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Which ecological factors influence the level of intraspecific diversity within post-glacial fishes? A case study using *Coregonus* **and** *Salvelinus*

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Abstract

- 1. While it is likely that ecological context is important, the factors that facilitate and maintain variable levels of intraspecific diversity in Salmonidae fishes across environments remain unclear.
- 2. Using a meta-analysis of sympatric ecotype assemblages from two salmonid genera—*Salvelinus* and *Coregonus*—we evaluated the importance of ecological factors determining the number of sympatric ecotypes (i.e. 2–7) and the level of trait divergence between them.
- 3. We found that ecotype diversity increased with lake depth and surface area in both *Coregonus* and *Salvelinus*. Further, diversity in *Coregonus* increased with latitude, while the number of ecotypes in *Salvelinus* assemblages was linked to climatic seasonality.
- 4. In comparing the two genera, we found elevated divergence in traits related to ontogeny (i.e. age and body shape) in *Salvelinus* and gill raker count in *Coregonus.* Trait divergence in life history traits (i.e. age and body length) in *Salvelinus* increased with seasonality, whereas contrasting relationships of latitude to body length and gill rakers were found in *Coregonus*. We also found similar levels of divergence in trait variance in the two genera, suggesting that among-ecotype differences in phenotypic variability are not more common in one genus than the other.
- 5. Overall, ecosystem characteristics, including lake location, climate and morphometry, are clearly important for where these genera have diversified, but the variables that are most closely associated with intraspecific diversity differ between the two genera studied and depend on whether diversity is quantified using number of ecotypes or trait divergence.

KEYWORDS

ecosystem, ecotype, lake, meta-analysis, Salmonidae

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1 | **INTRODUCTION**

Ecologically and phenotypically distinct ecotypes can coexist in sympatry, even at very early stages of evolutionary divergence (Schluter, [2000](#page-9-0); Skulason & Smith, [1995](#page-9-1)). Two or more ecotypes are usually only able to coexist if they occupy distinct ecological niches (Germain et al., [2018](#page-8-0)). Therefore, irrespective of whether ecotypes diverged in sympatry or came into secondary contact after a short period of evolution in allopatry, ecosystem characteristics are likely to be critical for where sympatric ecotype diversity can or cannot persist (Seehausen & Wagner, [2014](#page-9-2); Skulason & Smith, [1995\)](#page-9-1). The distribution of ecotype diversity is geographically uneven, suggesting that ecotype coexistence may be possible in some environments but not others (Wagner et al., [2014](#page-9-3)). While it is likely that ecosystem characteristics are important, the specific factors that determine the level of intraspecific diversity that can arise and be maintained in one environment compared to another remain unclear.

Several explanations have been put forward to elucidate why ecological diversity evolves in some places and some lineages but not in others. Species depauperate environments and available sources of food or predator-free space are thought to facilitate in situ diversification and colonization of diverging forms of the same species, particularly in lineages with trait plasticity or sufficient standing genetic and morphological variation to exploit them (Stroud & Losos, [2016](#page-9-4); Yoder et al., [2010](#page-9-5)). Abiotic ecosystem characteristics, such as latitude and climate, can have strong and sometimes opposing effects on rates of diversification and levels of sympatric diversity (Rabosky et al., [2018](#page-9-6)). In environments with distinct boundaries, such as lakes or islands, the size of the environment can impact both the number of intraspecific sympatric lineages and the distribution of traits within those lineages (Koene et al., [2020](#page-8-1); Losos & Schluter, [2000](#page-8-2); Recknagel et al., [2017](#page-9-7); Wagner et al., [2014](#page-9-3)). Ecological diversity can be quantified in multiple ways, including the number of diverging ecotypes, divergence in trait mean values and divergence in trait variability (Violle et al., [2012](#page-9-8)). To fully understand how ecosystem characteristics shape processes of diversification and community assembly, the role of these ecological factors in facilitating different forms of diversity at early stages of divergence needs to be evaluated.

