of the blade is shifted  
72 towards its distal end (Hornsey and Hide 1976; Seed and Harris 1980). *S. latissima* and some other species of  
73 Laminariales are known to live up to five years in the Arctic (Lüning 1990; Carlsen et al. 2007), and their holdfasts  
74 and stipes are perennial regions. Blades of *Saccharina* and *Laminaria* spp., on the other hand, are usually thought to  
75 be ephemeral substrate with the longevity of about a year, their growth starts in winter and reaches its peak during  
76 late spring and midsummer. The oldest portion of the blade is lost through abrasion and necrosis which is especially  
77 intense during autumn and winter (e.g., Lüning 1979, 1993; Seed and Harris 1980; Schaffelke and Lüning, 1994;  
78 Makarov and Shoshina 1996; Wiencke and Amsler 2012). Some authors, however, have reported that individuals of  
79 *S. latissima* and *Laminaria digitata* found in summer along the Greenland and Svalbard coasts had perennial blades  
80 (2-3 years), with prominent annual differentiation (Lund 1959; Borum et al. 2002; Carlsen et al. 2007). Based on  
81 these data, therefore, we expect kelp blades to be quite suitable substrate to study recruitment patterns, including  
82 seasonality of the fouling assemblage.

83 Seasonal changes in species richness and diversity of the assemblage associated with kelp have been  
84 demonstrated in many studies performed in temperate waters, but mostly for motile fauna. To our knowledge,  
85 however, there is only a single mention about winter state of kelp forests in the high Arctic (Berge et al. 2015), and  
86 data on seasonal changes in sessile epifauna at high latitudes are lacking since winter sampling under Arctic  
87 condition is logistically difficult. The recent evidence for reproduction in Arctic fauna throughout the year (e.g.  
88 Kuklinski et al. 2013), and activity across the food web in the middle of the Arctic winter (Berge et al. 2015),  
89 suggest seasonal studies can be revealing as to life-cycles of kelp-associated fauna, and food resources for nearshore  
90 food-webs.

91  
92 Based on the relative lack of data on epifauna on kelp blades, and their potentially important roles in  
93 nearshore ecosystem functioning, we assessed patterns of diversity, abundance and seasonal dynamics of fouling  
94 organisms associated with different regions of *S. latissima* and accompanying substrates (including stones of the  
95 barren ground) according to their longevity. Results will fill important knowledge gaps regarding winter kelp  
96 community structure in high Arctic, its seasonal changes and sessile fouling fauna in general. Moreover, such data  
97 create a basis for fruitful comparison for studies concerning settlement patterns of benthic invertebrates on artificial  
98 substrates. These data are needed to establish a baseline against which to evaluate inter-annual variability and  
99 climate change as kelp beds are predicted to expand in a warmer, ice-free Arctic (Krause-Jensen et al. 2012).

## 100 101 102 103 **Material and methods**

### 104 **Study area and sampling**

105 The material was collected in Kongsfjorden (78° 56'N 11° 56'E, near Ny-Ålesund), a glacial fjord on the west  
106 coast of Spitsbergen. Kongsfjorden opens to the West Spitsbergen Shelf and is dominated by the influence of  
107 Atlantic waters (warm and saline), Arctic waters (cold and fresh), and glacial melt with pronounced seasonality  
108 (Cottier et al., 2005; Svendsen et al., 2002; Willis et al., 2006). During the last ten years, most of Kongsfjorden was  
109 free of fast-ice during winter (our observations, see also <http://polarview.met.no>).

110 Samples were taken by SCUBA diving in different seasons – first September 2013, then January, May and  
111 September of 2014, and finally January 2015. Sampling dates were chosen for several reasons: (i) we expected to  
112 experience a range of blade maturity, from little new growth in January to maximum in May (see Wiencke et al.  
113 2007 and references therein); and (ii) these seasons represent a range of food supply and meroplankton abundance  
114 and composition (Willis et al., 2006; Piwosz, 2009; Walkusz et al., 2009; Kuklinski et al., 2013; Stübner et al.  
115 2016).

116 At each season, we took from 7 to 10 sets of samples from 2 m depth. Each set included a stone (hereafter  
117 SL-stone) with one or several individuals of *S. latissima* and understory macroalgae, and a sediment sample  
118 (hereafter, Gr-samples) taken by a cogged grab (sampling area 0.025 m<sup>2</sup>) near SL-stone. All stones found within  
119 sediment (grab) samples are referred hereafter as Gr-stones. In January and May, 2014 and in January, 2015 we also  
120 took six sets of samples at five meter depth. Stones with sessile organisms were selected from Gr-samples and  
121 weighed separately, the rest of Gr-samples were used for grain size analysis which was performed using the standard  
122 dry-sieve shaking method (Eleftheriou 2013).

123 When several individuals of *S. latissima* were attached to a single stone and it was impossible to separate  
124 their holdfasts we regarded the latter as a complex but single structure.

125 The thallus of *S. latissima* possesses several microhabitats corresponding to its region (blade, stipe, holdfast).  
126 In several SL- and GR-samples understory macroalgae were present. Data from each substrate type were analysed  
127 separately since their characters differ and are likely very important for fouling organisms. Substrate types were  
128 divided into three categories according to their longevity: (i) ephemeral (about a year – blades of *S. latissima*); (ii)  
129 short-living (about 3-4 years – holdfasts and stipes of *S. latissima*, understory macroalgae and large bryozoan  
130 colonies); (iii) stable (stones). Only individuals of *S. latissima* that were older than 3 years were taken in the  
131 analysis; age was determined according to Kain (1963).

132 Almost all blades of *S. latissima* in January and all blades in May possessed a very prominent constriction  
133 which was lacking in September (Online Resource 1). This constriction divided a blade in two portions, proximal  
134 and distal, which differed in thickness, colour and abrasion and hence were referred, respectively, as “0+” and “1+”  
135 year-zones. Although both stipes and holdfasts continue to grow (though not so fast) during the whole life-span of a  
136 kelp individual, it is impossible to divide these parts of *S. latissima* thallus into year-zones.

137 For each substrate type (blade, stipe and holdfast of *S. latissima*, thalli of other macroalgae and stones)  
138 surface area was calculated. Blades (different year-zones separately), stipes, and holdfasts (preliminary split at  
139 bifurcations into pieces) were photographed with a size scale, and the diameter of stipes and holdfast pieces was  
140 measured. Area of blades and length of stipes and holdfast pieces were assessed using ImageJ 1.48v package. Thalli  
141 area of other macroalgae was estimated using the same procedure. Surface area of stones and barnacle shells was  
142 calculated using an inelastic net marked in a grid of cm<sup>2</sup>.

