State-dependent migratory timing of post-spawned Atlantic

2	salmon (Salmo salar)
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Abstract

Atlantic salmon (*Salmo salar*) often survive spawning and migrate back to the sea to feed, either shortly after spawning in autumn or in the following spring. We conducted a four-year observational field study using telemetry to evaluate the determinants of migration timing in Atlantic salmon post-spawners (kelts). We found that individuals with low energy reserves migrated early to the risky but productive marine habitat, whereas individuals with greater energy reserves stayed in the safe but less productive river habitat until staying became energetically more costly than migrating. For males, the likelihood of overwintering in the river instead of migrating in autumn increased 27-fold with each increase in body condition index, whereas almost all females overwintered in the river. Among spring migrants, body condition was the strongest determinant of migration timing, and females left the river about five days later than males. Our study suggests that migration timing in Atlantic salmon kelts is the outcome of adaptive state-dependent habitat use, related to individual and sexual differences in energy allocation during spawning.

Keywords: telemetry, life-history, reproduction, iteroparity, behaviour

Introduction

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Anadromous salmonids reproduce in fresh water, but achieve the bulk of their growth at sea (Gross 1987). The marine part of the lifecycle is generally characterised by low survival and rapid growth, whereas the freshwater residency offers poorer growth (Jonsson and Jonsson 1993). This difference exists because both the predation pressure and production are higher in the marine than in the freshwater habitat (Gross 1987; Ward and Hvidsten 2010). As salmonids are ectotherms, they also grow faster with increasing temperature up to a growth optimum (Jobling 1994, Jobling et al. 2010). Therefore, salmonids grow especially well in the summer when both the temperature and the food availability are high at sea (Ingvaldsen and Loeng 2009; Rikardsen and Dempson 2010). In contrast to most Pacific salmon species, Atlantic salmon (Salmo salar L.1758) often survive spawning in the river (Jonsson et al. 1990; 1991a), and perform repeated marine feeding migrations to build new reserves for re-occurring spawning events (Fleming 1996). The low temperatures experienced by the surviving post-spawning salmonids after spawning (late autumn/winter) restrict their activity and growth potential (see Knutsen et al. 2004) both in the river and at sea. Therefore, the winter season is not ideal for re-building depleted energy resources. This may explain the large individual variation in the timing of feeding migration observed in post-spawning Atlantic salmon (kelts); they either migrate shortly after spawning in the autumn, or during the following spring/summer (Jonsson et al. 1990; 1991a; Niemelä et al. 2000; Webb et al. 2007). Here, we hypothesize that the variation in migration timing is the outcome of adaptive statedependent habitat use, related to differences in energy allocation during spawning. In order to address this hypothesis, we analysed data from a four-year observational field study based on remote sensing of individual behaviour. We tagged a total of 352 Atlantic salmon, and

registered their post-spawning freshwater survival and migratory timing by means of acoustic and radio telemetry. For Atlantic salmon, the likelihood of repeat spawning is negatively related to energy depletion during the previous spawning event (Fleming 1996). The assetprotection principle (Clark 1994) predicts that individuals with low energy reserves and a low probability of future reproduction should accept risky, productive habitats, whereas individuals with large reserves and a high probability of future reproduction should choose safe but less productive habitats. Therefore, applying the asset-protection principle to the migration of Atlantic salmon kelts, we predicted that kelts with low energy reserves would migrate to the sea shortly after spawning, whereas kelts with larger reserves would stay in the river until reaching an energetic threshold, determined by a trade-off between metabolic costs and gains in fresh waters versus marine waters. We thus expected the migration timing to be governed by individual factors related to state-dependency, specifically the individual's body condition, size and sex. As large Atlantic salmon individuals tend to allocate relatively more energy to spawning than small individuals (Jonsson et al. 1997), and as males have higher energy expenditure than females (Jonsson et al. 1991b; Fleming et al. 1997; Niemelä et al. 2000), we expected that (i) individuals in poor body condition would leave the river early, (ii) males would leave the river earlier than females, and (iii) large individuals would leave the river earlier than small individuals.

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Material and Methods

Study area

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The study was conducted in Alta River and Fjord in northern Norway (70°N, 23°E, Fig. 1) during 2007-2010. Alta River has a catchment area of 7400 km² and a mean annual water discharge of 75 m³ s⁻¹. It supports a large Atlantic salmon population with annual in-river catches between 6 and 32 t, with a mean mass of 5.2 kg for individual Atlantic salmon (Ugedal et al. 2008). Most spawners spend one or three years in the sea, where one-seawinter fish (1SW) are mainly males (76%, mean 2.3 kg) and three-sea-winter fish (3SW) are mainly females (77%, mean 10.0 kg) (Ugedal et al. 2008). Repeat spawning is more common in females than in males, and the large majority of both female and male repeat spawners spend one year at sea before returning to spawn again (Halttunen et al. 2011). The river length accessible to Atlantic salmon is 46 km, and a hydropower plant was constructed above this stretch in 1987. Winter freeze-up usually occurs in November, and spring ice break-up normally occurs in early May. There are numerous suitable areas for kelts to overwinter (deep pools, still waters), and a minimum discharge of 20-30 m³ s⁻¹ is maintained by the hydropower plant during the winter. The peak spawning run of Atlantic salmon occurs from mid-June through July, and spawning commences during the first half of October (Ugedal et al. 2008).

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Tagging

The post spawning survival and seaward-migration timing of Atlantic salmon kelts was assessed by the use of radio and acoustic telemetry (Cooke et al. 2004). Both methods enable monitoring movements of individual fish by using electronic tags that transmit individually coded signals. Radio telemetry was used to assess the post-spawning survival and the overall

seaward-migration timing of kelts. The method was chosen because it enabled tracking of fish through ice during the winter. Acoustic telemetry was used for a higher resolution study of out-migration timing versus energy reserves in the spring. The method was chosen because it enables tracking fish in saline water.

