

1 ORIGINAL PAPER

2 **Comparative vegetation survey with focus on cryptogamic covers**  
3 **in the high Arctic along two differing catenas**

4

5 **Ramona Kern<sup>1</sup>, Vivien Hotter<sup>1</sup>, Aline Frossard<sup>2</sup>, Martin Albrecht<sup>1</sup>, Christel Baum<sup>6</sup>,**  
6 **Bjorn Tytgat<sup>3</sup>, Lotte De Maeyer<sup>3</sup>, David Velazquez<sup>4</sup>, Christophe Sepey<sup>5</sup>, Beat Frey<sup>2</sup>,**  
7 **Michael Plötze<sup>7</sup>, Elie Verleyen<sup>3</sup>, Antonio Quesada<sup>4</sup>, Mette M. Svenning<sup>5</sup>, Karin Glaser<sup>1</sup>,**  
8 **Ulf Karsten<sup>1\*</sup>**

9

10 <sup>1</sup>University of Rostock, Institute of Biological Sciences, Applied Ecology and Phycology, D-18059 Rostock,  
11 Germany

12 <sup>2</sup>Swiss Federal Research Institute WSL, CH-8903 Birmensdorf, Switzerland

13 <sup>3</sup>Ghent University, Department of Biology, Protistology and Aquatic Ecology, B-9000 Gent, Belgium

14 <sup>4</sup>Universidad Autónoma de Madrid, Department of Biology, E-28049 Madrid, Spain

15 <sup>5</sup>UiT The Arctic University of Norway, Department of Arctic and Marine Biology, Faculty of Biosciences,  
16 Fisheries and Economics, N-9037 Tromsø, Norway.

17 <sup>6</sup>University of Rostock, Soil Sciences, Faculty of Agricultural and Environmental Sciences, D-18059 Rostock,  
18 Germany

19 <sup>7</sup>ETH Zurich, Institute for Geotechnical Engineering, ClayLab, CH-8093 Zürich, Switzerland

20

21

22 **\*Correspondence:**

23 Corresponding Author

24 Prof. Dr. Ulf Karsten, email: ulf.karsten@uni-rostock.de

25

26

27 **Keywords**

28 **Arctic, Svalbard, cryptogamic cover, soil, moisture, tundra, vegetation survey**

29

30

31 **Abstract**

32 Although cryptogamic covers are important ecosystem engineers in high Arctic tundra, they  
33 were often neglected in vegetation surveys. Hence we conducted a systematic survey of  
34 cryptogamic cover and vascular plant coverage and composition at two representative, but  
35 differing Arctic sites (Ny-Ålesund, Svalbard) along catenas with a natural soil moisture

36 gradient, and integrated these data with physical-chemical soil properties. Soil samples were  
37 taken for comprehensive pedological and mineralogical analyses. Vegetation surveys were  
38 conducted based on classification by functional groups. Vascular plants were identified to  
39 species level. Correlation and multivariate statistical analysis were applied to determine the  
40 key environmental factors explaining vegetation patterns along the soil moisture gradients.  
41 We observed significant differences in gravimetric water, soil organic matter and nutrient  
42 contents along the moisture gradients. These differences were coincident with a shift in  
43 vegetation cover and species composition. While chloro- and cyanolichens were abundant at  
44 the drier sites, mosses dominated the wetter and vascular plants the intermediate plots.  
45 Twenty four vascular plant species could be identified, of which only six were present at both  
46 sites. Cryptogamic covers generally dominated with maximum areal coverage up to 70% and  
47 hence should be considered as a new additional syntaxon in future ground-truth and remote  
48 sensing based vegetation surveys of Svalbard. Multivariate analysis revealed that soil  
49 moisture showed the strongest relation between vegetation patterns, together with  $\text{NH}_4\text{-N}$  and  
50 pH. In conclusion, soil moisture is a key driver in controlling cryptogamic cover and  
51 vegetation coverage and vascular plant species composition in high Arctic tundra.

52

## 53 **Introduction**

54 Cryptogamic covers are, together with dwarf shrubs, forbs and graminoids, the dominant  
55 primary producers in High Arctic tundra biomes (Breen and Levesque 2006; Williams et al.  
56 2017). Cryptogamic covers consist of different functional community types such as biological  
57 soil crusts (biocrusts) that are generally considered as an early successional stage dominated  
58 by various microorganisms such as algae and protists, as well as bacteria, archaea and fungi  
59 (Elbert et al. 2012). Later successional stages of cryptogamic covers are dominated by lichens  
60 and mosses, respectively, depending on the water availability (Elbert et al. 2012).  
61 Cryptogamic covers reach an average areal coverage of 50 %, with a high local variability  
62 ranging from 18 up to even 90 %, making them the dominant vegetation type at many High  
63 Arctic locations (Pushkareva et al. 2016; Williams et al. 2017). Despite this, cryptogamic  
64 covers are often neglected in ground-based vegetation surveys and large-scale vegetation  
65 mapping using satellite imagery of Svalbard and other Arctic regions (Johansen et al. 2012;  
66 Johansen and Tømmervik 2014).

67 Cryptogamic covers are formed by living organisms and their by-products, creating a few  
68 millimeters to centimeter thick top-soil layer of inorganic particles bound together by organic  
69 materials. They are often regarded as 'ecosystem-engineers', as they form water-stable

70 aggregates that have important, multi-functional ecological roles in primary production,  
71 nutrient and hydrological cycling, mineralization, weathering, and the stabilization of soils  
72 (Castillo-Monroy et al. 2010). More in particular, on a global scale, cryptogamic covers  
73 significantly contribute to C fixation (about 7 % of the total terrestrial vegetation) and N  
74 fixation (about 50 % of the total terrestrial biological N fixation) (Elbert et al. 2012). Since  
75 both cyanobacteria and algae excrete extracellular polymeric substances (EPS) which glue  
76 soil particles together, they form a carpet-like crust that increases the resistance against soil  
77 erosion by wind and water. By capturing water, cryptogamic covers also control the moisture  
78 content and buffering capacity of soils against temperature fluctuations. As such, cryptogamic  
79 covers influence soil processes, thereby facilitating the colonization of previous barren  
80 substrates by vascular plants (Pushkareva et al. 2016; Williams et al. 2017). Cryptogamic  
81 covers are therefore regarded as an important component in ‘the greening of the  
82 Arctic’(Screen and Simmonds 2010).