143 All sessile macrobenthic organisms were identified to the lowest possible taxonomic level, counted and  
144 weighed (wet-weight) with accuracy of 1 mg (except bryozoans). We measured a square of bryozoan colonies (in  
145 mm<sup>2</sup>) and calculated their weight using species-specific coefficient (Denisenko, 1983). For solitary organisms  
146 density was estimated as the number of individuals per m<sup>2</sup>, for colonial organisms as the number of colonies per m<sup>2</sup>.  
147 Since densities during our survey were rather low there were no difficulties with counting number of hydrozoan or  
148 bryozoan colonies.

149 Whenever possible, we registered the presence of juvenile and adult stages: for bryozoans, ancestrula or small  
150 colonies including ancestrula plus up to four zooids were assigned as juveniles; for spirorbid polychaetes, recruits  
151 with tubes not yet coiled were ascribed as juveniles.

## 152 **Data analysis**

153 To estimate seasonal changes in the percentage of “0+” year-zone we used General Linear Model analysis  
154 (GLM; continuous predictor – total area of blade, categorical factors – season, depth) and a post-hoc test (Fisher  
155 LSD). We only included data from Januaries and May because in Septembers blades were composed of “0+” year-  
156 zone.

157 Species richness (hereafter, SR) was assessed with both three-way main effects analysis of variance  
158 (ANOVA; factors: substrate type, season, depth) based on all samples, and Spearman correlation. For animals, we  
159 calculated the Shannon diversity index ( $H'$ ; log base e) and D-index (effective number of species; see Jost 2006)  
160 based on density.

161 To compare assemblages on different substrates, we estimated density and biomass using the surface areas of  
162 corresponding substrates. To calculate total density and total biomass we used the sum of substrate surface area of  
163 all substrate types in the sample. The frequency of occurrence ( $F$ ) was calculated for each taxon as a percentage of  
164 samples where the taxon was found relative to the total number of samples. Rare species were defined as those  
165 occurring in only one (unique) sample. We identified dominant and subdominant species based on their frequency of  
166 occurrence ( $F \geq 75\%$ ) and density/or biomass ( $N \geq 15\%$  or  $B \geq 15\%$ ) in at least one of the seasons.

167 Since macroalgal communities are usually quantified by biomass only, we used this parameter when we  
168 referred to the fouling assemblage as a whole. When assessing animals' distributions we used both their density and  
169 biomass values. **Density of animals was chosen as a main abundance index when we traced seasonal changes.**

170 ANOSIM (based on Bray-Curtis dissimilarity matrix) was used to test differences in fouling assemblages on  
171 different substrates and in different seasons. ANOSIM tests were based on (i) biomass of all sessile organisms as  
172 well as (ii) density and (iii) biomass of animals only, and were made for two and five meter depth separately.  
173 Understory macroalgae from sediment samples from both depths as well as Gr-stones from five meter depth were  
174 excluded from this analysis due their low frequency of occurrence. SIMPER (based on Bray-Curtis dissimilarity  
175 matrix) analysis was also used to assess taxon contribution to sample heterogeneity at two meter depth.

176 We applied a three-way main effects ANOVA analysis (factors ‘substrate type’, ‘season’, ‘depth’) for  
177 assessing (i) biomass distribution of animals and macroalgae; (ii) biomass of dominant and subdominant species.  
178 Understory macroalgae from sediment samples were excluded from these analyses due their low frequency of  
179 occurrence. Distribution of juvenile densities were tested with use of two-way ANOVA (factors substrate type and  
180 season) based on the data from 2 m depth since we have a full seasonal set of samples. ANOVA tests were based on  
181 fourth-root transformed data and were followed with post-hoc tests (Fisher LSD).

182 PCA analysis was based on fourth-root transformed data of average density of animals on each substrate type  
183 per season. PCA analysis was chosen because it allows to estimate the portion of variation for each factor and to  
184 trace loadings for each species.

186 GLM analysis, ANOVA and post-hoc tests, and Spearman correlations were calculated with use of  
187 STATISTICA 7.0 software package; ANOSIM, SIMPER and PCA analyses were made using PAST 3.12 software  
188 package.  
189

## 190 Results

### 191 Substrate characters

192 The main sediment fraction was categorized as pebbles and cobbles at two meters depth, and silted fine sand  
193 at five meters depth. At five meters depth we found several individuals of *Saccharina latissima* with holdfasts  
194 attached to very fine pebbles and almost entirely embedded within soft sediments. Such holdfasts were rather  
195 thinner and longer than those attached to cobbles and boulders (Online Resource 2).

196 Frequency of occurrence of understory macroalgae (see the list in the legend in Online Resource 3) was  
197 variable and quite low, and usually did not exceed 15-20% for each season, except for *Chaetopteris plumosa* – 30%  
198 and 50% (January 2014 and 2015, respectively) and *Polyides rotundus* – 60% and 40% (September 2014 and  
199 January 2015, respectively). Average surface area of each algal substrate type varied both within and among  
200 seasons, but this value was several times larger for blades of *S. latissima* than for other substrates; average surface  
201 area of stones sampled was comparable to that of holdfasts and stipes (Table 1).

202 Usually in January and always in May, blades of *S. latissima* included both “0+” and “1+” year-zones while  
203 in September all blades lacked constriction and comprised only “0+” year-zone (Online Resource 1). In both  
204 January samplings there were several individuals of *S. latissima* without “0+” year-zone of blade and occurrence of  
205 such individuals was higher at five meters depth (5% and 30% in January 2014 at two and five meter depth,  
206 respectively, and 30% at 5 m depth in January 2015). The percentage of “0+” year-zone varied according to the  
207 season (GLM,  $F_{2,156} = 112.7$ ,  $p < 0.0001$ ) but was independent from total blade surface area ( $F_{1,156} = 0.21$ ,  $p = 0.65$ )  
208 and thus the growth pattern of *S. latissima* was considered to be isometric (Fig. 1). Depth did not affect the  
209 percentage of “0+” year-zone ( $F_{1,156} = 3.42$ ,  $p = 0.07$ ) but analysis revealed a significant interaction (season\*depth;  
210  $F_{2,156} = 5.53$ ,  $p = 0.005$ ). In Januarys at 5 m depth, the percentage of “0+” year-zone was lower than at 2 m (post-hoc  
211 test;  $p < 0.001$ ) but this index was similar at both depths in May (post-hoc test,  $p = 0.18$ ). Post-hoc tests also  
212 demonstrated that the percentage of “0+” year-zone was similar in January 2014 and 2015 ( $p = 0.81$  and  $p = 0.48$  for  
213 two and five meter depth respectively).