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Radio tagging

We tagged a total of 166 returning Atlantic salmon in the Alta Fjord and River between June and September 2008-2009 (Table 1). Multi-sea-winter (MSW) fish (56 males, 2-5SW, mean fork length 95 cm; 92 females, 2-6SW, mean fork length 94 cm) were captured in bag nets in the Alta Fjord during their spawning migration towards the Alta River from June to August 2008-2009. Uninjured specimens were placed in a 400 L tank with oxygenated water in the research vessel (time in tank 5-120 min), and thereafter externally tagged with radio transmitters (Model F2120, Advanced Telemetry Systems (ATS), MN, flat with outline dimensions 21×52×11 mm, mass in air 15 g, estimated battery lifetime 10 or 13 months). Individuals could be identified by using unique combinations of frequencies and pulse-rates. The fish were anaesthetised prior to tagging (2-phenoxy-ethanol, EC No. 204-589-7, SIGMA Chemical Co., MO, 0.5 ml L⁻¹, mean time 2 min 40 sec), and thereafter placed in a waterfilled tube with the head and gills submerged. The radio transmitters were attached through the musculature 3-4 cm below the dorsal fin by use of 0.8 mm stainless-steel wires, and anchored to a 10×40 mm plastic back plate. The mean tagging time was 1 min 50 sec, and the mean handling time was 5 min 20 sec. The tagged fish were left to recover in the oxygenated tank under observation (mean time 2 min 50 sec), whereupon they were released back into the fjord to resume their migration towards the river. One-sea-winter Atlantic salmon often escape bag nets due to the large mesh sizes. Therefore, the radio-tagged group was biased towards MSW males, and so the radio-tagged group was

more similar in size than the acoustically tagged group (see Table 1). To increase the number of 1SW males in the study, eight additional individuals (mean fork length 57 cm) were captured by rod and reel in the upper parts of the river between 6 and 30 September 2009, and tagged using similar procedures as described above. An additional 10 MSW Atlantic salmon (eight females, two males, mean fork length 92 cm,) were captured in the same area and tagged during the same period (Table 1).

Acoustic tagging

We tagged a total of 202 kelts (57 males, 1-3SW, mean fork length 65 cm; 145 females, 2-7SW, mean fork length 94 cm, Table 1) with acoustic tags (Vemco Inc., Canada, model V13-1L, cylindrical with dimensions 13 mm×36 mm, mass in air 11 g, and Thelma AS, Norway, model MP-13, 13 mm×31mm, mass in air 11 g, estimated battery life 13 or 37 months) in the Alta River during April - May 2007-2009. The kelts were captured in the river by angling with barbless treble hooks, anaesthetised and tagged internally according to Halttunen et al. (2010). All kelts were also externally tagged with modified Carlin tags with contact and reward information for the report of recaptures.

Tracking

Radio tracking

To detect radio-tagged fish descending the river, automatic stationary receivers storing information on tag ID and time of registration (R4500S Scientific receiver, ATS, connected to a four- or nine-element Yagi-antenna) were installed 1 and 8 km upstream from the river outlet during both study years (Fig. 1). In addition, stationary receivers were operative 15 and 19 km upstream from mid-October 2009. The scan time on the receiver for each frequency

used on radio transmitters was 5 sec, giving a total scan time ranging from 2 min 30 sec to 4 min 10 sec to cover all frequencies used and to be able to register fish within the detection area of the receiver (approximately 1 km).

The positions and possible predation events of the radio-tagged fish within the river were determined by manual tracking using a portable receiver (R410 Scanning receiver, ATS, connected to a three-element Yagi-antenna). The fish were positioned with an accuracy of approximately \pm 300 m, i.e. fish were localized to a specific pool in the river. Tracking was performed approximately every second week from spawning (October) until the river froze (~mid-November), on a monthly basis during winter, and every third week from ice break-up (~mid-April) until the end of July. Due to unstable ice conditions during the winter 2008-2009, tracking in December and January could not be performed because accessing these remote areas in wintertime requires snowmobile transport on the river ice.

Acoustic tracking

To detect acoustically tagged fish descending the river, eleven acoustic receivers (model VR2, Vemco Inc.) were deployed at 5 m depth, spaced 400 m apart, in an array across the Alta Fjord 4 km from river mouth (Fig. 1) between April and October 2007-2009. In addition, two receivers were deployed near the river mouth in the lower reaches of the river.

Information on the time of registration and the tag ID were stored in the acoustic receivers whenever an individual passed within the detection range. The receivers were able to detect fish at a range varying from 200-600 m, depending on environmental conditions. Hence, all fish passing the receiver arrays were within the detection range.

Sampling

A total of 144 kelts (46 males and 68 females, mean fork length 82 cm, range 50-112 cm) were sacrificed during tagging in the river (2007-2009), and their stomach contents were analysed. The water temperature and discharge in the river were measured hourly at Kista hydrographical station 28 km upriver, and the sea temperature was measured at an average of 10 days interval at Ingøy hydrographical station (N 71°08' E 24°01') at one meter depth.

Data analysis

The minimum survival of Atlantic salmon kelts

Because we chose small tags to minimize potential tag effects on behaviour (Thorstad et al. 2000), some of the batteries were dead before the fish left the river. Due to the uncertain fates of fish with undetected exit, we based the survival rate estimate strictly on confirmed survived fish, i.e. fish with detected exit and recaptured fish, and therefore our survival estimate is a minimum estimate. We considered fish reported recaptured and released by anglers in the river during 1 June – 31 August to be alive with a high survival probability, despite being caught and released (Halttunen et al. 2010). Predation was assessed from tagged carcasses detected in the vicinity of the river.