Post-glacial lakes are hotspots for recent divergent evolution and assembly of ecotypes, resulting in co-existence of ecotypes that have not fully speciated. Sympatric ecotypes in post-glacial lakes have repeatedly evolved in several families of fish, most commonly occurring as a pair of ecotypes in which one ecotype occupies a littoral (benthic) habitat and the other occupies a pelagic (limnetic) habitat (Robinson & Parsons, [2002](#page-9-9); Schluter & McPhail, [1993](#page-9-10)). Two salmonid genera with a circumboreal distribution—*Salvelinus* and *Coregonus*—exhibit an unusual level of ecotype diversity, as they commonly occur in assemblages of up to seven sympatric ecotypes (Smith & Skúlason, [1996](#page-9-11)). These multi-ecotype assemblages are broadly distributed, occupying lakes that span temperate deciduous, boreal, subarctic and arctic biomes (Bernatchez et al., [2010](#page-7-0); Jacobs et al., [2020](#page-8-3); Siwertsson et al., [2010](#page-9-12)). These genera exhibit exceptional ecological and phenotypic diversity, both within and among

lakes (Chavarie, Adams, et al., [2021](#page-7-1); Klemetsen, [2013](#page-8-4); McPhee et al., [2012](#page-8-5)). Sympatric ecotypes in *Salvelinus* and *Coregonus* additionally vary in the mechanisms underlying ecological divergence, with trait differences arising from phenotypic plasticity in some cases, and genetic differentiation maintained by reproductive isolation in others. Therefore, they also differ in their degree of reproduc-tive isolation (Hendry, [2009](#page-8-6)). Despite the variability in mechanisms of differentiation among ecotypes, sympatric ecotype diversity has evolved and assembled numerous times since the formation of the post-glacial lakes in which they occur (Hudson et al., [2011](#page-8-7); Jacobs et al., [2020](#page-8-3); Turgeon & Bernatchez, [2003](#page-9-13)). Although there are broad similarities between these genera, the traits and ecological drivers involved in ecotype divergence may differ between them.

We used the repeated evolution and assembly of varying levels of diversity in *Salvelinus* and *Coregonus* to assess what ecological factors are related to the number of sympatric ecotypes and the level of trait divergence between them. Specifically, we used a metaanalysis of morphological, life history and diet data from *Salvelinus* and *Coregonus* assemblages with a circumboreal distribution. First, we assessed whether abiotic features of lakes, including climate, location and lake dimensions, are related to the number of ecotypes that can coexist in sympatry in each genus. Second, we asked whether the extent of trait divergence among sympatric ecotypes varies with ecological factors, including ecotype richness and lake features, or between genera. Finally, we determined whether patterns of among-ecotype divergence in trait means match patterns of among-ecotype divergence in trait variance.

2 | **METHODS**

2.1 | **Data collection**

We performed a systematic literature review to identify papers with data on wild caught fish from lakes supporting two or more coexisting ecotypes, closely related species, or taxa (hereafter referred to as 'ecotypes') forming separate ecological and morphological clusters, from the genus *Salvelinus* or *Coregonus* (see details in Blain et al., [2023](#page-7-2); Appendix [S1](#page-12-0)). Following the full-text review, we retained 127 studies with data from 131 assemblages and included these in the meta-analysis (Figure [S1](#page-12-0)). We extracted usable trait data for 127 assemblages (Table [S4](#page-12-0)). For 124 lakes, we had both abiotic predictor variables and a count of the number of ecotypes in the lake, along with usable trait data for 91 of these lakes.

We recorded data for diet breadth, upper jaw length, lower jaw length, pectoral fin length, head length, eye diameter, snout length, body shape principal components 1 and 2, asymptotic length, depth of capture, dorsal fin length, anal fin length, gill raker length, gill raker count, gape width, caudal fin length, pelvic fin length, body length, age, $\delta^{13}C$ and $\delta^{15}N$. For each of these traits, we recorded the mean, sample size and standard error for each population. We also recorded the latitude, longitude, maximum depth and surface area of each lake. Where necessary, these physical features were extracted

from additional sources such as government databases and limnology studies. To capture lake morphometry, we estimated the depth ratio as 50 × maximum depth × $\frac{\sqrt{\pi}}{\sqrt{2}}$ $\frac{v}{\sqrt{\text{surface area}}}$ (Wetzel & Likens, [2000](#page-9-14)). We extracted annual mean temperature, temperature seasonality, annual mean precipitation and precipitation seasonality from the WorldClim database using each lake's latitude and longitude (Fick & Hijmans, [2017](#page-8-8)). In WorldClim, temperature seasonality is estimated as the standard deviation of annual temperature multiplied by 100 and precipitation seasonality is the coefficient of variation for precipitation. We divided the extracted temperature seasonality values by 100, so that the estimates in our data set had a similar range to those in the precipitation data set, resulting in values that ranged from 36.7 to 177. Values for precipitation seasonality in our data set ranged from 10 to 94.