83 In the Arctic, water availability depending on habitat (micro)topography is, as elsewhere, a  
84 key driver in controlling the vegetation density and species composition (Zhang et al. 2004).  
85 An illustration of representative Arctic vegetation toposequences are given in Figure 7.  
86 Elevated ridges are generally exposed to wind, so that snow is easily blown away, leaving  
87 behind only a thin snow layer as source of melt water, in turn leading to rather dry ridge soils.  
88 The slopes directly beneath these ridges benefit from meltwater runoff and thus represent  
89 mesic sites. As the snow gets blown away, it accumulates in snow beds below the exposed  
90 ridges. In spring and summer, melting of these snow banks results in a soil moisture gradient  
91 that increases downhill. Eventually, excessive meltwater gathers in depressions, supplying  
92 wetlands, lakes and ponds (Elvebakk 1994; Walker 2000). This moisture gradient is reflected  
93 in the dominant vegetation types in Arctic tundra biomes. Dry exposed ridges are covered  
94 with open vegetation mainly consisting of prostrate dwarf-shrubs such as *Dryas octopetala*  
95 and lichenized biocrusts. In more moist habitats, prostrate dwarf-shrubs like *Salix polaris* and  
96 scattered herbs like *Saxifraga oppositifolia* and *Oxyria digyna* are the dominant vascular  
97 plants. In between these vascular plants cryptogamic covers can reach a surface coverage of  
98 up to 63% (e.g. station Brandal, Williams et al. 2017). Towards the wettest sites, pleurocarp  
99 mosses (and hence moss dominated cryptogamic covers) take over along with grasses and  
100 sedges (Elvebakk 1999). In addition, Pushkareva et al. (2015) reported that the soil water  
101 content shaped the cyanobacterial community composition of Arctic biocrusts. The increase  
102 of soil water content resulted in higher cyanobacterial richness.

103 Not only is the gradient in vegetation functional types directly influenced by the nutrient and  
104 organic carbon concentrations of the underlying soils, the vegetation itself also exerts a strong  
105 control on the remineralization of organic matter by microorganisms present (Berg and  
106 Smalla 2009; Vimal et al. 2017). In general, wet tundra is characterized by higher N and C  
107 contents compared to dry systems, but the available information is contradictory between  
108 studies, probably as a result of patchiness in vegetation types and soil properties (Chapin III  
109 and Shaver 1981; Edwards and Jefferies 2013).

110 We aimed to assess the relation between physico-chemical soil parameters along catenas on  
111 the composition and coverage of cryptogamic covers and vascular plants in the High Arctic  
112 tundra. Both catenas ranged from a wet site (wetland or close to a lake) to a hill or an elevated  
113 ridge, respectively, at two sampling sites (Knudsenheia (KH) and Ossian-Sarsfjellet (OS))  
114 (see also Fig. 7). Both sites represent typical settings in the High Arctic: one is a coastal plain  
115 with a soft slope towards a wetland (KH), the other an elevated ridge with a steep slope  
116 towards a permanent lake (OS). We assumed that soil moisture is one of the key factors  
117 influencing the vegetation type. As the water availability influences and is influenced by  
118 various soil parameters and vegetation, we conducted in-depth analyses of various  
119 pedological parameters as well as vegetation surveys including cryptogamic covers and  
120 vascular plants.

121

## 122 **Materials and Methods**

### 123 **Study sites**

124 The Ny-Ålesund Research Station (Svalbard, Norway, 78°55'26.33''N, 11°55'23.84''E), with  
125 contributions from many institutions and countries, is a model system for the High Arctic.  
126 Ny-Ålesund represents a coastal terrestrial environment, which is characterized by a variety of  
127 different geological features, soil and glacier types, and hence habitats such as polar semi-  
128 desert, wet moss tundra, and ornithogenic soils. Because of the West Spitsbergen Current,  
129 which flows along the West coast of Svalbard and transports warm Atlantic water masses into  
130 the Arctic Ocean, Ny-Ålesund shows relatively mild climatic conditions compared to other  
131 regions at the same latitude. A weather station was established in July 1974 by the Norwegian  
132 Meteorological Institute ([www.met.no](http://www.met.no)), which is located 8 m a.s.l., 100 m away from Ny-  
133 Ålesund. The meteorological data over the last two decades show a mean summer and winter  
134 temperature of 8 °C and -14 °C, respectively. However, longer cold periods between -20 °C  
135 and -35 °C can occur during winter. The annual precipitation over the last two decades  
136 averages 470 mm with 70 % typically falling between October and May, when snow cover is

137 usually complete, while the other 30 % are typically represented by scattered rain. Two sites  
138 were selected in the study area and established as permanent sampling sites, namely (1)  
139 Knudsenheia (KH), a wetland located approximately three km north-east of Ny-Ålesund, and  
140 (2) Ossian-Sarsfjellet (OS), a Nature Reserve approximately 12 km north-west of Ny-Ålesund  
141 across Kongsfjorden (Figs 1, 2). At each site a catena was established, which represent two  
142 different common settings in High Arctic tundra. KH is a typical coastal plain with a wetland  
143 at the flattest point (26 – 36 m a.s.l.). In contrast, the catena in OS ranges from a permanent  
144 lake to an elevated ridge (100 – 113 m a.s.l.), a typical setting for inland areas (Fig. 1). More  
145 details on cryptogamic cover vegetation types along changing altitudes in Svalbard are given  
146 by Williams et al. (2017).

147

#### 148 **Experimental design and sampling**

149 The catena in KH starts from the north to north-eastern littoral zone of a shallow pond and  
150 runs towards the south-southwest along a gradient in decreasing soil moisture (Figs 1C, 7). In  
151 OS, the wet plots are situated on the north to north-western shore of the lake Sarsvatnet. The  
152 catena was installed along a north by western orientation and culminates in a dry exposed  
153 ridge (Fig. 1D, 7).

154 Along each catena, three sub-sites were selected, namely dry, intermediate and wet. In each  
155 sub-site, three permanent replicate plots of 1 m<sup>2</sup> square were established (3 x 3 replicate plots  
156 per catena) (Table 1). Each 1 m<sup>2</sup> plot was further divided into four quadrats (50 x 50 cm) of  
157 which three were randomly used for soil sampling. Two different soil depths were  
158 consequently sampled with a sterilized spoon: the top layer (0-1 cm) and the subsoil (5-10  
159 cm). In total, 54 soil samples (3 sub-sites differing in soil moisture content x 3 plots  
160 (replicates) x 3 quadrats x 2 soil depths) were collected per site. Each soil sample was filled  
161 into sterile plastic bags wherein the samples were homogenized by hand. Subsamples of these  
162 pooled samples were dried at 60 °C within 24 h after collection for subsequent soil analyses.  
163 The sampling campaign took place in summer 2017.

164

#### 165 **Pedological characterisation**

166 Directly next to each dry, intermediate and wet plot a soil profile was excavated (about 1 m  
167 distance to the respective plots) as a rectangular pit down to 40 cm depth. Particularly the  
168 thickness of the O and A horizons was visually inspected based on characteristic colour  
169 changes and measured using a ruler. Soils were classified according to IUSS Working Group  
170 WRB (2015) protocol (The International Union of Soil Sciences, <https://www.iuss.org>).