Analyses of seaward-migration timing of Atlantic salmon kelts

Two separate models were used to test the determinants for the timing of seaward-migration; a logistic regression model to assess the probability of the radio-tagged fish migrating either during autumn (defined here as before end of January) or spring (defined her as after onset of April), and a linear regression model to explain *sea entry date* (in Julian days using 1 May as day 1) of the acoustically tagged fish in the spring. For both analyses, we were interested in

assessing the influence of sex, length (fork length in cm), body condition index and their interactions. The body condition index consisted of the residuals of log [mass] regressed on log [length], and was chosen in order to avoid co-linearity between the explanatory variables length and body condition (R = 0.34). The body condition index was calculated from length and mass measured at tagging, i.e. it represented the body condition in two distinct phases of the spawning migration; before spawning for the radio tagged fish and after spawning and overwintering for acoustically tagged fish. The variable *year* was included to account for among-year variation in the environment, which can be considerable; for example, the timing and intensity of the flood peak varied greatly among the study years (range 400-816 m³s⁻¹, Fig. 2). The variable tag site was included as a covariate in the linear model explaining sea entry date to control for distances between tagging sites and the sea. For the logistic regression model predicting spring migration, only 2 females out of 63 left in autumn (3%) compared with 9 out of 21 males (43%). Therefore, we only analysed the probability of spring migration for males. Because we had very few males in this analysis, we could not assess the influence of interactions. In addition, to prevent type II errors owing to the low power of the test (4 estimates -intercept, year, condition, length- for 21 data points), we performed stepwise removal of least significant variables. We performed a Pearson χ^2 statistic to assess the goodness-of-fit of the model, which showed no evidence of overdispersion (P = 0.99). We present odds ratio (with 95% confidence intervals), which is a measure of effect size in logistic regression (Nakagawa and Cuthill 2007). For continuous variables like *length* and *body condition index*, the odds ratio is the odds of migration occurring in the spring with an increase of one unit, e.g. the ratio of the odds when length equals 100 cm to the odds of it occurring when it equals 101 cm. An odds ratio of 1 indicates that spring migration is equally probable with change of one unit. When the ratio moves towards 0, spring migration is more likely to occur with a decrease of one unit, whereas when

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it moves towards infinity, it is more likely to occur with an increase of one unit. To assess the performance of the logistic model, we report the area under the receiver operating characteristic curve (AUC), which measures the overall predictive accuracy of the model independent of a specific threshold (Fielding and Bell 1997, R package "lrm"). AUC values vary from 0.5 to 1, and represent the percentage of randomly drawn pairs (i.e. one of each group) that the model classifies correctly. Generally, the predictive accuracy of a model is classified as "worthless" with an AUC close to 0.5 and "excellent" with an AUC close to 1.0. All variables in both models were centred to allow interpretation of main effects independently of interactions (Schielzeth 2010). In addition, all continuous variables were standardized in order to provide standardized parameter estimates that allow comparing the magnitude of the effects, and hence, the importance of each variable (Schielzeth 2010). Because the models included a categorical variable, we used two standard deviations to standardize the continuous variables, therefore permitting comparison between categorical and continuous variables (Gelman 2008; Schielzeth 2010). All statistical analyses were conducted using the open source R language for statistical computing and graphics (R Development Core Team 2009).

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Results

The minimum survival of Atlantic salmon kelts

The minimum survival of Atlantic salmon kelts was 63% (98/156 individuals included in the analysis, both study years pooled). Out of these survivors, 13% (13/98) left the river in the autumn, and the remaining 87% (85/98) left in the spring. The majority of the autumn migrants (10 of 13, 77%) were MSW males. The minimum survival rate was higher for MSW females (74%) than for MSW males (43%; χ^2 -test, P < 0.001), but there was no difference in survival between MSW females and 1SW males (62%; Fisher's exact test, P = 0.70).

Overall seaward-migration timing of Atlantic salmon kelts

Despite no difference in the body condition between the males and females (t test assuming equal variances: t = 1.99, P = 0.29, df = 80) upon entry, almost all females started their migration in spring (97%) whereas almost half of the males migrated in the autumn (43%). Even though salmons had a slight tendency to enter later with lower body condition (Pearson correlation coefficient: r = -0.20, P = 0.08, n = 76), the river entry date was not correlate with the exit date (Pearson correlation coefficient: t = -0.05, t = 0.09). Among males, the complete model (with low power because it is getting towards saturation with about 5 data points available for each estimate) showed a tendency for males in better condition to leave in the spring (Table 2). Model reduction, increasing the power of the analysis, confirmed the effect of condition on the probability of spring migration (odds ratio [CI] = 26.7 [2.2;1265.0], Wald t = 2.1, t = 0.036). Males were 27 times more likely to migrate in the spring with each increase of one unit in body condition index (Fig. 3). In general, the probability of males migrating to sea in the spring was 50% if their body condition index was close to 0 or

positive, meaning fish of average or higher than average condition (Fig. 3). Overall, the model had an excellent predictive power, as indicated by the high AUC value (Table 2).

Migration timing among spring migrating Atlantic salmon kelts

In total, 188 (93%) of all the acoustically-tagged spring migrants were detected on the acoustic arrays in the fjord. Among sex, body condition, and fork length, body condition had the strongest influence on migration timing of detected fish: fish in better body condition migrated later, and this effect was stronger for females than for males (Table 3, Fig. 4a). Sex was the second strongest predictor but its effect was three times smaller than that of body condition (Table 3). On average, females migrated five days later than males (Table 3). Fish length did not affect sea entry date in the spring, but its interaction with sex was a strong predictor (Table 3, Fig. 4b); males entered the sea slightly later with increasing fork length, whereas females' sea entry date did not vary with length. Sea entry date varied among years, but variation among tag sites was negligible (Table 3). The median migration date during the study period was 9 June (Fig. 2).