2.2 | **Effect size estimates for differences among all ecotypes**

We used standardized metrics to compare ecotypes' phenotypic differentiation among lakes, as there is variation in the specific methodology with which different researchers measure traits, such as body length and gill raker counts. We calculated three effect size estimates for trait differences among ecotypes in each lake: (1) adjacent standardized mean difference (SMD; Borenstein et al., [2009](#page-7-3)), (2) outer SMD and (3) log-transformed ratio of coefficients of variation (lnCVR; Senior et al., [2020](#page-9-15)). Adjacent SMD was estimated as the difference between adjacent ecotypes from the same lake, in which each ecotype was compared with the ecotype closest in value (Figure [1](#page-2-0)). To capture the full range of trait variation within each lake, outer SMD was estimated as the difference between ecotypes in the same lake with the smallest and largest values for a trait (Figure [1](#page-2-0)). InCVR was calculated as the ratio between the coefficients of variation of the two sympatric ecotypes with the smallest and largest trait values, to determine the differences in trait variance associated with the most phenotypically distinct ecotypes in a system. lnCVR was used to estimate the difference between ecotypes in within-ecotype variation (O'Dea et al., [2019](#page-8-9)), while adjacent SMD and outer SMD are measures of the differences between ecotypes in trait mean values. All analyses were performed in R version 4.0.5 (R Core Team, [2021](#page-8-10)).

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To calculate adjacent SMD's we ordered ecotypes from a lake by mean trait value, then calculated the SMD for each ecotype compared to the ones with adjacent lower and higher values. In a group with three ecotypes, for example, there would be two initial comparisons (small to medium and medium to large), from which we calculated a meta-analytic mean to estimate the adjacent SMD. For each initial comparison, an SMD was calculated using *escalc*() from the metafor R package (Viechtbauer, [2010](#page-9-16)). To allow for downstream comparisons among adjacent SMD's, we additionally used the standard error and sample size for each mean trait value to calculate the variance associated with the SMD using the *escalc*() for adjacent SMD's from two-ecotype lakes (Viechtbauer, [2010](#page-9-16)). For lakes with three or more ecotypes, the measures for some ecotypes were used in multiple pairwise comparisons. Therefore, we first calculated a variance/covariance matrix of pairwise comparisons in each lake (Gleser & Olkin, [2009](#page-8-11), equations 19.18 and 19.19) as well as a metaanalytic mean estimate of the adjacent SMD and associated variance (Gleser & Olkin, [2009](#page-8-11), equation 19.21).

Outer SMD's were calculated as the SMD between ecotypes with the smallest and largest trait values within a lake that were reported in a particular paper. Both SMD and the associated variance were calculated using *escalc*() (Viechtbauer, [2010](#page-9-16)). lnCVR was estimated by taking the log-transformed ratio of the coefficients of variation for the largest and smallest ecotypes in a lake. We calculated both an effect size estimate (Senior et al., [2020](#page-9-15), equation 6) and associated variance (Senior et al., [2020](#page-9-15), equation 16) for lnCVR.

In cases where multiple studies have been performed on ecotypes from a lake, there were multiple estimates of each adjacent SMD, outer SMD and lnCVR for that lake. In these cases, we performed a fixed-effects meta-analysis using *rma*() to generate one estimate per effect size for each lake (Viechtbauer, [2010](#page-9-16)).

Several of the physical and ecological features of lakes that we extracted exhibited high correlations (Figure [S2](#page-12-0)). To account for this when fitting traits together in a model, we first ran a principal components analysis (PCA) with eight lake features: (1) lake latitude, (2) log-transformed lake surface area, (3) log-transformed maximum

2.3 | **Ecological lake features**

FIGURE 1 Estimation of effect sizes. Adjacent SMD was calculated by combining the SMD estimates between ecotypes and the one immediately larger or smaller than it within an assemblage. Outer SMD and lnCVR were both estimated between the ecotypes with the largest and smallest trait values within each assemblage, with outer SMD comparing ecotype trait means and lnCVR comparing ecotype trait variances. SMD, standardized mean difference.

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lake depth, (4) log-transformed lake depth ratio (Z_r), (5) mean annual precipitation, (6) seasonality of annual precipitation, (8) mean annual temperature and (8) seasonality of annual temperature. We used a correlation matrix for the PCA because variables were on different scales. We retained the first four principal components, which cumulatively explained 92% of the variance. PC1 was associated with climate variables, with higher PC1 values indicating higher mean annual temperature and precipitation and lower seasonality in temperature and precipitation (Figure [S3](#page-12-0)). Lake dimensions contributed most to PC2, with higher PC2 values indicating larger and deeper lakes. PC3 values decreased with depth ratio, whereas PC4 values increased with latitude.