171 For water content determination of each sample, about 20 g of fresh soil from one of the 50 x  
172 50 cm quadrats were sieved (2 mm mesh) and weighed. Afterwards, the soil was dried at 105  
173 °C overnight. After weighing again, the gravimetric soil water content was calculated. This  
174 dried soil fraction was subsequently combusted at 450 °C for 5 h for the assessment of the  
175 amount of the soil organic matter content. The moisture content was expressed as percentage  
176 of total fresh mass and the organic matter content percentage of total dry mass.

177 The pH of each soil sample was measured in an aqueous soil-extract (soil:aqua ratio of 1:2)  
178 with a glass electrode connected to a pH meter (FEP20-FiveEasy Plus, Mettler-Toledo  
179 GmbH, Switzerland).

180 Soil texture (percentage of sand, silt, clay) of each soil sample was determined following the  
181 „sieve-pipette“ approach (Gee and Bauder 1986). This method is a combination of wet  
182 sieving of the fraction  $>63 \mu\text{m}$  and the pipette sampling method for the silt ( $2\text{--}63 \mu\text{m}$ ) and  
183 clay ( $<2 \mu\text{m}$ ) fractions. In a column, the sediment concentration, as a function of time, was  
184 monitored by timed withdrawals of samples with a pipette at certain heights and at a constant  
185 temperature. The sieve-pipette method measures the mass percentage for the defined grain-  
186 size classes.

187 For nutrient analysis, dried soil subsamples (see above) were sieved (2 mm mesh). These  
188 subsamples were stored at room temperature prior to further analysis. For ammonia and  
189 nitrate analysis 0.5 g of dried and ground soil was extracted with 20 mL of  $0.01 \text{ mol L}^{-1} \text{ CaCl}_2$   
190 for 2 h on a vertical shaker. Afterwards, each extract was filtered with a GF92 glass-fiber  
191 filter (Whatman) and the filtrate was frozen at  $-20 \text{ °C}$  until measurement with a continuous  
192 flow analyser (Alliance Instruments, Salzburg, Austria) using the manufacturer's protocol for  
193 both compounds. Two soluble labile inorganic phosphate fractions according to Hedley et al.  
194 (1982) were extracted by a two-step fractionation scheme, the first consisting of the water-  
195 extract and the second one of the bicarbonate-extract. Five grams of pooled dried and ground  
196 soil samples were transferred into 20 mL of ultra-pure de-ionized water and incubated on a  
197 vertical shaker for 24 h. The tubes were then centrifuged for 5 min at 5,000 rpm (Megafuge,  
198 Heraeus) and the supernatant was filtrated with glass-fiber filters (MN 616 G - phosphate-  
199 free), resulting in the water-extract. The soil-pellet was re-suspended in 20 mL  $0.5 \text{ mol L}^{-1}$   
200  $\text{NaHCO}_3$  solution and put again onto a vertical shaker for 24 h, followed by centrifugation  
201 and filtration as in the first extraction step. The bicarbonate-filtrate was neutralized (pH 7)  
202 prior to measurement. The filtrates and neutralized filtrates were then measured for their P  
203 concentrations using the colorimetric molybdenum blue method at  $\lambda = 885 \text{ nm}$  (Murphy and  
204 Riley 1962). The soil total carbon (TC) and total nitrogen (TN) were determined from these

205 dried and ground soil subsamples by dry combustion using a CNS VARIO EL analyser  
206 (Elementar Analysensysteme GmbH, Germany).

207 Mineralogy of bulk samples was determined on randomly oriented powder specimens with X-  
208 ray diffraction (XRD) analysis. The samples were air dried, crushed in a jaw breaker <400 µm  
209 and split representatively. An aliquot of about 2 g was milled in ethanol to a grain size below  
210 20 µm with a McCrone micronizing mill and dried afterwards at 65 °C. For frontloading  
211 preparation, about 1 g of the powdered material was gently pressed in a sample holder for  
212 packing, sample-height adjustment and forming a flat surface. Preferred orientation was  
213 avoided by using a blade for surface treatment. A second sample preparation was carried out  
214 producing oriented specimens for enhancement of the basal reflexes of layer silicates, thereby  
215 facilitating their identification. The changes in the reflex positions in the XRD pattern by  
216 intercalation of different organic compounds (e.g. ethylene glycol) and after heating were  
217 used for identification in particular of smectite.

218 X-ray diffraction measurements were conducted with a Bragg-Brentano X-ray diffractometer  
219 (D8 Advance, Bruker AXS, Germany) using CoK $\alpha$  (35 kV, 40 mA) radiation. The instrument  
220 was equipped with an automatic theta compensating divergence slit and a Lynx-Eye XE-T  
221 detector. The powder samples were step-scanned at room temperature from 2 to 80°2Theta  
222 (step width 0.02°2Theta, counting time 2 s per step). The qualitative phase analysis was  
223 carried out with the software package DIFFRACplus (Bruker AXS). The phases were  
224 identified on the basis of the peak positions and relative intensities in the comparison to the  
225 PDF-2 data base (International Centre for Diffraction Data).

226 The quantitative amount of the mineral phases was determined with Rietveld-analysis. This  
227 full pattern-fitting method consists in the calculation of the X-ray diffraction pattern and its  
228 iterative adjustment to the measured diffractogram. In the refinements phase specific  
229 parameters and the phase content were adapted to minimize the difference between the  
230 calculated and the measured X-ray diffractogram. The quantitative phase analysis was carried  
231 out with Rietveld program Profex/BGMN (Döbelin and Kleeberg 2015).

232

### 233 **Vegetation survey**

234 All vegetation surveys were carried out, together with the soil sampling, in late July and early  
235 August 2017. At both field sites, each of the 18 permanent sampling plots was evaluated by  
236 manual inspection and documentation using a digital camera (see Online Resource 1 and 2).  
237 Additionally, three vegetation survey plots of 25 x 25 m were established along the soil  
238 moisture gradient in both field sites. The vegetation of these plots was recorded following the

239 point intercept method (Levy and Madden 1933) to determine the proportions of eight ground  
240 cover functional groups according to the approach of Williams et al. (2017). These included  
241 biocrusts (typically dominated by cyanobacteria, which cause a dark color), chlorolichen  
242 (with an algal photobiont), cyanolichen (with a cyanobacterial photobiont), moss, vascular  
243 plant, litter (dead plant material, reindeer and goose droppings), rock, and bare soil. Litter,  
244 rock, and bare soil were later on summarized as ‘unvegetated area’. Twenty-five squares of 25  
245 x 25 cm (=625 cm<sup>2</sup>) were randomly selected within each established vegetation survey plot,  
246 and the functional groups in each square were determined by 25 point measurements (Levy  
247 and Madden 1933). In total, 625 point measurements per vegetation survey plot were  
248 undertaken. The vascular plant species on the vegetation survey and the experimental plots  
249 were determined after Rønning (1996), and the names corrected according to the Plant List  
250 2013 ([www.theplantlist.org](http://www.theplantlist.org)).