River and sea temperatures, feeding and predation in the river

The river winter temperatures (1 November – 30 April) in 2007-2010 ranged between 0.0-3.0°C (average of 0.4°C), while the sea temperatures ranged between 4.1-7.8°C (average of 5.6°C). The river temperatures surpassed the sea temperatures each year in the middle of June (Fig 2). None of the sampled kelts had stomach contents, and we observed no predation on the tagged fish.

Discussion

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Consistent with our expectations, fish in poorer condition left the river earlier than fish in better condition, both when comparing between autumn and spring migrating males, and among spring migrating females and males. This indicates that there may be an inherent advantage to prolonging the freshwater residency instead of initiating the seaward-migration as soon as possible. We suggest that the advantage may be related to the differences between the river and marine habitat during winter, in relation to metabolic costs vs. gains, and predation risk. The river is a low-cost, low-gain, and low-risk habitat for overwintering kelts. Fish that overwinter in the river at temperatures close to 0 °C should have very low metabolic costs (Jobling 1994). We found no evidence for metabolic gains through feeding (because kelts had empty stomachs), and the kelts risk little or no predation in the river, as they are virtually untouchable under the protective ice cover. The known predators of Atlantic salmon in the River Alta are white-tailed sea eagles *Haliaeetus albicilla* and European otters *Lutra* lutra, and as a rule, they feed on dead fish after spawning before the river freezes (E. Halttunen, personal observations). The marine environment, on the other hand, is a productive but more risky habitat for kelts. Despite lower marine production during winter, there are prey (fish, amphipods, euphausiids) available in the sea throughout the year. In addition, even though marine predation on adult Atlantic salmon is not considered an important factor for mortality (reviewed by Mather 1998), large predators like seals and whales are known to predate upon adult salmon (Ward and Hvidsten 2011, own unpublished data). Therefore, we suggest that individuals that can energetically afford to do so, should overwinter in the river and migrate to the sea when increased temperatures improve feeding opportunities. This hypothesis is further supported by the fact that the median outmigration date among the spring migrants coincided roughly with the equalising of temperatures in the river and along the outer coast, therefore rendering the metabolic costs of staying in the river

higher than that of migrating to the sea. As Atlantic salmon migrate between the river and marine habitats, where they cover large distances (Klemetsen et al. 2003), optimizing each leg of the migration to the lowest possible risk and the highest possible gain is likely important for the overall survival and lifetime fitness.

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Consistent with our expectations, males left the river earlier than females, both when comparing autumn to spring migrants and spring migrants among each other. Atlantic salmon males generally invest more of their somatic energy (36%) in spawning than females (25%, Jonsson et al. 1991b) as a result of active and aggressive behaviour and investments in secondary sexual traits (Fleming 1998). This somatic energy loss, when added to injuries acquired during intense male-male competition during spawning, may lead to poor physiological status and high post-spawning mortality among males (Jonsson et al. 1990; Baglinière et al. 1991; Fleming 1998). However, Atlantic salmon males display two alternative reproductive strategies. They return from their sea feeding migration either after one winter as small individuals, or after multiple winters as large individuals. Small males invest less of their total energy in spawning (40-50%) compared to large males (up to 70%), and larger males fight more violently (Fleming 1996; Jonsson et al. 1997; Jonsson and Jonsson 2003). Accordingly, the mortality among large males in this study was higher than among females and small males, and almost all of the autumn migrants were large males. Small males may minimize their involvement in the spawning event to conserve their energy, therefore increasing their chances of surviving to spawn again (Seamons and Quinn 2010). This seems to be the case in River Alta, where small males have a much higher repeat spawning probability than large males (Halttunen 2011).

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Almost all females (98%) overwintered in the river and started their ocean migration in the spring, despite large variations in body condition when entering the river. This shows a clear difference in the overall migration strategy between the sexes, and most likely, a difference in energy allocation during spawning. Atlantic salmon females generally allocate most of their energy reserves in egg production (~30%; Jonsson and Jonsson 2003) and little in aggression and competition, therefore suffering lower post-spawning mortality (Jonsson et al 1990; Baglinière et al. 1990, 1991; Fleming 1998). Consequently, females have a higher repeat-spawning probability than males (Heggberget 1989; Jonsson et al. 1990; Niemelä et al. 2000). Females in River Alta can spawn multiple times (Halttunen 2011). Therefore, according to the asset protection principle (Clark 1994), females should choose safe but less productive habitats like the river after spawning, and hence prioritise the next spawning occasion because they have a high future reproductive probability. As the large majority of the repeat spawning females and males spend a year in the sea before returning, i.e. spawn alternate years (Halttunen 2011), the difference in the migration behaviour between sexes cannot result from a difference in repeat spawning.

Based on the knowledge that larger fish tend to use more energy than small fish during spawning (Jonsson et al. 1997), we expected large individuals to leave the river earlier than small individuals, and this was certainly true for the overall migration pattern of males, where almost all of the autumn migrants were large males. However, we observed the contrary among the spring migrating males; larger males left the river at a later date than smaller males (Fig 4b). For spring migrating females, length had no effect on migration timing (Fig 4b). There might be several factors accounting for large size not advancing the outmigration date in the spring. Especially for females, large size is an asset worth protecting, as size correlates positively with fecundity (Fleming 1996). Large fish also have a lower metabolic

rate in relation to body size than small fish (Jobling 1994), and might use relatively less energy than small fish during overwintering. In addition, large fish might be more dependent on good feeding conditions to meet the demands of growth compensation. Growth compensation (Jobling 1994, 2010; Ali et al. 2003) is a phenomenon occurring in Atlantic salmon after a period of starvation, induced by the loss of body mass and reduced lipid reserves. It leads to high feeding rates, and a growth rate above that normally recorded in fish that feed regularly. Growth compensation may have far-reaching consequences, as there are trade-offs between growth and other fitness-related traits, such as physiological costs in the form of reduced swimming performance, immune capacity and developmental stability (Jobling 2010; Ali et al. 2003). Therefore, waiting for good growth conditions before starting the feeding migration might be crucial for adult post-spawners, and especially for large ones.