2.4 | **Intraspecific diversity and abiotic factors**

To test whether the number of ecotypes depended on abiotic features, we used sequential ordinal regression, with three levels in our response variable (two ecotypes, three ecotypes, four or more ecotypes) and the four PC axes summarizing abiotic variation across lakes as the predictors. We fit the full model and all reduced models using *brm*() from the brms R package. To compare models, we estimated model weights using leave one out cross validation (LOOCV) with *loo*(), then applied model averaging via stacking with *loo_model_ weights*(). We summed the weights of models containing each predictor variable and considered predictors with weights greater than 0.7 to have high relative importance (RI), following Wagner et al. ([2014](#page-9-3)). We then re-fit the ordinal regression models including only high RI predictors to estimate effect sizes.

2.5 | **Trait differentiation among populations**

For each trait, we estimated differences in trait values among sympatric populations using separate random effects meta-analyses using the function *rma*() with restricted maximum likelihood (REML) for each of the three effect size estimates (adjacent SMD, outer SMD, and lnCVR), as this approach has improved accuracy over maximum likelihood for fixed effects estimation (Luke, [2017\)](#page-8-12). Adjacent SMD and outer SMD were both estimates of divergence in trait means across an assemblage, while lnCVR was an estimate of divergence in trait variance. To estimate divergence for 'head lengths', we included measures for upper jaw length, lower jaw length, gape width, orbit length, head length and snout length in a multivariate random effects meta-analysis using *rma.mv*() with the trait measured and lake included as random effects. We used the same approach to estimate divergence for 'fin lengths', using lengths of pectoral, dorsal, caudal, anal and pelvic fins.

We tested the relationship between trait divergence and abiotic factors for a subset of traits that are frequently measured in each genus (i.e. body length in both genera, gill raker count in *Coregonus* and age in *Salvelinus*). For each trait, we fit a meta-analytic linear model with the four PC axes summarizing abiotic variation among

lakes as fixed effects, using maximum likelihood (ML) because likelihoods from models fit with REML would not be comparable under our AICc-based model selection approach (Luke, [2017](#page-8-12)). Using *glmulti*(), we estimated the AICc scores and weights for the full model and all reduced models (Calcagno, [2010](#page-7-4)). Following Wagner et al. ([2014](#page-9-3)), we then estimated the relative importance of each effect as the sum of AICc weights of models including that effect. We then re-fitted the meta-analytic linear models with REML and included only high RI predictors to estimate effect sizes.

3 | **RESULTS**

3.1 | **Number of ecotypes and abiotic factors**

In both *Coregonus* and *Salvelinus*, the number of sympatric ecotypes increased with habitat size but were differently affected by other lake features. A greater number of ecotypes in *Coregonus* assemblages increased in probability with higher values of PC2 (lake dimensions; slope $= 0.92$, 95% CI $= [0.46, 1.47]$) and PC4 (latitude; slope = 1.97, 95% CI = [1.00, 3.26]), whereas PC1 (climate) and PC3 (depth ratio) had low relative importance (RI) as predictors (Figure [2](#page-4-0); Table [S1](#page-12-0)). In *Salvelinus*, high ecotype diversity decreased in probability with PC1, meaning that there was a greater probability of few ecotypes in lakes with higher mean temperature and precipitation and lower seasonality (slope = -0.12 , 95% CI = [−0.26, 0.01]; Figure [2](#page-4-0)). The probability of a greater number of sympatric *Salvelinus* ecotypes increased with PC2 (lake dimensions; slope = 0.37, 95% CI = [0.16, 0.58]), but PC3 and PC4 had low RI as predictors (Figure [2;](#page-4-0) Table [S1](#page-12-0)).

3.2 | **Trait divergence among sympatric ecotypes**

Most traits exhibited a similar extent of divergence among sympatric ecotypes in both *Coregonus* and *Salvelinus*. Mean trait divergence was similar between *Coregonus* and *Salvelinus* for most traits when quantified as adjacent SMD's (body length, capture depth, $\delta^{13}C$, δ^{15} N; Figure [3](#page-4-1)) and outer SMD's (age, body length, capture depth, δ15N; Figure [S4a](#page-12-0)). However, *Coregonus* ecotypes showed greater divergence in gill raker counts than *Salvelinus*, with higher values for both adjacent and outer SMD (adjacent SMD: $2.14 \times$ higher, outer SMD: 2.18x higher; Figure [3](#page-4-1)). Additionally, when estimated as outer SMD, δ^{13} C exhibited more divergence in *Coregonus* than *Salvelinus* (1.93x higher; Figure [S4a](#page-12-0)). On the other hand, both adjacent and outer SMD's values for overall body shape were higher in *Salvelinus* than *Coregonus* (adjacent SMD: 3.95× higher, outer SMD: 3.96× higher; Figure [2](#page-4-0), Figure [S4a](#page-12-0)), as was adjacent SMD for age (1.66× higher; Figure [3](#page-4-1)). Among-ecotype differences in trait variance (lnCVR) were similar between *Coregonus* and *Salvelinus* in all traits measured (Figure [S4b](#page-12-0)).