251

## 252 **Statistical analyses**

253 All statistical analyses were done using R version 3.4.0 (R-Development-Core-Team 2017).  
254 The mean of the replicate quadrants (see above) was calculated and used for further statistical  
255 analysis. After a Shapiro-Wilk test for normality, analysis of Variance (one-way ANOVA)  
256 was performed to reveal significant differences of the measured soil parameters between the  
257 subsites (wet, intermediate, dry) in both regions, with a threshold of significance at 95 %. The  
258 soil parameters were normalized ( $X_{norm}=(X_i-X_{min})/(X_{max}-X_{min})$ ) for cluster and multivariate  
259 analyses. A cluster analysis based on the Bray-Curtis dissimilarity was conducted to visualize  
260 differences within and between the sites according to the measured soil parameters using the  
261 Vegan package (Oksanen et al. 2018) implemented in R.

262 With the data obtained via the point intercept method, the percentage areal coverage by each  
263 functional group was calculated for every plot and displayed in a stacked bar plot. Moreover,  
264 differences in vegetation coverage between the plots were visualized by non-metric  
265 multidimensional scaling (nMDS) using the Vegan package and Bray-Curtis dissimilarity  
266 index implemented in R. To reveal correlations between the ground coverage of the  
267 vegetation classes and soil parameters, permutational multivariate analysis of variance  
268 (PERMANOVA) (with the “adonis” function in R) was applied using the Bray-Curtis  
269 dissimilarity index, including a permutation test with 1000 permutations. Soil parameters that  
270 were significantly correlated with vegetation ground cover were added to the plot.  
271 Subsequently, the data on ground coverage were statistically analyzed via pairwise  
272 PERMANOVA implemented in the RVAideMemoire package (Hervé 2018) followed by



273 Bonferroni correction to compare the ground coverage composition along the transect and  
274 among the investigated sites. The presence/absence data of the vascular plants were visualized  
275 with a Venn diagram using the Venn diagram package (Dusa 2017) implemented in R.  
276

## 277 **Results**

### 278 **Physical and chemical soil properties**

279 The cluster analysis based on all physical and chemical soil parameters (Table 1) revealed a  
280 clear separation between the KH and OS sites, as well as between the dry, intermediate and  
281 wet plots in both regions (Fig. 3). One exception is KHd.1 which clustered separately from  
282 the other dry KH plots, because of its much lower moisture content (14.7 % compared to 35.8  
283 and 33.4 %, respectively). Especially for OS, the cluster analyses revealed a clear difference  
284 between each subsite (dry, intermediate and wet) but a close similarity between the replicate  
285 plots. The differences between the subsites are also reflected in the mean values of the  
286 physical and chemical soil parameters (Table 1).

287 The different KH plots were characterized by higher sand (75.8-81.4 %) and lower silt content  
288 (6.5-13 %) compared to the OS subsites (61-71.7 % sand and 19.4-28.2 % silt). The clay  
289 content was more or less similar between both sites (7.5-14.8 %) (Table 1). The gravimetric  
290 soil water content in KH ranged from 28.0 % (of the wet weight) in the dry plots, to 63.4 % in  
291 the intermediate plots and 70.0 % in the wet plots (Table 1). The high difference in water  
292 content might be explained by the lower amount of clay in the dry plots compared to the  
293 intermediate and wet plots. In OS, the soil water content along the moisture gradient was  
294 generally lower due to its elevation. While the dry plots exhibited 23.5 % of soil water  
295 content, the intermediate and wet plots had 34.2 and 48.3 %, respectively (Table 1). Except  
296 for the dry plots in KH, which had a soil pH of 5.5, the intermediate and wet plots exhibited  
297 pH-values between 6.8 and 7.1, respectively. In OS, the pH in the dry plots was with 7.1  
298 higher compared to the intermediate (6.4) and wet plots (6.9) (Table 1). The soil organic  
299 matter (SOM) content was higher in the subsites in KH compared to those in OS. In KH the  
300 dry plots exhibited 29.4 % SOM (of the dry weight), whereas the intermediate and wet plots  
301 had SOM values of 43.8 and 39.2 %, respectively (Table 1). In contrast, the SOM in OS  
302 varied from 12.9 % in the dry plots to 31.4 % in the intermediate plots. Here, the wet plots  
303 contained 15.8 % SOM (Table 1). TN values were lower in soil samples in the dry plots of  
304 KH and OS (0.66 and 0.34 %, respectively) compared to the intermediate (1.15 and 1.00 %,  
305 respectively) and wet (1.06 and 0.50 %, respectively) plots of both sites (Table 1). By  
306 contrast, the soil TC values were highest in the intermediate plots. While in KH the TC

307 content was 22.79 %, OS showed 18.06 %. Soils of the dry plots in both sites contained 13.22  
308 and 6.15 % TC, respectively. The wet plots showed 18.51 % TC in KH and 7.95 % in OS.  
309 The TC values well reflected the SOM data (Table 1).

310 The  $\text{NH}_4\text{-N}$  contents were always higher than those of  $\text{NO}_3\text{-N}$ . The  $\text{NH}_4\text{-N}$  values ranged  
311 along the water availability gradient between 30.31 and 49.17  $\text{mg kg}^{-1}$  dry weight in KH and  
312 between 25.45 and 69.61  $\text{mg kg}^{-1}$  dry weight in OS with a tendency of higher amounts in the  
313 dry subsites. The  $\text{NO}_3\text{-N}$  contents ranged from 16.24 to 48.65  $\text{mg kg}^{-1}$  dry weight in the soil in  
314 KH and from 4.97 to 30.71  $\text{mg kg}^{-1}$  dry weight in OS. The OS intermediate and wet plots  
315 exhibited with 46.78 and 10.34  $\text{mg kg}^{-1}$  dry weight much lower values compared to the dry  
316 plots (23.62  $\text{mg kg}^{-1}$  dry weight) (Table 1). In contrast to both nitrogen compounds,  $\text{P}_{\text{labile}}$   
317 contents were always much lower with values between 3.02 and 4.97  $\text{mg kg}^{-1}$  dry weight in  
318 KH, and between 1.82 and 2.59  $\text{mg kg}^{-1}$  dry weight in OS (Table 1).

319 The O horizon varied in thickness between 1 and 4 cm among the different plots, i.e. each soil  
320 was conspicuously covered by organic material. The depth of the respective soil horizon is  
321 given in Table 1. The A horizon consisted mainly of dark decomposed organic materials  
322 (humus) and was thinner in KH (>8 and >13 cm) compared with OS (between >20 and >22  
323 cm depth) (Table 1).