Variation among years affected the overall (autumn/spring) migration timing of Atlantic salmon kelts, likely due to annual changes in environmental factors. The river temperatures Atlantic salmon experience prior to and during spawning will have an immediate effect on their energy expenditure through metabolic costs (Jobling 1994). As we suggest that outmigration is related to body condition, this could explain the differences between years in the likelihood of migrating in the autumn. Environmental factors had also a significant effect on the out-migration during spring. The migration timing of kelts varied in accordance to the timing of the flood peak (earliest in 2009 and latest in 2008, Table 3, Fig. 2), and hence the spring flood may be an important environmental cue for kelt migration initiation. This concurs with the migratory timing of Atlantic salmon smolts, which synchronize their outmigration with environmental cues in an attempt to match the most profitable growth conditions at sea (Hvidsten et al. 2009; Rikardsen and Dempson 2010). However, despite the clear effect of environmental cues, the out-migration of kelts spanned out over a period of 6-

12 weeks, whereas the smolts migrate over a more concentrated time period (Jensen et al. 2012). This might be because kelts are considerably larger than smolts, and therefore not dependent on the safety of a crowd under migration. Also, kelts have larger mouths, and therefore, a larger feeding repertoire than smolts, whereas smolts are dependent on certain seasonal prey types (e.g. zooplankton; Hvidsten et al. 2009; Rikardsen and Dempson 2010). Finally, even though kelts might also benefit from entering the sea in the most productive season, they might be forced to leave the river earlier than would be optimal for feeding opportunities at sea due to poor body condition.

On a more general note, the morphology and hydrography of rivers can play an important role in the migration timing of post-spawning Atlantic salmon. The decision to overwinter in the river is likely dependent on suitable overwintering areas and adequate water discharge throughout the winter (as documented in *e.g.* brown trout *Salmo trutta* and Arctic charr *Salvelinus alpinus*; Olsen et al. 2006; Jensen and Rikardsen 2012), as well as on a favourable temperature regime and relative differences between river and sea temperatures. Therefore, the individual condition and the future reproductive prospects of the fish work most likely in concert with environmental factors in determining the overall migration timing of spent Atlantic salmon. This study, however, demonstrates the importance of physical condition in migration timing, and provides a theoretical framework for future salmonid migration studies.

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423 References

424 Ali, M., Nicieza, A., and Wootton, R. J. 2003. Compensatory growth in fishes: a response to 425 growth depression. Fish Fish. 4: 147–190. doi: 10.1046/j.1467-2979.2003.00120.x. 426 Baglinière, J.L., Maisse, G., and Nihouarn, A. 1990. Migratory and reproductive behaviour of 427 female adult Atlantic salmon, Salmo salar L., in a spawning stream. J. Fish Biol. **36**(4): 511–520. doi:10.1111/j.1095-8649.1990.tb03553.x. 428 429 Baglinière, J.L., Maisse, G., and Nihouarn, A. 1991. Radio-tracking of male adult Atlantic 430 salmon, Salmo salar L., during the last phase of spawning migration in a spawning 431 stream (Brittany, France). Aquat. Living Resour. 4(3): 161–167. doi: 432 10.1051/alr:1991017. 433 Clark, C.W. 1994. Antipredator behavior and the asset-protection principle. Behav. Ecol. 434 **5**(2): 159-170. doi:10.1093/beheco/5.2.159. 435 Cooke S.J., Hinch S.G., Wikelski M., Andrews R.D., Kuchel L.J., Wolcott T.G., Butler P.J. 436 2004. Biotelemetry: a mechanistic approach to ecology. Trends Ecol. Evol. 19(6): 437 334–343. doi: 10.1016/j.tree.2004.04.003. 438 Fielding, A.H., and Bell, J.F. 1997. A review of methods for the assessment of prediction 439 errors in conservation presence/absence models. Environ. Conserv. 24(1): 38-49. doi: 10.1017/S0376892997000088. 440 441 Fleming I.A, 1996. Reproductive strategies of Atlantic salmon: ecology and evolution. Rev. 442 Fish Biol. Fish. **6**: 379–416. doi: 10.1007/BF00164323. 443 Fleming, I.A. 1998. Pattern and variability in the breeding system of Atlantic salmon (Salmo 444 salar), with comparisons to other salmonids. Can. J. Fish. Aquat. Sci. 55(Suppl. 1): 445 59-76. doi: 10.1139/d98-009.

446	Fleming, I.A., Lamberg, A., and Jonsson, B. 1997. Effects of early experience on the
447	reproductive performance of Atlantic salmon. Behav. Ecol. 8 (5): 470-480. doi:
448	10.1093/beheco/8.5.470.
449	Gelman, A. 2008. Scaling regression inputs by dividing by two standard deviations. Statist.
450	Med. 27 : 2865–2873. doi:10.1002/sim.3107.
451	Gross, M.R. 1987. Evolution of diadromy in fishes. Am. Fish. Soc. Symp. 1: 14-25.
452	Halttunen, E., Rikardsen, A.H., Thorstad, E.B., Naesje, T.F., Jensen, J.L.A., and Aas, Ø.
453	2010. Impact of catch-and-release practices on behavior and mortality of Atlantic
454	salmon (Salmo salar L.) kelts. Fish. Res. 105(3): 141-147. doi:
455	10.1016/j.fishres.2010.03.017.
456	Halttunen, E. 2011. Staying alive – the survival and importance of Atlantic salmon post-
457	spawners. Ph.D. thesis, University of Tromsø, Tromsø, Norway. Munin Open
458	Research Archive: http://munin.uit.no/handle/10037/3536
459	Heggberget, T.G. 1989. The population structure and migration system of Atlantic salmon,
460	Salmo salar, in the River Alta, North Norway. A summary of the studies 1981-1986.
461	In Proceedings of the Salmonid Migration and Distribution Symposium. Edited by
462	E.L. Brannon and B. Jonsson. University of Washington, School of Fisheries, Seattle
463	pp. 124-139.
464	Hvidsten, N.A., Jensen, A.J., Rikardsen, A.H., Finstad, B., Aure, J., Stefansson, S., Fiske, P.,
465	and Johnsen, B.O. 2009. Influence of sea temperature and initial marine feeding on
466	survival of Atlantic salmon Salmo salar post-smolts from the Rivers Orkla and Hals,
467	Norway. J. Fish Biol. 74 (7): 1532-1548. doi: 10.1111/j.1095-8649.2009.02219.x.
468	