In *Coregonus*, adjacent SMD for body length and gill raker count declined with an increasing number of ecotypes per lake (Figure [4](#page-4-2),

FIGURE 2 The number of sympatric populations by abiotic variables. (a,b) Marginal effects of abiotic predictor variables on the number of sympatric ecotypes found in an assemblage. Each line shows how the posterior probability of observing a particular level of ecotype diversity varies with an abiotic predictor. Shaded regions around lines represent 95% confidence intervals. (c) Relative importance of abiotic factors for predicting the number of populations present in a lake. Relative importance was calculated as the sum of weights for all models that included a factor, based on leave one out cross-validation. PC1 primarily corresponds to variation in temperature and precipitation, PC2 to lake surface area and depth, PC3 to the ratio of lake depth to surface area and PC4 to latitude.

FIGURE 3 Mean trait differences between phenotypically adjacent ecotypes. Each point represents the meta-analytic mean difference among sympatric ecotypes within a lake. Error bars indicate 95% confidence intervals. Results for outer SMD and lnCVR are shown in Figure [S1](#page-12-0). SMD, standardized mean difference.

FIGURE 4 Ecological factors and body length divergence. (a) Relationship between the number of ecotypes present in a lake and the adjacent SMD's (standardized mean differences) in body length between sympatric ecotypes. (b) Relative importance of abiotic factors in predicting adjacent SMD's in body length between sympatric ecotypes. PC1 primarily corresponds to variation in temperature and precipitation, PC2 to lake surface area and depth, PC3 to the ratio of lake depth to surface area and PC4 to latitude. Results for other traits, outer SMD and lnCVR are in Appendix [S1](#page-12-0).

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Figure [S5](#page-12-0), Table [S2](#page-12-0)). Adjacent SMD for body shape PC1, $\delta^{15}N$, and δ13C exhibited no significant relationship to number of *Coregonus* ecotypes (Figure [S5](#page-12-0), Table [S2](#page-12-0)). *Coregonus* outer SMD for gill raker count was positively related to number of ecotypes, but outer SMD was not related to number of ecotypes for body length, body shape PC1, δ^{15} N or δ^{13} C (Figure [S6](#page-12-0), Table [S2](#page-12-0)). Number of ecotypes did not predict trait variance (lnCVR) for any of the traits for which it was estimated in *Coregonus* (gill raker count, body length $\delta^{15}N$ and $\delta^{13}C$; Figure [S7,](#page-12-0) Table [S2](#page-12-0)).

In *Salvelinus*, adjacent SMD for body length declined with number of ecotypes (Figure [4](#page-4-2), Table [S2](#page-12-0)), whereas adjacent SMD for age, asymptotic length, depth of capture, δ^{13} C and δ^{15} N showed no relationship with the number of ecotypes (Figure [S5](#page-12-0), Table [S2](#page-12-0)). There was no relationship in *Salvelinus* between number of ecotypes and either any of the traits quantified for outer SMD (age, asymptotic length, depth of capture, δ^{13} C, δ^{15} N, body shape PC1; Figure [S6](#page-12-0); Table [S2](#page-12-0)) or lnCVR (age, asymptotic length, depth of capture, $\delta^{13}C$, δ^{15} N; Figure [S7](#page-12-0) and Table [S2](#page-12-0)).

3.3 | **Trait divergence and abiotic factors**

For *Coregonus*, PC4 (latitude) had high relative importance (RI) for body length divergence among ecotypes, with adjacent SMD, outer SMD and lnCVR all decreasing at higher latitudes (adjacent SMD: slope= -2.18 , 95% CI = [-2.98 , -1.38]; outer SMD: slope= -1.66 , 95% CI = [-2.63, -0.69]; lnCVR: slope=-0.26, 95% CI = [-0.47, −0.05]; Figure [4](#page-4-2), Figures [S9](#page-12-0) and [S10](#page-12-0)). PC1, PC2 and PC3 all had low RI for adjacent SMD, outer SMD and lnCVR of body length. Gill raker outer SMD increased with PC4 (latitude: slope = 1.34, 95% CI $=$ [0.85, 1.82]), whereas PC1, PC2 and PC3 did not exhibit high RI (Figure [S9](#page-12-0)). No PC axes had high RI for gill raker adjacent SMD or gill raker lnCVR (Figure [S8](#page-12-0), Figure [5\)](#page-5-0).