324

### 325 **Mineralogical soil properties**

326 Quartz was the dominant mineral in all soils, ranging from 47.6 to 73.8 % of the dry weight in  
327 KH and from 33.6 to 56.8 % in OS (Table 2). The dry plots in both sites always showed the  
328 highest percentage of Quartz (Table 2). Dolomite/Ankerite was the second most abundant  
329 mineral and varied between 8.0 and 31.7% of the dry weight in KH and between 4.8 to 22.6%  
330 in OS. Na-Plagioclase was present in medium concentrations ranging from 5.0 to 9.2 % at  
331 KH, and from 7.1 to 13.3 % at OS. Calcite, Muscovite and Biotite were present in much lower  
332 concentrations at KH (0.6 to 3.8 %) compared to OS (3.0 to 12.5 %), while Chlorite and K-  
333 Feldspar occurred in low values (Table 2).

334

### 335 **Vegetation and cryptogamic cover survey**

336 Biocrusts were the dominant vegetation form in both sites, whereas cyanolichens were sparse  
337 (Fig. 4 A, B). In the wet plots of KH up to 40 % of the surface was overgrown by biocrusts. In  
338 OS, chlorolichens were the second most dominant functional group, which were twice as  
339 abundant compared to KH. Mosses showed a reverse pattern and were the second most  
340 abundant vegetation type in KH with an occurrence twice of that in OS. One sixth of the

341 surface in both sites was covered with vascular plants. Unvegetated area was, however, more  
342 dominant than biocrusts in OS, and even twice as abundant in KH (Fig. 4 A).

343 Vegetation ground cover composition in the dry, intermediate and wet plots significantly  
344 differed in each of the two sites, as indicated by pairwise PERMANOVA ( $p \leq 0.001$ ; Fig. 4  
345 B) and nMDS (Fig. 5). In addition, each of the three plots in KH differed significantly from  
346 the respective subsite in OS ( $p \leq 0.01$ ). Multivariate analysis of the vegetation classes and the  
347 respective soil parameters for each plot in KH and OS revealed that wet plots from both sites  
348 were quite similar to each other and dominated by moss (e.g. *Racomitrium lanuginosum*) and  
349 biocrusts (Fig. 5). However, large differences in the ground cover composition were observed  
350 between the two sites KH and OS for the dry plots and also, although not as prominent, for  
351 the intermediate plots.

352 The dry plots of both catenas were about one third unvegetated (stones, bare soil, litter), while  
353 mosses were almost absent. In OS, vascular plants covered another third of the dry area and  
354 were three times as common as in KH, where they covered only 10 % of the soil surface. In  
355 OS biocrusts and chlorolichens appeared in equal amounts but were a bit scarcer than in KH.  
356 In KH, cyanolichens were as numerous as biocrusts and chlorolichens, and three times as  
357 frequent as in OS (Fig. 4 B). In the intermediate plots, KH was dominated by biocrusts.  
358 Mosses, vascular plants and unvegetated area were equally common, whereas lichens covered  
359 a small surface. In OS, ground coverage in the intermediate plots was completely different. It  
360 was dominated by vascular plants and unvegetated area. Unlike in KH, cyanolichens but also  
361 mosses were scarce. Chlorolichens, however, made up almost one fifth of the total area and  
362 were three times as frequent as in the KH intermediate plots (Fig. 4 B). Biocrusts were  
363 prevailing in the wet plots of both field sites. Mosses covered about one third of the wet  
364 ground in KH, and one fifth in OS. Chlorolichens covered one fifth of the ground in OS,  
365 whereas in KH they were negligible. Vascular plants covered one fifth in KH, but were scarce  
366 in OS. Cyanolichens were almost absent in both field sites (Fig. 4 B). In summary, a  
367 significant shift in ground cover composition could be observed along the catenas. Based on  
368 pairwise PERMANOVA in both field sites KH and OS, biocrusts ( $p \leq 0.05$ ) and mosses ( $p \leq$   
369  $0.001$ ) increased with increasing soil moisture, whereas cyanolichens ( $p \leq 0.05$ ) and  
370 unvegetated area ( $p \leq 0.01$ ) decreased. Chlorolichens also decreased, but this was significant  
371 only in KH ( $p \leq 0.001$ ). In OS, vascular plants decreased with increasing soil moisture ( $p \leq$   
372  $0.001$ ), while only an increasing yet statistically insignificant trend ( $p > 0.05$ ) was observed in  
373 KH.

374 Three soil parameters were significantly correlated with the change in the vegetation cover  
375 (PERMANOVA), namely moisture (explained variance: 30 %,  $p=0.001$ ), pH (explained  
376 variance: 20.5 %,  $p=0.002$ ) and ammonium concentration (explained variance: 13 %,  $p=0.012$ ).

378 Altogether, 24 different vascular plant species belonging to 11 families could be observed  
379 (Table 3). Six of these were present in both field sites, 14 species were exclusive to KH, and  
380 four species were only observed in OS (Table 3, Fig. 6). In KH, the species richness was  
381 higher in the dry and intermediate plot (13 species) compared to the wet plot (nine species). In  
382 OS, all plots harbored six to seven different plant species (Table 3, Fig. 6). The dwarf shrubs  
383 *Saxifraga oppositifolia* and *Salix polaris* were present in both study sites and almost all plots,  
384 whereas the graminoid *Luzula nivalis*, as well as the forbs *Cerastium arcticum*, *Draba alpine*,  
385 *Minuartia rubella*, *Papaver dahlianum* and *Pedicularis hirsuta* could only be observed in the  
386 dry plot of KH (Online Resource 1). A summarizing scheme of both Arctic vegetation  
387 toposequences along the catenas and soil moisture gradients in KH and OS is shown in Figure  
388 7.

389

## 390 **Discussion**

### 391 **Soil properties - carbon**

392 Soil organic matter (SOM) values along both moisture availability gradients ranged between  
393 29.4 and 43.8 % of the dry weight in KH, and between 12.9 and 31.4 % in OS. Interestingly,  
394 the intermediate plots had higher SOM values compared to the wet and dry sites. This is in  
395 agreement with studies from Arctic tundra soils in northern Alaska (Mercado-Díaz et al.  
396 2014) and central Northwest Territories in Canada (Chu and Grogan 2010), which reported  
397 similar SOM values of 29.2-34.9 % and 34.5-46.5% SOM, respectively. The corresponding  
398 TC contents for each plot were always approximately half of those of the SOM (Table 1).  
399 Arctic tundra vegetation is characterized by a significant transfer of fixed C below ground  
400 into storage organs (e.g. roots, rhizomes, tillers etc.) at the end of the growing season as part  
401 of their energy conservation and overwintering mechanism. Consequently, most of the plant C  
402 ends up in the soil (e.g. through litter or exudates), where it is recycled through  
403 microbiological activity, gradually being released by respiratory processes and thus returning  
404 to the atmosphere. Therefore, high vegetation coverage leads to enhanced SOM accumulation.  
405 However, the process of decomposition in the Arctic is generally very slow, mainly because  
406 of low temperatures, as well as due to a lack of moisture in well drained soils or excess water  
407 where drainage is inhibited (Harden et al. 2012). Plant-derived SOM gradually accumulates,

408 forming more mature soils, or in wetlands such as bogs, lack of oxygen through waterlogging,  
409 causes formation as peat.