- Ingvaldsen, R., and Loeng, H. 2009. Physical oceanography. *In* Ecosystem Baretns Sea.
- 470 Edited by E. Sakshaug, G. Johnsen and K. Kovacs. Tapir Academic Press,
- 471 Trondheim, Norway. pp. 33–64.
- Jensen, A.J., Finstad, B., Fiske, P., Hvidsten, N.A., Rikardsen, A.H., and Saksgård, L. 2012.
- 473 Timing of smolt migration in sympatric populations of Atlantic salmon (Salmo salar),
- brown trout (*Salmo trutta*), and Arctic charr (*Salvelinus alpinus*) Can. J. Fish. Aquat.
- 475 Sci. **69**: 711–723. doi:10.1139/F2012-005.
- Jensen, J.L.A., and Rikardsen, A.H. 2012. Archival tags reveal that Arctic charr Salvelinus
- 477 *alpinus* and brown trout *Salmo trutta* can use estuarine and marine waters during
- winter. J. Fish Bio. **81**(2), 735–749. doi:10.1111/j.1095-8649.2012.03343.x.
- Jobling, M. 1994. Fish bioenergetics. Chapman & Hall, London, UK.
- Jobling, M. 2010. Are compensatory growth and catch-up growth two sides of the same coin?
- 481 Aquacult. Int. **18**: 501–510. doi: 10.1007/s10499-009-9260-8.
- Jonsson B, and Jonsson N. 1993. Partial migration: niche shift versus sexual maturation in
- 483 fishes. Rev. Fish Biol. Fish. **3**: 348–365.
- Jonsson, N., and Jonsson, B. 2003. Energy allocation among developmental stages, age
- groups, and types of Atlantic salmon (*Salmo salar*) spawners. Can. J. Fish. Aquat.
- 486 Sci. **60**(5): 506-516. doi: 10.1139/f03-042.
- Jonsson, N., Jonsson, B., and Hansen, L.P. 1990. Partial segregation in the timing of
- 488 migration of Atlantic salmon of different ages. Anim. Behav. 40(2): 313-321. doi:
- 489 10.1016/S0003-3472(05)80926-1.
- Jonsson, N., Hansen, L.P., and Jonsson, B. 1991a. Variation in age, size and repeat spawning
- of adult Atlantic salmon in relation to river discharge. J. Anim. Ecol. **60**(3): 937-947.
- 492 Available from http://www.jstor.org/stable/5423 [accessed 3 May 2011].

493	Jonsson, N., Jonsson, B., and Hansen, L.P. 1991b. Energetic cost of spawning in male and
494	female Atlantic salmon (Salmo-salar L). J. Fish Biol. 39(5): 739-744. doi:
495	10.1111/j.1095-8649.1991.tb04403.
496	Jonsson, N., Jonsson, B., and Hansen, L.P. 1997. Changes in proximate composition and
497	estimates of energetic costs during upstream migration and spawning in Atlantic
498	salmon Salmo salar. J. Anim. Ecol. 66(3): 425-436. Available from
499	http://www.jstor.org/stable/5987 [accessed 3 May 2011].
500	Mather, M.E. 1998. The role of context-specific predation in understanding patterns exhibited
501	by anadromous salmon. Can. J. Fish. Aquat. Sci. 55: 232-246. doi: 10.1139/d98-002.
502	Nagelkerke, N.J.D. 1991. A note on a general definition of the coefficient of determination.
503	Biometrika 78 (3): 691-692. doi: 10.1093/biomet/78.3.691.
504	Nakagawa, S., and Cuthill, I.C. 2007. Effect size, confidence interval and statistical
505	significance: a practical guide for biologists. Biol. Rev. 82 (4):591-605. doi:
506	10.1111/j.1469-185X.2007.00027.x.
507	Niemelä, E., Mäkinen, T.S., Moen, K., Hassinen, E., Erkinaro, J., Länsman, M., and
508	Julkunen, M. 2000. Age, sex ratio and timing of the catch of kelts and ascending
509	Atlantic salmon in the subarctic River Teno. J. Fish Biol. 56(4): 974-985. doi:
510	10.1006/jfbi.1999.1223.
511	Olsen, E.M., Knutsen, H., Simonsen, J.H., Jonsson, B., and Knutsen, J.A. 2006. Seasonal
512	variation in marine growth of sea trout, Salmo trutta, in coastal Skagerrak. Ecol.
513	Freshw. Fish 15 (4): 446–452. doi: 10.1111/j.1095-8649.2004.00285.x.
514	R Development Core Team. 2009. R: A Language and Environment for Statistical
515	Computing. R Foundation for Statistical Computing, Vienna, Austria.