214 Rarely, some substrates (more often, “0+” year-zone and stipe of *S. latissima*) were not colonized by sessile  
215 organisms (i.e., they were empty) and their portion varied during the survey. All understory macroalgae were empty  
216 in May, 2014.  
217

### 218 Species richness and species composition

219 In total, 111 taxa of sessile biota were recorded including 80 taxa (57 to species level) of animals and 31 taxa  
220 (23 to species level) of macroalgae; Online Resource 3 gives the full list of taxa. Such a huge gap between the  
221 number of identified species and the number of registered taxa was due to numerous juveniles in the samples,  
222 especially bryozoans, which often possess identification characters only at adult stage.  
223

224 Species richness (hereafter, SR) on every substrate type was highly variable and ranged from 0 to 39 species  
225 per sample. Among animals, Bryozoa was the most species-rich group (56 taxa and 43 species) and among epiflora,  
226 brown algae (Orchophyta: Phaeophyceae – 16 taxa and 11 species) and red algae (Rhodophyta: Floridiophyceae –  
227 11 taxa and 10 species) demonstrated the highest diversity. Mean and total values of SR of macrobenthos were the  
228 lowest on blades and stipes and the highest on SL-stones (Fig. 2, Table 1, see also Online Resources 4 and 5). SR  
229 pattern was similar both for animals and macroalgae and at the both depths (Spearman correlation,  $p < 0.05$  for all  
230 paired combinations).

231 A three-way main effects ANOVA demonstrated that SR of animals was significantly influenced by all three  
232 factors: ‘substrate type’ ( $F_{6,41} = 15.01$ ,  $p < 0.0001$ ), ‘depth’ ( $F_{1,41} = 10.96$ ,  $p = 0.002$ ), and ‘season’ ( $F_{4,41} = 2.9$ ,  $p = 0.038$ ).  
233 Substrate type had the highest effect size (63.2%) while effect sizes of season and depth were rather low (8.2% and  
234 7.7%, respectively). SR of epiflora demonstrated similar patterns, but only substrate type had a significant effect  
235 (three-way main effect ANOVA,  $F_{6,41} = 17.32$ ,  $p < 0.0001$ ; effect size = 68.5%). Post-hoc tests (Fisher LSD,  $p < 0.05$ )  
236 confirmed that SR of sessile macrozoobenthos on SL-stones and SR of macroalgae both on SL- and Gr-stones  
237 differed significantly from other substrates.

238 Seasonal changes in total SR of animals at two meters depth were similar on SL-stones and holdfasts  
239 (Spearman correlation, 0.98,  $p = 0.0004$ ) with the highest values of total SR detected in winter samples both in 2014  
240 and 2015 (Fig. 3; Table 1). Contrary to this, we failed to trace any trend in seasonal fluctuations in SR of animals on  
241 blades and stipes (Fig. 3) and total SR were considerably lower on these substrates (Table 1). Total SR of  
242 macroalgae was the lowest in September, 2013 and January, 2014 but increased to the highest values in the same  
243 seasons a year later (Fig. 3). SR of macroalgae varied on different substrate types with the lowest values registered  
244 for stipes (a single species *Styctiosyphon tortilis* was found only once – in May, 2014).

245 Species composition was highly variable on each substrate type. Almost half (46%) of registered taxa  
246 demonstrated substrate preferences (Online Resource 3). Two species (*Buskia nitens* and *Cerceis spirillum*) were  
247 found only on understory macroalgae while about 66% of taxa never resided on them (the majority of such species  
248 were recorded on stones). Nine species were found only on *Saccharina latissima* (seven only on holdfasts, two both

249 on holdfasts and stipes) while 36 taxa were resided solely on stones, of which two species inhabited only Gr-stones,  
250 21 taxa occurred only on SL-stones, and 13 taxa were found both on SL- and Gr-stones. There were no taxa specific  
251 for blades or/and stipes. We registered 13 taxa residing on all regions of the *S. latissima* thalli and both SL- and Gr-  
252 stones; the rest of the 56 taxa inhabited several substrate types in different combinations.

253 For animals, both Shannon diversity index ( $H'$ ) and D-index (effective number of species) had the highest  
254 values for stones and the lowest for blades and stipes (Table 1), and demonstrated seasonal fluctuations both at two  
255 and five meters depth. A three-way main effects ANOVA revealed that D-index was significantly influenced by  
256 substrate type ( $F_{6,44}=16.94$ ,  $p<0.0001$ ; effect size=72.3%) but was independent of season ( $F_{6,44}=0.56$ ,  $p=0.69$ ) and  
257 depth ( $F_{6,44}=0.63$ ,  $p=0.43$ ). Values of D-index for blades and stipes significantly differed from other substrates (post-  
258 hoc test; Fisher LSD,  $p<0.05$ ).

259 Within a sample, we found up to six species of understory macroalgae with different thallus morphologies  
260 and surface area. SR tended to be higher when more understory algae were recorded, but the frequency of  
261 occurrence of understory algae was too low to address this trend statistically.

## 262 Occurrence and abundance of fouling organisms

263 Total biomass of sessile organisms on blades and stipes was considerably lower than that on stones and  
264 holdfasts (Fig. 4). During our survey we also found noticeable seasonal fluctuations in total biomass, which was  
265 rather high in January, 2014 and September, 2014 and lower in other seasons. It is not surprising that biomass of  
266 macroalgae on both blades and stipes was close to zero and was rather low on holdfasts (except for September, 2013  
267 and January, 2014) compared to stones. Biomass on all parts of *Saccharina latissima* was dominated by animals. A  
268 similar pattern was revealed for SL-stones in September, 2013 and January, 2014 while in the rest of the seasons the  
269 biomass ratio of animals and algae was opposite. Gr-stones were dominated by macroalgae in May, 2014 and  
270 January, 2015 but in September, 2014 biomass ratio of macroalgae and animals was almost equal. Total biomass of  
271 sessile zoobenthos was the lowest in May, 2014 and the highest in January, 2014.