410

#### 411 **Soil properties - nitrogen**

412 In contrast to the large amount of soil organic C, Arctic soils store 8-15 Gt N which equals  
413 about 10 % of the global soil N content (Loisel et al. 2014). The TN contents in the present  
414 study ranged between 0.34 and 1.15 % of the soil dry weight with concomitant rather low  
415  $\text{NH}_4\text{-N}$  (31.25 – 56.89  $\text{mg kg}^{-1}$  soil dry weight) and  $\text{NO}_3\text{-N}$  (6.78 – 29.84  $\text{mg kg}^{-1}$  soil dry  
416 weight) amounts (Table 1), which are comparable to both nutrients in the Canadian tundra  
417 (Chu and Grogan 2010) and in cryosols from Siberia and Greenland (Wild et al. 2013). The  
418 C/N ratio (calculated from Table 1) ranged between 16 and 20, and indicated clear N  
419 limitation at all study sites, because the typical C/N stoichiometry for soils on a global scale is  
420 around 14 and those of soil microbial biomass between 8 and 9 (Cleveland and Liptzin 2007).  
421 The data of the present study agrees with Chapin et al. (2011) who assumed that N limitation  
422 is most common in Arctic ecosystems. The C/N was lowest at both wet study plots (16 to 17)  
423 compared to the dry and intermediate test plots with a C/N ratio of 18 to 20. As mineralization  
424 rates are generally low in Arctic biomes, only small proportions of this N are bioavailable  
425 (Wild et al. 2013). In addition, N availability also controls rates, directions and magnitudes of  
426 C fluxes in Arctic ecosystems under increasing warming (Chu and Grogan 2010), i.e. the soil  
427 C- and N-cycle are strongly interlinked. Recently,  $\text{NO}_3\text{-N}$  was reported to be an important N  
428 source for Arctic tundra plants (Lui et al. 2018). Consequently, the intermediate plot of OS  
429 with the lowest  $\text{NO}_3\text{-N}$  content (6.78  $\text{mg kg}^{-1}$  soil dry weight) (Table 1), might have the  
430 strongest N limitation for decomposition. This is in agreement with the largest SOM (31.4 %)  
431 accumulation at this plot within the OS area. Microbial mineralization of SOM is regarded as  
432 a main source for the annual N mobilization in Arctic soils (Schimel and Bennett 2004),  
433 which is supplemented by the biological fixation of atmospheric N (Hobara et al. 2006), as  
434 well as by atmospheric deposition of inorganic N compounds (Van Cleve and Alexander  
435 1981). Nevertheless, the annual N-requirements of the Arctic vegetation is about 2-3 times  
436 higher compared to all N-mobilization and input processes (Shaver and Chapin 1991), which  
437 supports the general view of N-limitation of Arctic vegetation (Reich et al. 2006). However,  
438 many of the calculations on N-budgets in Arctic soils were undertaken rather locally and  
439 already decades ago, and hence do not well reflect recent environmental changes in the  
440 tundra.

441

**442 Soil properties - phosphorus**

443 Although P is at least as important as N for the Arctic tundra vegetation (Giesler et al. 2012;  
444 Zamin and Grogan 2012) and soil microorganisms (Gray et al. 2014), it is not well understood  
445 how much P is available in Arctic soils. The present study revealed very low available P  
446 contents in Svalbard soils, ranging from only 1.8 to 5.0 mg kg<sup>-1</sup> dry weight (Table 1). Other  
447 Arctic soils such as in Canada or Alaska contain much higher P amounts between 17 and  
448 several hundred mg kg<sup>-1</sup> dry weight (Mercado-Díaz et al. 2014; Keller et al. 2007). Chemical  
449 and biological weathering of primary minerals like apatite is the main input of P in Arctic  
450 soils. This difference between our sites and other regions might thus be related to the soil  
451 mineral composition. In both KH and OS, the mineral composition is dominated by quartz,  
452 chlorite and plagioclase which are minerals lacking P. More detailed studies on  
453 biogeochemical cycling and budgets of P in Arctic soils are urgently needed.

454

**455 Vegetation and cryptogamic cover survey along the catenas**

456 The catenas in KH and OS differed in their overall areal cover by functional vegetation types  
457 (Fig. 4). Moisture content reflecting the local topography was significantly related to these  
458 community changes in vegetation (Fig. 5). Soil moisture in summer is mainly dependent on  
459 the soil structure, thawed permafrost layer and height above the water level in our catenas  
460 (permanent lake in OS, wetland in KH), since precipitation at that season is low (see Material  
461 and Methods) and melt water is only important in May and June. According to Elvebakk  
462 (1999), Arctic vegetation and topography are strongly correlated, since topography,  
463 influences water availability, in particular water runoff, which itself is strongly influenced by  
464 the vegetation type. Cryptogamic covers, in particular biocrusts and lichens, shape the soil  
465 surface by protecting fine-grained material from water erosion, thereby acting as water  
466 barriers for the underground layers leading to a higher runoff while mosses more likely trap  
467 water. Further, it is important to mention, that rooting depth of vascular plants is limited in  
468 Arctic soils because of permafrost and the relatively low A-horizon (maximum depth in KH  
469 13 cm, in OS 22 cm; Table 1), thus, mainly top soil moisture content shapes the vegetation.

470 A complete High Arctic toposequence consists of dry exposed ridges, mesic slopes and zonal  
471 snow beds, and ends up in a wet area. From the exposed ridge to the wet site, soil moisture  
472 increases, which affects the vegetation. Ridges and slopes are dominated by prostrate dwarf  
473 shrubs and rosette herbs, snow beds by forbs, and wet sites by mosses, grasses and sedges.  
474 Moreover, with increasing water availability, the vegetation becomes denser (Elvebakk 1999;  
475 Walker 2000). Apart from some small differences, we found a more or less similar vegetation

476 pattern in our study sites (Fig. 7). Herbs and lichens dominated the dry plots in KH, while  
477 prostrate dwarf shrubs were almost completely absent, which are the prevailing vegetation  
478 type on exposed ridges and mesic slopes according Elvebakk (1999). The absence of prostrate  
479 dwarfs in KH is likely related to the toposequence starting directly with a snow bed rather  
480 than an exposed ridge or a mesic slope. In addition, the intermediate plots all lie within the  
481 same topographic entity as the wet site. This is reflected in a similar vegetation composition,  
482 which is dominated by mosses and biocrusts (Fig. 7) and the lack of a significant increase in  
483 moisture content from the intermediate to the wet plots (Table 1).