516	Rikardsen, A.H., and Dempson, J.B. 2010. Dietary life-support: the food and feeding of
517	Atlantic salmon at sea. In Atlantic Salmon Ecology. Edited by Ø. Aas, S. Einum, A.
518	Klemetsen and J. Skurdal. Blackwell Publishing Ltd., Oxford, UK. pp. 115-145. doi:
519	10.1002/9781444327755.ch5.
520	Schielzeth, H. 2010. Simple means to improve the interpretability of regression coefficients.
521	Meth. Ecol. Evol. 1(2): 103–113. doi: 10.1111/j.2041-210X.2010.00012.x.
522	Seamons, T.R., and Quinn, T.P. 2010. Sex-specific patterns of lifetime reproductive success
523	in single and repeat breeding steelhead trout (Oncorhynchus mykiss). Behav. Ecol.
524	Sociobiol. 64 (4): 505–513. doi: 10.1007/s00265-009-0866-7.
525	Thorstad, E.B., Økland, F., and Finstad, B. 2000. Effects of telemetry transmitters on
526	swimming performance of adult Atlantic salmon. J. Fish Biol. 57(2): 531-535. doi:
527	10.1111/j.1095-8649.2000.tb02192.x.
528	Ugedal, O., Næsje, T.F., Thorstad, E.B., Forseth, T., Saksgård, L.M., and Heggberget, T.G.
529	2008. Twenty years of hydropower regulation in the River Alta: long-term changes in
530	abundance of juvenile and adult Atlantic salmon. Hydrobiologia 609(1): 9-23. doi:
531	10.1007/s10750-008-9404-2.
532	Ward, D.M., and Hvidsten, N.A. 2010. Predation: Compensation and Context Dependence. In
533	Atlantic Salmon Ecology. Edited by Ø. Aas, S. Einum, A. Klemetsen and J. Skurdal.
534	Blackwell Publishing Ltd., Oxford, UK. pp. 199-220. doi:
535	10.1002/9781444327755.ch8.
536	Webb, J., Verspoor, E., Aubin-Horth, N., Romakkaniemi, A., and Amiro, P. 2007. The
537	Atlantic salmon. In The Atlantic Salmon: Genetics, Conservation and management.
538	Edited by E. Verspoor, L. Strandmeyer and J. Nilsen. Blackwell Publishing Ltd.,
539	Oxford, UK. pp. 17-56.

Table 1. The number, average length, and average mass of tagged Atlantic salmon in Alta River and Fjord in 2007 – 2009, according to sex and tagging technique. Ranges (minimum-maximum) are given in parenthesis.

	Year	2007		2008			2009			
		n	Fork length (cm)	Mass (kg)	n	Fork length (cm)	Mass (kg)	n	Fork length (cm)	Mass (kg)
Radio	Males	-	-	-	25	98	11.9	41	83	10.1*
tagging;						(71-122)	(3.9-23.3)		(54-110)	(4.8-15.0)
tagged	Females	-	-	-	54	96	11.2	46	94	10.0*
before river						(83-116)	(6.5-17.8)		(82-113)	(6.8-17.0)
entry										
Acoustic	Males	27	62	1.6	0	-	-	30	67	2.7
tagging;			(54-82)	(0.9-3.4)					(57-102)	(1.3-10.3)
tagged	Females	33	93	6.1	73	93	6.0	39	97	7.1
before river			(75-115)	(2.6-12.9)		(75-116)	(2.9-11.5)		(89-108)	(4.0-10.2)
exit										
exit										

Note: *mass average based on 28 and 38 individuals, respectively, as data was not available for all tagged individuals

Table 2. Probability (odds ratio with 95% confidence interval - CI) of migrating in the spring (after onset of April) rather than in the autumn (before end of January) in relation with length and body condition while controlling for year, for radio tagged Atlantic salmon (n = 84) in Alta River and Fjord (2007 – 2009).

Predictor	Odds ratio	95% CI	Wald	P	Model
			χ^2		diagnostics*
Intercept	1.9	0.9 – 118.7	0.9	0.4	
Year	5.5	1.3 – 53.0	1.9	0.06	AUC = 0.93
Length	29.9	1.0 – 16596.7	1.5	0.13	$R^2 = 0.65$
Condition	265.2	4.32 – 576885.1	1.7	0.08	

Note: * The area under the receiver operating characteristic curve (AUC) and Naglekerke R² (Nagelkerke 1991) present the predictive and explanatory power of the model, respectively.

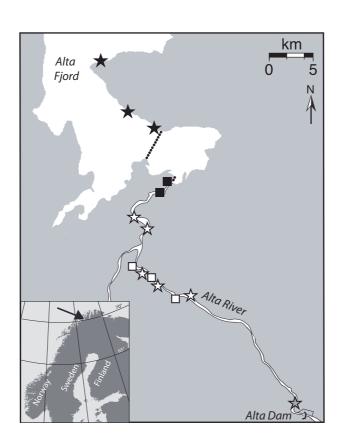
Table 3. Effects of sex, length and body condition, while controlling for year and tag site, on the migration timing (sea entry date) among spring migrants (n=186), for Atlantic salmon tagged with acoustic transmitters in Alta River (2007-2009).