272 ANOSIM tests based on biomass of all sessile organisms, as well as on density and biomass of animals,  
273 indicated that assemblages were significantly different on all substrate types at both depths (statistic R values varied  
274 from 0.5 up to 0.61). Within each substrate type at 2 m depth, seasonal fluctuations were not so prominent (statistic  
275 R values varied from 0.19 up to 0.3), while at 5 m depth the degree of dissimilarity among seasons was higher  
276 (statistic R values varied from 0.42 up to 0.62).

277 We found the highest dissimilarities in biomass distribution among substrates at 2 m (Table 2), while  
278 seasonal differences were somewhat lower (except for holdfasts). Sample heterogeneity was also high and  
279 comparable with seasonal dissimilarities (Table 2).

280 A three-way main effects ANOVA demonstrated that biomass both of animals and macroalgae was  
281 influenced by substrate type ( $F_{5,294}=24.7$ ,  $p<0.0001$ ; effect size = 27.2% for animals and  $F_{5,294}=35.07$ ,  $p<0.0001$ ;  
282 effect size = 35.7% for algae) and season ( $F_{5,294}=6.9$ ,  $p<0.0001$ ; effect size = 6.1% for animals and  $F_{5,294}=3.19$ ,  
283  $p<0.014$ ; effect size = 2.6% for algae). Biomass distribution was independent of depth.

284 Both density and biomass of sessile zoobenthos demonstrated high correlations with substrate type. Thus, the  
285 first component (21.4%) in the PCA analysis (based on average density of animals on each substrate type per  
286 season) corresponded to substrate type since blades and stipes had the lowest scores (from -3.79 up to -2.69) and  
287 SL-stones had the highest one (from 1.6 up to 8.9) while the rest of substrate types occupied intermediate positions  
288 (Fig. 5). This correspondence was also confirmed by the lowest loadings for the species that were often registered at  
289 the blades and stipes (-0.35 and -0.14 for *Celeporella hyalina* and *Circeis armoricana*, respectively), as well as by  
290 the highest loadings for species found mostly on stones (0.89 for *Tegella arctica*; see also the right frame on Fig. 5).  
291 Thus along the first component (from left to right) there was an “age-gradient” of substrates: from ephemeral  
292 (blades) through short-living (stipes and holdfasts), towards stable (Gr- and SL-stones). The second principle  
293 component (13.8%) corresponded to the differences between living and non-living substrates located directly on the  
294 bottom: holdfasts and understory macroalgae had the highest scores (always positive) while scores of Gr-stones  
295 (always) and SL-stones (usually) were negative. The lowest loadings were found for species more frequently  
296 inhabiting stones, while the highest loadings were registered for species usually occupying understory macroalgae  
297 and holdfasts (see the list in the left box on Fig. 5). Together, the first and second components explained about one  
298 third of dispersion, the rest of the components corresponded to the seasonal variation of density of fouling animals  
299 on stones.

300 We found high temporal heterogeneity in the structure of fouling community. Among macroalgae, dominant  
301 species differed within each season on different substrates and changed on each substrate during the survey (Fig. 6,  
302 Table 1, see also Online Resource 4) so we failed to trace any consistent pattern. Among animals, blades and stipes  
303 were dominated in biomass by a single species (*Circeis armoricana* and *Celeporella hyalina*, respectively) during  
304 the survey. Contrary to this, dominant species on other substrates varied in different seasons (Fig. 6, Table 1, and  
305 Online Resources 4 and 5). And since a dominant species in one of the seasons was usually registered as a  
306 subdominant in others, we refer to this collection of taxa as ‘leading species’. All dominant and leading species were  
307 found on at least three substrate types, among which stones and holdfasts were always represented (except for  
308 *Alcyonidium gelatinosum* that was not registered on holdfasts). In addition to this group, we distinguished several  
309 species (*Porella smitti*, *Tricellaria ternate*, *Crisia eburnea*, *Alcyonidium mamillatum*, *Porella minuta*, *Buschiella*  
310

311 *quadriangularis*) with high frequencies of occurrence ( $F \geq 70\%$ ), but with rather low density and biomass. Several  
312 species were rather rare on a given substrate type (Online Resource 3).

313 A three way main-effects ANOVA demonstrated that substrate type and season significantly influenced  
314 abundance indices of dominant and leading species (Table 3), but effect size of substrate type was higher. Only three  
315 species were significantly influenced by depth and, moreover, effect size of this factor was extremely low (about 2%  
316 of explained variation).

### 317 318 **Distribution of juveniles**

319 We found juvenile stages of bryozoans, barnacles and spirorbid polychaetes; and among them, eight taxa  
320 were identified to species level and seven to genus level (Online Resource 3).

321 Juveniles usually demonstrated the same substrate preferences as adults, but in several cases they were also  
322 registered on substrates where adults were absent (e.g., juveniles of *Dendrobeatia* sp. were found on stipes,  
323 *Tubulipora* sp. on stones, and *Balanus balanus* on blades). Juveniles of some taxa were recorded in specific  
324 season(s), e.g., only in September (*Alcyonidium* sp.), both in September and in January (*Circeis armoricana*,  
325 *Tubulipora* sp., *B. balanus*, and *Porella minuta*), both in January and May (*Tricellaria* sp.). Contrary to this,  
326 juvenile stages of some taxa (e.g., *Callopora craticula*, *Cribrilina annulata*, *Celeporella hyalina*, *Crisiella producta*,  
327 *Bushiella* sp., *Dendrobeatia* sp.) were present in all seasons during the survey, and on at least one type of perennial  
328 substrates. A two-way ANOVA (factors 'substrate type' and 'season';  $df_1 = 3$ ,  $df_2 = 2$ ) demonstrated that the  
329 distribution of juveniles of some taxa (*Bushiella* sp., *Dendrobeatia* sp., *C. hyalina*, *C. armoricana*, *C. annulata*,  
330 *C. producta*) was significantly influenced by both factors ( $p < 0.001$  in all cases except for *C. hyalina*, season,  
331  $p = 0.024$ ). Two taxa (*Alcyonidium* sp. and *Tricellaria* sp.) were significantly affected only by season ( $p < 0.005$  in  
332 both cases), and some taxa (*Paradexiospira* sp., *B. balanus*, *C. craticula*, *Patinella* sp.) were only influenced by  
333 substrate type ( $p < 0.005$  in all cases). Distribution of juveniles of several species (*C. hyalina*, *C. armoricana*,  
334 *C. annulata*, *C. producta* and *Bushiella* sp.) were also influenced by interactions of factors (substrate\*season), and  
335 highest densities of juveniles were recorded for different substrates in various seasons.