For *Salvelinus*, body length adjacent SMD declined with PC1 (climate; slope = −0.30, 95% CI = [−0.51, −0.09]), PC2 (lake dimensions; slope = −0.47, 95% CI = [−0.80, −0.14]) and PC4 (latitude; slope = −0.56, 95% CI = [−1.01, −0.11]), whereas PC3 had low RI (Figure [4](#page-4-2), Table [S3](#page-12-0)). Body length outer SMD and age outer SMD both decreased with PC1 (climate) in *Salvelinus* (body length: slope = −0.65, 95% CI = [−0.97, −0.34]; age: slope = −0.20, 95% CI = [−0.33, −0.07]), but all other PC axes had low RI for these traits (Figure [S9](#page-12-0) and Table [S3](#page-12-0)). For age lnCVR, PC3 (depth ratio) had high RI, with age lnCVR decreasing in lakes with lower depth ratio values (slope = −0.15, 95% CI = [−0.28, −0.02]; Figure [5\)](#page-5-0). For age adjacent SMD and body length lnCVR, no abiotic predictor variables had high RI (Figure [S8](#page-12-0), Figure [S10](#page-12-0) and Table [S3](#page-12-0)).

4 | **DISCUSSION**

We provide evidence supporting a role for lake ecosystem characteristics in determining intraspecific diversity in *Coregonus* and *Salvelinus*. Although abiotic features of lakes determined the number of ecotypes in both genera, it was lake dimensions (depth and surface area) and latitude that mattered for *Coregonus* while lake dimensions and climate variables (temperature and precipitation) were important for *Salvelinus*. A relationship of ecotype diversity to lake depth and surface area has been found previously among ecotype assemblages from a restricted geographic region and within the same species (ex. *Salvelinus alpinus*; Koene et al., [2020](#page-8-1); Recknagel et al., [2017;](#page-9-7) Woods et al., [2012](#page-9-17)); our results show that this relationship holds on a circumboreal scale and across both genera. Both biogeographical models and empirical patterns indicate that in situ diversification should increase with habitat size in discrete systems, such as islands and lakes (Losos & Schluter, [2000](#page-8-2); Schluter & Pennell, [2017;](#page-9-18) Valente et al., [2020;](#page-9-19)

FIGURE 5 Ecological factors and divergence among sympatric ecotypes in trait variability. (a) Each point represents the lnCVR in gill raker number among sympatric *Coregonus* ecotypes for one lake. (b) Each point represents the lnCVR in age among sympatric *Salvelinus* ecotypes for one lake. (c) Relative importance of abiotic factors in predicting lnCVR's in gill rakers (*Coregonus*) and age (*Salvelinus*) between sympatric ecotypes. PC1 primarily corresponds to variation in temperature and precipitation, PC2 to lake surface area and depth, PC3 to the ratio of lake depth to surface area and PC4 to latitude.

Wagner et al., [2014](#page-9-3)). In habitats with distinct boundaries, geographical size (surface area and depth) should be a proxy for ecological opportunity but also facilitate the evolution of reproductive isolation if larger habitats allow ecotypes to be spatially distinct during reproduction (Vonlanthen et al., [2009](#page-9-20)). The factors underlying a relationship between habitat size and diversification—opportunity for geographical isolation, availability of distinct resources, and the potential for a larger overall population size—should also enable coexistence of previously diverged ecotypes (Chesson, [2000](#page-8-13); Losos & Ricklefs, [2009](#page-8-14)). Nonetheless, in both *Salvelinus* and *Coregonus*, multi-ecotype assemblages have evolved even in lakes that are very small, suggesting that although ecosystem size may be indicative of ecological opportunity in a lake ecosystem, these two lake traits are frequently decoupled.

The importance of lake surface area and depth for diversification and maintenance of multiple sympatric ecotypes in *Coregonus* is likely related to the presence of profundal ecotypes in the large ma-jority of assemblages with three or more ecotypes (Blain et al., [2023](#page-7-2); Præbel et al., [2013](#page-8-15)). Ecotype pairs in *Coregonus* may not experience the same limitations on habitat size as populations sustaining higher levels of intraspecific diversity, as they typically consist of two shallow water ecotypes—one pelagic and one littoral (Bernatchez et al., [1996](#page-7-5); Öhlund et al., [2020](#page-8-16); Schluter & McPhail, [1993](#page-9-10)). Thus, small and shallow lakes may lack a large enough profundal habitat for a third ecotype to evolve or establish following colonization. Habitat size may directly influence reproductive isolation, as sympatric *Coregonus* ecotypes often spawn at different depths, which has been linked to genetic divergence within multi-ecotype assemblages (Ingram et al., [2012](#page-8-17)).