484 In contrast to KH, OS exhibited a complete High Arctic toposequence, consisting of exposed  
485 ridges, mesic slopes and zonal snow beds ending up in a wet area along the moisture transect  
486 (Elvebakk 1994). The prostrate dwarf shrubs such as *Dryas octopetala* and *Cassiope*  
487 *tetragona*, which are typical for exposed ridges and slopes, dominated the plant communities  
488 in the dry and intermediate plots. Biocrusts, mosses and lichens were the dominant vegetation  
489 in the wet plots (Fig. 7). These vegetation patterns are well reflected in the moisture content  
490 of the top soils (Table 1), as the soil water content significantly increased from the dry to the  
491 intermediate plots in both field sites. The dominance of cyanobacteria-dominated biocrusts at  
492 the wet plots can be explained by the dominant form of atmospheric water supply being a key  
493 driver of biocrust community structure - while terrestrial green algae can use water vapor as  
494 the only water source, liquid water (rain or melt water) is a prerequisite for the development  
495 of cyanobacteria (Lange et al. 1986). The conspicuously different vegetation between KH and  
496 OS can be explained by differences in the toposequence including site-specific physical and  
497 chemical parameters, but also by regional microclimatic conditions as a result of the  
498 difference in exposure of the transects. KH is an open plain with the glaciers Vester  
499 Brøggerbreen and Mørebreen to the South and West, and Kongsfjorden to the North. Strong  
500 katabatic winds from the glaciers towards the sea are quite common and likely have a cooling  
501 effect on this study site. OS on the other hand is relatively sheltered by surrounding mountain  
502 ridges and hence might have a milder climate than KH. This is also evident from differences  
503 in the vascular plant species composition between both sites. More in particular, OS is  
504 protected as a nature reserve because it is the most northern limit of a number of vascular  
505 plants (e.g. *Comastoma tenellum*, *Tofieldia pusilla*) in Svalbard as a result of its particular  
506 microclimatic conditions (Birkeland et al. 2017).

507 The plant community around Ny-Ålesund (including KH) was described as the *Cetrariella*  
508 *delisei-Saxifraga oppositifolia* association within the *Luzulion nivalis* alliance (Elvebakk  
509 1994; Øvstedal et al. 2009). In the dry plots of KH, *Saxifraga oppositifolia* and a lichen

510 strongly resembling *Cetrariella delisei* indeed grew extensively. For OS, the *Dryado-*  
 511 *Caricetum rupestris* and *Cassiopo tetragonae-Dryadetum octopetalae* associations were  
 512 reported (Elvebakk 1994). Both associations belong to the *Caricion nadinae* alliance and are  
 513 typical for exposed ridges and mesic slopes, respectively (Elvebakk 1994). These findings are  
 514 in accordance with the vegetation in OS: *D. octopetala* and *Carex rupestris* appeared to be  
 515 very numerous in the dry plots, and *Cassiopo tetragona* and *D. octopetala* in the intermediate  
 516 plots. Hence, a *Dryado-Caricetum rupestris* association in the dry plots seems to shift towards  
 517 a *Cassiopo tetragonae-Dryadetum octopetalae* association in the intermediate plots along the  
 518 OS transect. All vegetation communities found in KH and OS prefer slightly acidic to slightly  
 519 alkaline substrate (Elvebakk 1994; Øvstedal et al. 2009), which corresponds to the measured  
 520 pH values from soils along the transect (Table 1).

521 Chlorolichens were generally more abundant than cyanolichens and covered up to 20 % of the  
 522 dry plots in KH, which is in agreement with other sites on Svalbard (Williams et al. 2017).  
 523 Both lichen groups differ in their ecosystem functions. Chlorolichens are known as soil  
 524 stabilizer and effective preventer of soil erosion, as high primary producer already under high  
 525 air humidity alone and producer of C-rich metabolites that can be leached into the soil  
 526 (Williams et al. 2017). In contrast, cyanolichens are less effective soil stabilizer, which  
 527 typically exhibit high primary production under liquid water conditions, and which leach N-  
 528 rich metabolites into the soil (Williams et al. 2017). The low precipitation in the Ny-Ålesund  
 529 region and lack of melt water during the summer season thus explains the higher abundance  
 530 of chlorolichens over cyanolichens. Similar lichen patterns were described for the west coast  
 531 of Greenland (Heindel et al. 2019). Typical chlorolichen taxa associated with biocrusts are  
 532 *Cetraria muricata*, *Cladonia pyxidata*, *Lepraria cf. neglecta*, and *Psora rubiformis*, which are  
 533 part of the about 600 known lichen species of the flora of Svalbard (Elvebakk and Hertel  
 534 1997). However, these lichen numbers are based on total numbers and not only those  
 535 associated with biocrusts.

536 Our most intriguing observation was that intermediate and wet plots in KH and the wet plots  
 537 in OS were dominated by biocrusts and mosses. This would assign these plots to a wetland  
 538 association. However, the Svalbard wetland vegetation is poorly studied and biocrusts have  
 539 not been included into its flora characterization (Elvebakk 1994; Walker et al. 2009).  
 540 Therefore, we propose the integration of biocrusts into vegetation associations in the form of  
 541 a new syntaxon. The already used terms ‘lichen’, ‘bryophyte’, and ‘cryptophyte’ (Weber et al.  
 542 2000) should be modified to ‘lichen’, ‘moss’ and ‘biocrust’ to define the vegetation in a more



543 realistic and consistent way, as they are clearly too abundant in the Arctic tundra to be  
544 neglected.

545 Although biocrusts were the dominating vegetation type in the wet plots, only dark biocrusts  
546 were detected. The missing light biocrusts are defined as an early developmental stage with  
547 low biodiversity (Pushkareva et al. 2016). The dominant phototrophic organisms in light  
548 biocrusts are filamentous green algae and cyanobacteria. These communities stabilize the soil  
549 beneath and thereby facilitate the colonialization by other non-filamentous microalgae and  
550 cyanobacteria (Weber et al. 2016). Dark biocrusts are at a later successional stage and possess  
551 a higher biodiversity (Weber et al. 2016). The substrate stability and properties due to  
552 colonialization by biocrusts is fundamental for the even later succession of mosses, lichens  
553 and ultimately vascular plants (Breen and Levesque 2006; Langhans et al. 2009). This has  
554 been shown in a vegetation study of a glacier foreland on Svalbard which ran for over 40  
555 years and showed that biocrusts were eventually replaced by vascular plants (Hodkinson et al.  
556 2003). Dark biocrusts were common in both field sites (14 % in OS, 42 % in KH). This  
557 indicates well-developed biocrusts in the studied sites (Pushkareva et al. 2016) which in turn  
558 reflects low disturbance by mechanical processes like cryoturbation (Pushkareva et al. 2016;  
559 Yoshitake et al. 2010).

560

## 561 **Conclusions**

562 Our findings highlight the importance of cryptogamic covers in Arctic tundra, which have  
563 been largely neglected in earlier vegetation surveys. We suggest that besides lichens and  
564 mosses, in particular biocrusts should be considered as a new additional syntaxon in future  
565 Arctic vegetation mapping. In the face of global change particularly at high latitudes we  
566 further suggest that long-term studies of the dynamics in the vegetation composition are  
567 necessary to better understand the crucial role cryptogamic covers and in particular biocrusts  
568 play in the 'greening of the Arctic'. In addition, soil moisture could be identified as an  
569 ecological key factor controlling vegetation type and coverage.