336 On blades, we recorded juveniles of six taxa (*C. armoricana*, *C. hyalina*, *C. producta*, *Patinella* sp.,  
337 *C. annulata*, *Tubulipora* sp.) but not in May, when they were registered on perennial substrates. During the survey,  
338 their densities were about half that of adults (Fig. 7). In Januarys, juveniles of all species were found within the '1+'  
339 year-zone of blades.

## 340 341 342 **Discussion**

### 343 **Methodological issues**

344 Traditionally, abundance of sessile biota is estimated using percent cover as a main index. The principal  
345 advantage of this method consists in possibility to describe both community structure and abundance without time-  
346 consuming procedure of calculating density and biomass, particularly since colonial animals present challenges  
347 here. Doubtless, spatial competition on stable substrates may be very high and sometimes it can be difficult to count  
348 the number of colonies of hydrozoans or bryozoans. However, percent cover data are only comparable when  
349 substrates are more or less equal in size, which is not always the case. Therefore, different methods for assessing  
350 species number and/or abundance of epibionts were used in previous studies, such as number of epifaunal organisms  
351 per plant (Norton 1971), area of substrate surface (e.g., Rozycki and Gruszczynski 1986), weigh of basibiont (e.g.,  
352 Kain 1971; Lippert et al. 2001), volume of basibiont (e.g., Edwards et al. 1980; Smith et al. 1996; Christie et al.  
353 2003, 2009; Blight and Thompson 2008) or length of stipe (e.g., Christie et al. 2003). Such a diversity of methods  
354 hinders the comparability of different studies of species abundance. Most of these methods are applicable to the  
355 study of abundance of mobile epifauna, but one of the most important factors for studying the distribution of sessile  
356 organisms is the area of substrate surface. This method is also of particular importance when comparing diversity  
357 and abundance across different substrates.

### 358 359 **Phenology of blade extension**

360 We documented growth of *Saccharina latissima* in Kongsfjorden in early January (though not all individuals  
361 had started to grow by mid-January). The blade at this time includes a small '0+' year-zone and a large '1+' year-  
362 zone. Growth slows down in autumn and the distal part of the blade is destroyed due to physical forces, so by the  
363 end of September the blade is composed only of '0+' year-zone.

364 Our results are in agreement with previously reported data from European coast (e.g., Lüning 1993;  
365 Schaffelke and Lüning 1994; Makarov and Shoshina 1996). Carlsen et al. (2007), however, reported 3-year-old  
366 blades both of *S. latissima* and *Laminaria digitata* from the same locality (Kongsfjorden, west Spitsbergen), and  
367 blades of 2-3 years were also reported from Greenland coast (Lund 1959; Borum et al. 2002). Such a difference is  
368 quite surprising and difficult to explain. Perhaps such individuals were rather rare and were collected in sheltered  
369 localities and/or deeper where blade abrasion is rather slow. In addition they probably were a specific target in these  
370 studies while we sampled randomly and did not find such old blades. However, since in all these cases material was  
371 collected in high Arctic glacial fjords with mid-summer decreases in light and nutrient loads due to glacial runoff,  
372 one can speculate that such abrupt changes in abiotic drivers might cause non-cyclic slowing down or even stoppage  
373 of blade growth, resulting in constriction formation. Although such a phenomenon was never reported for

374 Laminariales species, interaction of environmental factors (such as irradiance, temperature, nutrients, etc.) can affect  
375 algal growth in ways that are not fully understood (reviewed in Wiencke et al. 2007, 2009; Gomez et al. 2009;  
376 Wiencke and Amsler 2012).

### 377 **Assessment of species richness**

378 Epifauna associated with all Laminariales species often demonstrate large similarities across studies,  
379 especially regarding SR and species composition (Schultze et al. 1990; Lippert et al. 2001; Carlsen et al. 2007;  
380 Wlodarska-Kowalczyk et al. 2009). Based on this, we expected to find similar levels of species diversity in our  
381 survey and earlier studies, especially those conducted in the same region and locality (Kongsfjorden, west coast of  
382 Spitsbergen). Data from two previous studies (Lippert et al. 2001, Carlsen et al. 2007) allow a detailed comparison  
383 with our results for sessile epifauna. Strikingly, the number of zoobenthic taxa recorded on *S. latissima* in our survey  
384 was at least two times higher than in either of the other studies. Nevertheless, the higher level taxonomic  
385 compositions found the present study and reported by Lippert et al. (2001) are consistent, and the most abundant  
386 taxonomic groups are similar, e.g. we registered 80% of the bryozoan species listed by Lippert and coauthors. On  
387 the other hand, species composition reported by Carlsen et al. (2007) differs significantly: sponges, barnacles and  
388 hydrozoans were not mentioned in their study, only about 50% of bryozoan species coincided, and spirorbid  
389 polychaetes were represented by other species. Data reported by Rozycki and Gruszczynski (1986) for four species  
390 of overstory kelp collected on the west coast of Spitsbergen revealed a similar pattern: SR in our survey is three  
391 times as high while taxonomic composition is comparable. Such differences in both cases are most likely due the  
392 combination of two factors: small numbers of samples in these studies (8, 20 and 56, respectively) and spatial  
393 heterogeneity. Even at small scales (one to few meters) it is typical for the distribution of biota and assemblage  
394 composition to vary considerably (e.g., Arroyo et al. 2004; Anderson et al. 2005; Kuklinski et al. 2006; Yakovis et  
395 al. 2008; Miguel et al. 2010).

396 In our survey we found 111 taxa in 68 samples including 110 individuals of *S. latissima* which is quite far  
397 from assessing true species richness. Even in an extensive study of motile and sessile epifauna associated with  
398 macroalgae, and based on rather numerous samples (in total, 403 samples including 356 individuals of three species  
399 of overstory kelp), the observed number of motile and sessile taxa was 208, a value much lower than Chao2  
400 estimated value (Wlodarska-Kowalczyk et al. 2009). This research was also performed in Svalbard waters  
401 (Hornsund), and demonstrated diverse sessile fauna among which Bryozoa and Hydrozoa were the most species rich  
402 groups, with 70 and 37 species, respectively (compared to 56 and 8 species in the present study). Unfortunately,  
403 Wlodarska-Kowalczyk and coauthors did not provide any details of epifauna distribution on a given species of  
404 macroalgae (or at least mention them to be located either on overstory or understory kelp) and this prevents us from  
405 further comparison. Besides differences in number of samples, such a huge gap in the species richness probably  
406 resulted from different sampling depth and larger number of sites sampled.