Latitude emerged as an important predictor of divergence in *Coregonus*, with assemblages at higher latitudes exhibiting greater ecotype diversity, more divergence in gill rakers, and reduced divergence in body length. Patterns in ecotype diversity and gill raker divergence are likely linked, as habitat use in *Coregonus* ecotypes tends to be accompanied by diet specialization on pelagic zooplankton or benthic macroinvertebrates (Bernatchez et al., [1996](#page-7-5); Lundsgaard-Hansen et al., [2013](#page-8-18); Schluter & McPhail, [1993](#page-9-10)). Pelagic versus benthic resource use is reflected in δ^{13} C isotopes and closely tied to gill raker morphology, both of which exhibited high divergence between ecotypes in this study (Figures [5](#page-5-0) and [S8](#page-12-0); Østbye et al., [2006](#page-8-19); Harrod et al., [2010](#page-8-20); Siwertsson et al., [2013](#page-9-21)). Productivity and its variation across an annual cycle may drive a relationship between gill raker divergence and latitude, as pelagic zooplankton are more diverse and abundant in high productivity ecosystems and during the summer while benthic macroinvertebrates are comparatively more plentiful in low productivity ecosystems and available year-round (Hayden et al., [2019](#page-8-21); Kahilainen et al., [2004](#page-8-22), [2005](#page-8-23); Sierszen et al., [2003](#page-9-22)). Perhaps higher latitude environments allow both benthic and pelagic resources to be more favourable at different times of the year (Svanbäck & Bolnick, [2007\)](#page-9-23), facilitating divergence in morphology and indirectly influencing differentiation in spawning time. Whitefish species pairs exhibit greater phenotypic divergence in lakes with

greater seasonality in the community structure of benthic prey (Landry & Bernatchez, [2010](#page-8-24)). Alternatively, resource variation across the annual cycle could favour generalist tactics, and therefore, a less divergent set of sympatric phenotypes capable of exploiting resources from multiple habitats (Chavarie et al., [2016;](#page-8-25) Hayden et al., [2014](#page-8-26)). Higher temperatures are associated with shifts from food webs fueled by benthic to pelagic resources (Hayden et al., [2019](#page-8-21)), so lakes at different latitudes could vary in their source of productivity and its effect on divergence.

In contrast to trophic ecology, life history trade-offs may instead be higher at lower latitudes. Relationships among latitude, growth rates and body length are variable in Salmonids, as fish at lower latitudes might grow faster due to a longer and warmer growing season, while fish may have a counter-gradient relationship or faster annual growth rate at high latitudes (Chavarie et al., [2010](#page-8-27); Power et al., [2005;](#page-8-28) Rypel & David, [2017\)](#page-9-24). Nonetheless, it is possible that growing season constraints at higher latitudes result in fewer or more similar life history tactics within a particular lake environment, resulting in less divergence among ecotypes.

For *Salvelinus*, the number of sympatric ecotypes were best predicted by lake surface area and depth, which suggests that lake size may enhance ecological opportunity in this genus (Eloranta et al., [2015](#page-8-29)). *Salvelinus* assemblages commonly contain a combination of piscivorous, planktivorous and benthivorous ecotypes—all diets associated with distinct usage of habitat zones within a lake (Blain et al., [2023](#page-7-2)). Habitat coupling is common in *Salvelinus* and occurs when individual fish can exploit multiple habitats within a lake. Because habitat coupling can reduce specialization, this connectivity in habitat use is expected to reduce ecotype divergence (Eloranta et al., [2013](#page-8-30); Marklund et al., [2019](#page-8-31)). Habitat coupling is most common in small lakes with a circular shape, making specialization and divergence more likely in larger lakes with reticulate coastlines (Dolson et al., [2009](#page-8-32)). Further, the number of ecotypes was higher in lakes with greater climate seasonality and lower mean temperature and precipitation, suggesting that harsher, more variable climates promote divergence in *Salvelinus*. These lakes may exhibit greater variation in resource availability, contain fewer potential competitors, and more closely match the optimal thermal environment for *Salvelinus* species—all of which have the potential to facilitate ecotype divergence.