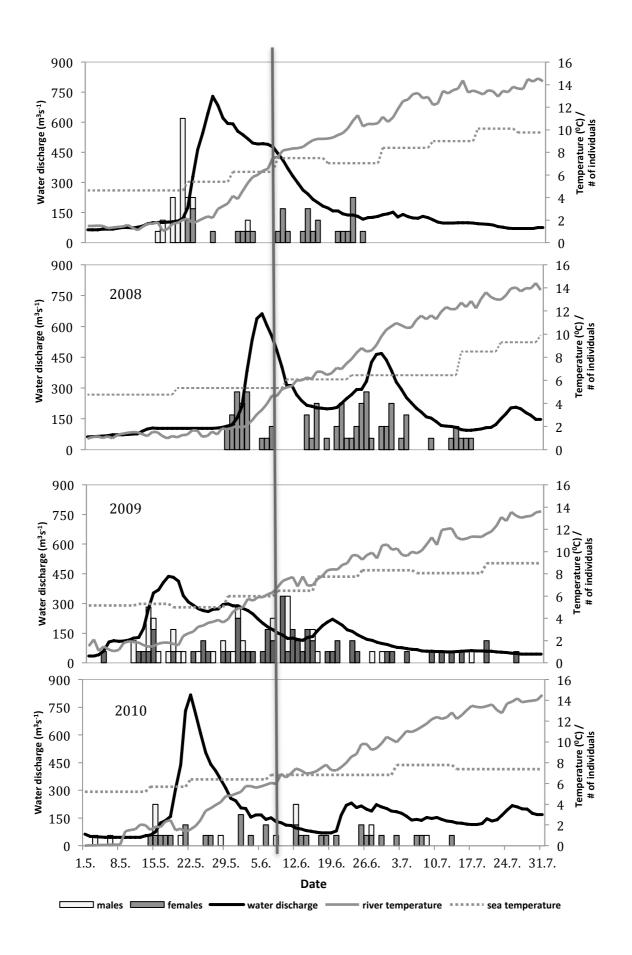
Predictor	Level	Estimate	95% CI	df	F	P
Intercept		40.9	33.2 – 48.7			
Year	2008	5.5	-1.8 – 12.9	2	41.6	<0.001
	2009	-5.8	-12.6 – 0.9			
Tag site	2	-1.4	-9.6 – 6.9	5	0.9	0.5
	3	-1.5	-7.5 – 4.4			
	4	-1.1	-8.4 - 6.2			
	6	-2.2	-10.1 – 5.6			
	7	4.0	-3.0 – 11.0			
Sex	F	4.8	0.1 - 9.4	1	54.2	<0.001
Condition		12.6	8.0 - 17.3	1	50.6	<0.001
Length		4.6	-2.7 – 11.9	1	1.4	0.2
Sex * Condition	F	8.1	0.7 - 15.5	1	7.9	0.005
Sex * Length	F	-8.4	-15.7 – -1.2	1	4.6	0.03
Length * Condition		-6.2	-19.5 – 7.2	1	0.8	0.4

Note: F = female, Level = category level which is associated with the estimate. CI = confidence interval, df = degrees of freedom. Model $R^2 = 0.55$.

Figure captions

- **Fig. 1.** Map of the lower part of the Alta River and Fjord showing the position of the acoustic receiver arrays (dotted line and single dots), the stationary radio loggers (black and white squares; the white squares depicts radio loggers in use only in 2008), and the tagging sites both in the river and the fjord (stars; the black stars mark the tagging sites for radio tagged Atlantic salmon, the white stars for acoustically tagged Atlantic salmon, and the grey star an additional tagging site for radio tagged Atlantic salmon in 2009).
- **Fig. 2**. The marine migration start of acoustically tagged kelts in 2007-2009 in Alta River in relation to water discharge, river temperature, and sea temperature. The grey vertical line marks the median migration date across all years (9 June).
- **Fig. 3.** The probability (\pm SE) for Atlantic salmon kelts in Alta River of migrating in autumn (0) or spring (1) in relation to body condition in 2009-2010. The thin dotted lines underline the body condition and length values where fish had a 50% probability of migrating in spring rather than in autumn. Vertical lines on the x-axis represent the distribution of the data, i.e. each line represents a tagged male.
- **Fig. 4.** Sea entry date (in Julian days using 1 May as day 1) of Atlantic salmon kelts in Alta River in relation to body condition (a) and fork length (b). Lines represent model predictions (± SE), and dots the data.





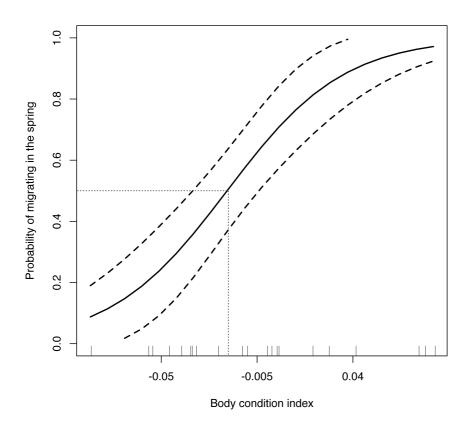


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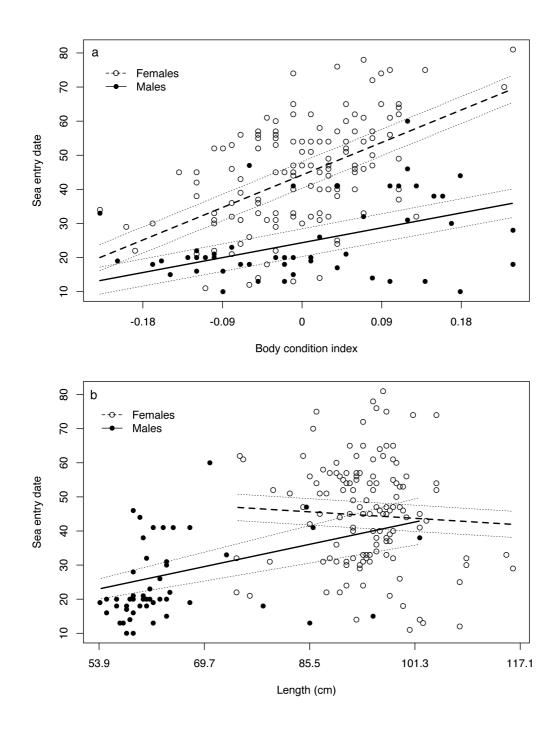


Fig. 4. Sea entry date (in Julian days using 1 May as day 1) of Atlantic salmon kelts in Alta River in relation to body condition (a) and fork length (b). Lines represent model predictions $(\pm SE)$, and dots the data.