407 Habitat complexity and structural heterogeneity are well known to beget diversity and increase richness and  
408 abundance in different habitats (e.g., Dean and Connell, 1987; McCoy and Bell 1991; Gee and Warwick 1994;  
409 Knowles and Bell 1998; Kelaher and Castilla 2005; Matias et al. 2010; Kovalenko et al. 2012; St.Pierre and  
410 Kovalenko 2014; Loke and Todd 2016). Abundance, biomass, and diversity of motile epifauna are directly related to  
411 algal species composition (e.g. Attrill et al. 1996; Parker et al. 2001; Christie et al. 2003; Norderhaug et al. 2007).  
412 As shown by Leclerc et al. (2015), the abundance of mobile fauna in kelp epiphytes may interfere with habitat  
413 complexity on the understory and kelp size-density structure (density of adults hosting abundant epiphytes).  
414 Armitage and Sjøtun (2016) demonstrated that shared macroalgal epiphytes lead to higher similarity between the  
415 zoobenthic communities associated with *Codium fragile* and *Fucus serratus*. Unfortunately, such data lack for  
416 sessile epifauna, anyway, one can speculate that (i) substrate surface area available for recruits increases with the  
417 habitat complexity; (ii) understory macroalgae alter hydrodynamic regime within the bottom water layer making it  
418 more turbulent which is supposed to facilitate in the food uptake by suspension feeders (e.g., Shimeta and Jumars  
419 1991; Vogel 1994).

420 Thus, the higher values of SR and somewhat different species lists reported by Wlodarska-Kowalczyk and  
421 coauthors (2009) may be due to higher habitat complexity and structural heterogeneity owing to the presence of  
422 more numerous understory macroalgae with diverse thallus structure in their samples. Understory macroalgae in our  
423 survey were dominated by filamentous and branching forms (see species list in Online Resource 3), and their  
424 morphology and limited surface area impede settlement by sessile biota. Although the occurrence of understory  
425 macroalgae in our samples was rather low, we noted that SR tended to be higher when more understory algae were  
426 recorded. This observation, and the quite different species lists of understory algae in the present study and that  
427 reported by Wlodarska-Kowalczyk et al. (2009), we suggest that SR and diversity of sessile biota associated with  
428 overstory kelp is also highly influenced by species composition and abundance of understory macroalgae.

### 429 **Effect of substrate type**

430 It has been shown that the number of fouling species on different species of Laminariales is often lower on  
431 blades and stipes than on holdfasts (Norton 1971; Withers et al. 1975; Schultze et al. 1990; Christie et al. 2003,  
432 2009; Blight and Thompson 2008; Wlodarska-Kowalczyk et al. 2009); and our data confirm this observation (see  
433 Fig. 2). In the present study, SR, density and biomass of epibionts on blades and stipes were two orders of  
434 magnitude lower than on other substrate types (see Fig. 2, 3, 4). According to our results, substrate type, reflecting  
435  
436

437 longevity, surface characteristics, and likelihood of disturbance, among other factors, more strongly influenced  
438 distribution of fouling biota than seasonality or depth. This hierarchy of factors was found for SR, diversity, and  
439 abundance indices of the whole assemblage, as well as for dominant and leading species (see Table 3, Fig. 6).

440 Numerous studies have demonstrated the effect of substrate type on SR, diversity, species composition of  
441 epifauna (e.g., Seed and O’Connor 1981; Lippert et al. 2001; Kuklinski and Barnes 2005; Villegas et al. 2008). Our  
442 data are in agreement with them. In the present study, SL-stones housed the most diverse biota compared to living  
443 substrates (97 and 61 taxa, respectively, out of 111 – see Online Resource 3), which is probably due to higher  
444 substrate stability and, perhaps, surface quality, in particular, rugosity. Numerous fouling settlers prefer small  
445 cavities, and rough and convex surfaces (e.g., Seed and O’Connor 1981; Ward and Thorpe 1989; Kuklinski and  
446 Barnes 2005). Gr-stones demonstrated a lower ratio of substrate-confined species and lower values of total SR  
447 compared with SL-stones, making GR-stones more similar to ephemeral substrates. Although species composition,  
448 density, and biomass of fouling biota on Gr-stones were similar to those on SL-stones, their values were highly  
449 variable during the survey so that Gr-stone samples formed an elongated cluster along first component axis in PCA  
450 analysis (see Fig. 5) which corresponded with an “age-gradient” of substrate type. Such features are very likely to be  
451 explained by somewhat smaller size of some Gr-stones (see Table 1) and their lower weight resulting in more  
452 frequent overturning caused by strong wave action, and, in addition, smaller surface area available for recruits. We  
453 have indirect confirmation of disturbance (overturning) frequencies of small Gr-stones during our survey as the  
454 presence of dead bryozoan colonies with destroyed frontal walls of zooids which were more numerous in samples  
455 collected in September, 2014 and January, 2015, sampling periods that were preceded by strong storms. These  
456 results confirm the data of numerous previous studies reported that the stone size is of prime importance for  
457 colonization patterns (e.g., Osman 1977; Sousa 1979; Wilson 1987; Kuklinski et al. 2006; Grzelak and Kuklinski  
458 2010). As suggested before, assemblages on small stones are likely in a state of constant transformation and are  
459 composed of opportunistic species with very short life cycles (Kuklinski et al. 2006; Kuklinski 2009).

460 Bryozoans in Svalbard waters show no indication of being substrate-specific (Kuklinski and Barnes 2005).  
461 Contrary to this, some bryozoan species in this study were only found on a single substrate type (Online Resource  
462 3). Realizing that we had a smaller number of samples, a limited depth range of sampling, and a single sampling  
463 site, we do not suggest such species are substrate-specific, but instead designate them as ‘substrate-confined’. The  
464 proportion of such substrate-confined species was higher for stones compared to *Saccharina latissima* and, they  
465 resided on perennial regions of kelp thallus and not on blades. Absence of substrate-confined species on blades  
466 probably resulted from the ephemeral and unstable nature of the substrate, which prevents the majority of sessile  
467 animals from finishing their life-cycles.

468 Although the percentage of empty substrate was rather low during the survey (see Table 1), we never  
469 registered examples of spatial competition (overgrowth), even on stones. This suggests that epifaunal communities  
470 on all substrates were in early stages of colonization.