Patterns of life history trait and body shape divergence in *Salvelinus* pointed to the importance of ontogeny in the evolution and development of *Salvelinus* ecotype diversity. *Salvelinus* ecotypes often employ alternative diet-associated life history strategies, in which fish that grow to a large size tend to be piscivorous and have a longer lifespan, while invertivores typically grow to a smaller size with a shorter lifespan (Bjorklund et al., [2003](#page-7-6); Parsons et al., [2011](#page-8-33)). The evolution of these alternative strategies was likely driven by differing predation rates, as gape-limited predators can consume the smaller, younger invertivores but not the larger, older piscivores (Reznick et al., [1990](#page-9-25); Smalås et al., [2013](#page-9-26)). Our results are consistent with the idea that plasticity or genetic variation in developmental trajectories is a key prerequisite for sympatric divergence in

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Salvelinus (Parsons et al., [2011](#page-8-33); Skúlason et al., [2019](#page-9-27)). Body lengths differ markedly among *Salvelinus* ecotypes with different life history trajectories (Alekseyev et al., [2009](#page-7-7)), and we found that divergence in this trait increased with temperature and precipitation seasonality. Annual variation in productivity, and therefore resource availability, may allow coexistence of ecotypes with alternative life history strategies. Alternatively, there may be a correlation between seasonality and another causal variable, such as reduced species richness (Jablonski et al., [2017](#page-8-34)). If seasonal environments are more species depauperate, they likely contain greater ecological opportunity (Losos, [2010](#page-8-35)), as there are likely fewer competitors of other species for both piscivorous and invertivorous *Salvelinus* ecotypes.

Phenotypic divergence was similar overall between *Salvelinus* and *Coregonus*, suggesting a degree of similarity, or similar amonglake variability, in the phenotypic effects resulting from processes driving diversification and coexistence. The consistency between genera is somewhat surprising, given prior observations that *Salvelinus* tends to be more ecologically generalist and has uniquely high phenotypic variation facilitated by exceptional levels of plasticity (Chavarie, Adams, et al., [2021](#page-7-1); Klemetsen, [2013](#page-8-4)), and that *Coregonus* ecotypes might be more constrained in their direction of divergence (Blain et al., [2023](#page-7-2)). *Salvelinus* and *Coregonus* tend to diversify in different lakes, with only two lakes in this data set containing ecotype assemblages from both, and this geographical separation might also have been expected to lead to differences in phenotypic variability between the genera. Ecotypes can diverge in both trait means and trait variance (Chavarie, Howland, et al., [2021](#page-8-36); Violle et al., [2012](#page-9-8)), but we found similar levels of phenotypic variability across the two genera despite more common coexistence of generalist with specialist phenotypes in *Salvelinus* (Chavarie et al., [2016](#page-8-25)). Nonetheless, *Coregonus* and *Salvelinus* exhibited a few differences in trait divergence that appear to be linked to their different patterns of resource exploitation and drivers of intraspecific diversity. *Coregonus*, which exhibited greater gill raker divergence, has a feeding ecology that is typically restricted to lower trophic level prey (i.e. zooplankton and/or benthic invertebrates), whereas *Salvelinus*, which exhibited elevated age and body shape divergence, has wider niche use as they can be piscivorous and omnivorous (Klemetsen, [2013](#page-8-4)).

In sum, we found that habitat size predicts the number of ecotypes in *Coregonus* and *Salvelinus* assemblages, while latitude and climate variables predict the level of divergence among some traits. These relationships may be critical to ecotype persistence, or lack thereof, as environments change. Ecotype diversity affects ecosystem function (Des Roches et al., [2018](#page-8-37)), and this may be especially true in high latitude and high altitude lakes that tend to be relatively species depauperate. Temperature and precipitation seasonality, which are associated with trait divergence in these genera, are expected to shift with climate. Changes in productivity may also change resources distributions across habitats in lakes of different sizes. Looking forward, it will be important to evaluate the mechanisms by which these abiotic variables facilitate the astonishing levels of intraspecific diversity expressed in *Salvelinus* and *Coregonus*.

AUTHOR CONTRIBUTIONS

Conceptualization: SAB, CEA, PAA, RK and LC. Developing methods: SAB, CEA and LC. Data analysis, preparation of figures and tables: SAB. Conducting the research: SAB and LC. Data interpretation, writing: SAB, CEA, PAA, RK and LC.

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CONFLICT OF INTEREST STATEMENT

The authors have no conflicts of interest to declare.

DATA AVAILABILITY STATEMENT

Phenotypic traits and lake characteristic data are available on Dryad at <https://doi.org/10.5061/dryad.pg4f4qs01>. Scripts used for data analysis are available on GitHub: [https://github.com/stephblain/](https://github.com/stephblain/salmonids_ecological_factors) [salmonids_ecological_factors](https://github.com/stephblain/salmonids_ecological_factors).

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SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

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