471 Among animals, we can distinguish two groups of dominant and leading species: the first was specific for  
472 blades+stipes and the second for stones, while the list of leading species on holdfasts was a mixture of both groups  
473 (see Table 1, Online Resources 4 and 5). Since the same trend was also found for species composition, we suggest  
474 that assemblage of epibionts on holdfasts is highly influenced by that on stones.

475 Some dominant species are consistent with those previously mentioned for Svalbard waters. Thus,  
476 *Celeporella hyalina* was reported to dominate macroalgae (Lippert et al. 2001; Kuklinski and Barnes 2005) and in  
477 the present study this species dominated on stipes of *S. latissima* and was a subdominant on blades and holdfasts,  
478 and on understory kelp. The bryozoan assemblages on shallow subtidal stones in Arctic waters were dominated by  
479 *Harmeria scutulata* (e.g., Barnes and Kuklinski 2003; Kuklinski and Barnes 2005; Kuklinski et al. 2006; Kuklinski  
480 2009). During our survey, *H. scutulata* was dominant or subdominant species on SL-stones, and only once took a  
481 subdominant position on Gr-stones. The other leading species on stones (*Cribrilina annulata*, *Tegella arctica*,  
482 *Electra arctica*) have also been shown to be abundant in some studies, but contrary to our data, their density was  
483 reported to be at least an order of magnitude lower compared to that of *H. scutulata* (e.g., Barnes and Kuklinski  
484 2003; Kuklinski and Barnes 2005; Kuklinski et al. 2006; Kuklinski 2009).

485 In the present study, SR and abundance demonstrated a clear tendency to increase in accordance with  
486 substrate longevity, from ephemeral (blades) through short-lived (holdfasts and understory macroalgae) towards  
487 stable (stones) substrate. Short-lived substrates are quite specific: they are perennial and grow during their life-span,  
488 increasing their surface area which provides empty space for recruits. On the other hand, different parts of the  
489 *S. latissima* thallus vary in rigidity; morphology (shape); and, as a consequence, local water flow velocity and  
490 regime; and also exometabolite composition (some of them may attract recruits and some may prevent them from  
491 settlement). All these points may affect the attractiveness of different parts of kelp thallus and result in higher  
492 diversity and abundance of biota associated with the holdfast.

#### 493 **Seasonal dynamics and recruitment**

495 During the survey, we found the highest values of SR of sessile zoobenthos in both Januaries but we  
496 registered very low values of epifloral SR in January, 2014 and the highest in January, 2015. Unfortunately there are  
497 scarce data on seasonal changes of sessile biota associated with kelp and mainly they concern temperate waters.  
498 Contrary to our findings, Withers et al. (1975) mentioned richer epifauna in summer comparing to winter, and  
499 Christie et al. (2003) specified that SR was not affected by seasonal changes but abundance of epifauna was

500 significantly related to season. Carlsen et al. (2007) reported that total SR of epifauna on blades of *Saccharina*  
501 *latissima* and *Laminaria digitata* was significantly higher in May than in late August (considering only sessile  
502 species from their list, 15 and 6 taxa, respectively). We sampled in May and late September and found an opposite  
503 pattern (Fig. 3). There are several possible explanations for these dissimilarities: (i) interannual variability; (ii)  
504 spatial heterogeneity; (iii) the difference of one month between autumn sampling dates (we sampled a month later)  
505 resulting in colonization of blades by newly settled recruits which probably had not yet appeared in August.

506 SR of epifauna in the present study demonstrated seasonal trends that were similar for SL- and Gr-stones and  
507 holdfasts: it was higher in both Januarys and lower in May and both Septembers with comparable values of the same  
508 seasons in subsequent years (see Fig. 3). Collation of species lists for these substrates reveals that the highest values  
509 of SR in Januarys were due the presence of rare species (Online Resource 3), which most likely results from their  
510 settlement between September and January. The decrease of SR in May and September may be caused by the effect  
511 of abiotic factors (e.g., abrupt drop of salinity and increase of sedimentation rate in summer are quite typical in  
512 glacial bays, and runoff from the Red River in autumn flows directly over the kelp bed we studied), juvenile  
513 mortality and/or predation impact.

514 Seasonal fluctuations in total biomass of fouling assemblages were connected with different relative  
515 importance of macroalgae and zoobenthic organisms' abundance in different seasons. The fouling community in  
516 May and September, 2014 was dominated in biomass by understory macroalgae due their intensive growth during  
517 spring and summer. Low values of macroalgal biomass recorded in September, 2013 are probably explained by  
518 interannual variability and spatial heterogeneity of a small scale. The same factors seem to be more probable to  
519 affect also the biomass ratio of macroalgae and animals in corresponding seasons (Septembers and Januarys). Low  
520 frequency of occurrence of all species of understory macroalgae in the samples and short period of our observations  
521 prevent us from any further speculation.

522 Seasonal variability in the abundance of juvenile forms on holdfasts and stones registered in the present study  
523 (see Fig. 6) likely reflects spatial patterns in adult populations, seasonal recruitment, and/or differential juvenile  
524 mortality. Unfortunately, data on life cycles or life history of many benthic species from the Arctic are limited, but  
525 life-history strategies of dominant taxa in this study varied considerably. Settlement peaks of some taxa in mid-  
526 winter, and lack of distinct seasonality in other taxa, highlight the little-recognized complexity of larval strategies  
527 for Arctic benthos, where larval appearance is often presumed to be associated with the spring phytoplankton bloom.

528 We found distinct seasonality in occurrence of some juvenile stages that likely reflects seasonality in the  
529 reproduction of such species. Similarly, those species with juveniles found in both in September and January likely  
530 have either a long recruitment period with very slow growth, or stoppage of growth when food supply is low. The  
531 presence of juveniles of some taxa in all seasons at least on one type of perennial substrates indicates year-round  
532 reproduction in such species, which was also confirmed by presence of larvae in the ovicells of some bryozoans  
533 (e.g., *Callopora craticula*, *Cribrilina annulata*, *Celeporella hyalina*) during every sampling period. These species  
534 are known to be opportunists with fast growth. Absence of juveniles of these species on blades in May (see Fig. 7)  
535 despite their presence on perennial substrates at the same time let us suggest a year-round reproduction of these  
536 species and, at the same time it is very likely that their larvae avoid blades during spring presumably due a specific  
537 exometabolites composition and biofilm on them. Many authors pointed the differences of epifaunal assemblages  
538 associated with different species of macroalgae (e.g., Seed and O'Connor, 1981; Blight, Thompson, 2008; Parker et  
539 al., 2001; Christie et al., 2009; Grzelak and Kuklinski, 2010). For example, even for kelp species with similar thallus  
540 structure difference in SR of epibionts was reported; this was explained by higher production of antifouling  
541 chemicals by *Laminaria ochroleuca* compared to *L. digitata* (Blight, Thompson, 2008). Later, Hellio et al. (2000,  
542 2002) found that exudates of *L. ochroleuca* had higher levels of antimicrobial activity comparing to *L. digitata*,  
543 particularly against microalgae and marine fungi which together with bacteria are significant contributors to biofilm  
544 formation and thus affect the subsequent attachment of other epibiont organisms.

545 Only a few of the animals recorded in our study (bryozoan *Electra arctica*, cirripedian *Balanus balanus*) have  
546 planktotrophic larvae which spend quite long periods in the water column (pelagic larvae), while the rest of sessile  
547 epifauna – spirorbid polychaetes, ascidians, sponges, hydrozoans and the rest of bryozoans – have lecithotrophic  
548 larvae. Spirorbid larvae are regarded as pelagic in spite of their very short swimming period (e.g., from 15 minutes  
549 up to few hours reported for *Spirorbis spirorbis* – Knight-Jones 1951, 1953; see also an extensive review in  
550 Kupriyanova et al. 2001). However, it was shown that presence of a favourable substrate reduced the duration of  
551 pelagic phase, for example, up to 91% of released larvae of *Spirorbis rupestris* metamorphosed even without  
552 becoming pelagic (Gee 1963). Swimming period of both sponge and solitary ascidian larvae varies in different  
553 species within 2-24 hours (e.g., Berrill 1930; Svane and Young 1989; Ereskovsky, 2010), and although they are  
554 assumed to be pelagic, they have not been registered in meroplankton samples in Kongsfjorden (e.g., Piwosz, 2009;  
555 Walkusz et al. 2009; Kuklinski et al. 2013; Grenvald et al. 2016; Stübner et al. 2016). In the few studies concerning  
556 behaviour of hydrozoan planulae, it was demonstrated that they have short swimming or crawling period (usually  
557 10-12 hours but up to 2-3 days – Williams 1965; Sommer 1992; Orlov 1996, 1997; Migotto 1998) and only very  
558 rarely can they be transported by ambient currents (see Migotto 1998). Bryozoan lecithotrophic larvae also have a  
559 short swimming period which they spend within the bottom water layer (for different bryozoan species, from several  
560 hours up to 3-4 days – Shunatova, unpublished data; see also Ryland 1974, 1976; Reed 1991). Pearse (1969) and  
561 Mileikovsky (1971) termed such larval type as demersal and both of them suggested that it may be more common  
562 for benthic invertebrates than previously thought (see also Gerrodette 1981).

563 Based on these data, it is likely that the distance traveled by demersal larvae does not exceed a few meters  
564 (and usually is less than one), so that they are more or less restricted to the area where maternal colony (or  
565 individual) is located. This perspective can help to explain (i) the higher similarity of SR, species composition, and  
566 abundance of sessile epifauna associated with SL-stones and holdfasts, and (ii) the lower SR and abundance on  
567 blades and stipes. This is consistent with observations of Christie et al. (2003) who found that epifauna on the stipe  
568 of *Laminaria hyperborea* was richer in regions closer to the holdfast. Meyer et al. (2017), also noted very high  
569 spatial heterogeneity of settlement on panels deployed for different periods in a 12-month period in several Svalbard  
570 fjords, which they interpreted to reflect short larval periods and, thus, close relationships between settlers and nearby  
571 adult residents. Thus, larval characteristics are likely of the primary importance for colonization patterns of fouling  
572 assemblages. The presence of demersal larvae in most registered species and close spatial arrangement of holdfasts  
573 and SL-stones, as well as blades and stones of the barren ground, result in similar patterns in species composition  
574 and abundance for each pair of substrates.

## 575 576 **Conclusions**

577 This study demonstrated that SR, diversity, and abundance were higher on holdfasts than on blades or stipes  
578 of kelp in a Svalbard fjord. Substrate type, reflecting longevity, surface characteristics, and likelihood of  
579 disturbance, among other factors, more strongly influenced distribution of fouling biota than seasonality or depth.  
580 This hierarchy of factors was found for SR, diversity, and abundance indices of the whole assemblage, as well as for  
581 dominant and leading species. Among numerous differences between discussed, substrate types their longevity had  
582 quite prominent effects on species composition. Additionally, presence of demersal larvae in most fouling species  
583 and close spatial arrangement of holdfasts and SL-stones as well as blades and stones of the barren ground result in  
584 similar patterns in species composition and abundance for each pair of mentioned substrates. Based on the data  
585 reported for motile epifauna and our qualitative observations, we suggest that SR and diversity of sessile biota  
586 associated with overstory kelp is also highly influenced by species composition and abundance of understory  
587 macroalgae. But this point deserves to be a subject of a separate study.

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## 598 599 600 **Conflict of interest**

601 The authors declare that they have no conflict of interest.

## 602 603 **Ethical approval**

604 All applicable international, national, and/or institutional guidelines for the care and use of animals were followed.

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### Figure captions

**Fig. 1** Portion of “0+” year-zone of the blades of *Saccharina latissima* (percentage of total surface area of blade)  
*Error bars* show standard error.

**Fig. 2** Average and total species richness on different substrate types (all seasons together)  
*Error bars* show standard error.

**Fig. 3** Seasonal dynamic of species richness in zoo- and phytobenthos at 2 m depth

**Fig. 4** Seasonal dynamic of total biomass of macrobenthic organisms at 2 m depth

**Fig. 5** Biplot of substrates distribution in Principal Component Analysis (based on four-root transformed data on density of macrozoobenthos at 2 m depth)  
Dotted lines correspond to loading vectors of species that significantly contributed to the first and second components.

**Fig. 6** Seasonal dynamic of total biomass of dominant and leading species (left column – animals, right column – algae) at 2 m depth

**Fig. 7** Distribution of adults and juveniles of leading species on blades of *Saccharina latissima*  
“0+” and “1+” year-zones of the blades of *Saccharina latissima* are marked as “0+” and “1+”, correspondingly.  
*Error bars* show standard error.