570

## 571 **Author Contributions**

572 RK, VH, AF, BT, DV, CS, BF, EV, AQ, MS, KG and UK all contributed to the study design  
573 as well as sample and data collection during the joint summer expedition 2017 in Ny-  
574 Ålesund. MA, CB, MP, AF and LDM analyzed samples for specific parameters. RK, VH, KG  
575 and UK undertook all statistical analysis. RK, VH and UK wrote the first version of the  
576 manuscript with contributions from all coauthors.

577

**578 Funding**

579 This study was funded through the 2015-2016 BiodivERsA COFUND call for research  
580 proposals, with the national funders of Belgium (BELSPO BR/175/A1/CLIMARCTIC-BE),  
581 Germany (DFG KA899/33-1), Norway (The Research Council of Norway 270252/E50),  
582 Spain (MINECO, PCIN2016-001, CTM2016-79741) and Switzerland (SNSF  
583 31BD30\_172464).

584

**585 Conflict of Interest Statement**

586 The authors declare that the research was conducted in the absence of any commercial or  
587 financial, as well as non-financial relationships that could be construed as a potential conflict  
588 of interest.

589

**590 Acknowledgements**

591 The authors are grateful to the staff at the AWIPEW station, Ny-Ålesund for excellent  
592 technical and logistic support during the summer campaign 2017.

593

**594 Electronic Supplementary Material**

595 The Electronic Supplementary Material for this article can be found as Online Resource 1 and  
596 2:

597

**598 References**

599 Berg G, Smalla K (2009) Plant species and soil type cooperatively shape the structure and  
600 function of microbial communities in the rhizosphere. *FEMS Microb Ecol* 68:1-13.

601 doi:10.1111/j.1574-6941.2009.00654.x

602 Birkeland S, Skjetne IEB, Brysting AK, Elven R, Alsos IG (2017) Living on the edge:  
603 conservation genetics of seven thermophilous plant species in a high Arctic archipelago.

604 *AoB PLANTS* 9:plx001. doi:10.1093/aobpla/plx001

605 Breen K, Levesque E (2006) Proglacial succession of biological soil crusts and vascular  
606 plants: biotic interactions in the High Arctic. *Can J Bot* 84:1714–1731.

607 doi.org/10.1139/b06-131

608 Castillo-Monroy A, Maestre F, Delgado-Baquerizo M, Gallardo A (2010) Biological soil  
609 crusts modulate nitrogen availability in semi-arid ecosystems: insights from a

- 610 Mediterranean grassland. *Plant Soil* 333:21–34. doi.org/10.1007/s11104-009-0276-7,  
611 2010.
- 612 Chapin III FS, Shaver GR (1981) Changes in soil properties and vegetation following  
613 disturbance of Alaskan Arctic tundra. *J Appl Ecol* 18:605-617. doi: 10.2307/2402420
- 614 Chu H, Grogan P (2010) Soil microbial biomass, nutrient availability and nitrogen  
615 mineralization potential among vegetation-types in a low Arctic tundra landscape. *Plant*  
616 *Soil* 329:411-420. doi.org/10.1007/s11104-009-0167-y
- 617 Cleveland C, Liptzin D (2007) C:N:P stoichiometry in soil: is there a “Redfield ratio” for the  
618 microbial biomass? *Biogeochem* 85: 235-252. doi 10.1007/s10533-007-9132-0Döbelin  
619 N, Kleeberg R (2015) Profex: a graphical user interface for the Rietveld refinement  
620 program BGMN. *J Appl Crystal* 48:1573-1580. doi.org/10.1107/S1600576715014685
- 621 Dusa A (2017) Venn: Draw Venn Diagrams.
- 622 Edwards KA, Jefferies RL (2013) Inter-annual and seasonal dynamics of soil microbial  
623 biomass and nutrients in wet and dry low-Arctic sedge meadows. *Soil Biol Biochem*  
624 57:83-90. doi.org/10.1016/j.soilbio.2012.07.018
- 625 Elbert W, Weber B, Burrows S, Steinkamp J, Büdel B, Andreae MO, Pöschl U (2012)  
626 Contribution of cryptogamic covers to the global cycles of carbon and nitrogen. *Nat*  
627 *Geosci* 5:459-462. doi.org/10.1038/ngeo1486, 2012.
- 628 Elvebakk A (1994) A survey of plant associations and alliances from Svalbard. *J Veg Sci*  
629 5:791-802. doi.org/10.2307/3236194
- 630 Elvebakk A, Hertel H (1997) A catalogue of Svalbard lichens. In Elvebakk A, Prestrud P  
631 (eds) A catalogue of Svalbard plants, fungi, algae, and cyanobacteria, Norsk Polarinstitutt  
632 Skrifter, Oslo, Norway, pp 271-411.
- 633 Elvebakk, A. (1999). Bioclimatic delimitation and subdivision of the Arctic. In: Nordal I,  
634 Razzhivin VY (eds) The species concept in the High North — A Panarctic Flora  
635 Initiative. The Norwegian Academy of Science and Letters, Oslo, Norway, pp 81–112.
- 636 Gee GW, Bauder JW (1986) Particle-size analysis. In: Klute A (ed) *Methods of soils analysis*,  
637 part 1: Physical and mineralogical methods. *Agronomy Monograph*, 9, 2nd ed. American  
638 Society of Agronomy/Soil Science Society of America, Madison, WI, USA, pp 383-411.
- 639 Giesler R, Esberg C, Lagerström A, Graae BJ (2012) Phosphorus availability and microbial  
640 respiration across different tundra vegetation types. *Biogeochem* 108:429–445.  
641 doi.org/10.1007/s10533-011-9609-8

- 642 Gray ND, McCann CM, Christgen B, Ahammad SZ, Roberts J, Graham DW (2014) Soil  
643 geochemistry confines microbial abundances across an Arctic landscape; implications for  
644 net carbon exchange with the atmosphere. *Biogeochem* 120:307–317.  
645 doi.org/10.1007/s10533-014-9997-7
- 646 Harden JW, Koven CD, Ping CL, Hugelius G, McGuire AD, Camill P, Jorgenson T, Kuhry P,  
647 Michaelson GJ, O'Donnell JA, Schuur EAG, Tarnocai C, Johnson K, Grosse G (2012)  
648 Field information links permafrost carbon to physical vulnerabilities of thawing. *Geophys*  
649 *Res Lett* 39:L15704. doi:10.1029/2012GL051958
- 650 Hedley MJ, Stewart JWB, Chauhan BS (1982) Changes in inorganic und organic soil-  
651 phosphorus fractions induced by cultivation practices and by laboratory incubations. *Soil*  
652 *Sci Soc Am J* 46:970-976. doi:10.2136/sssaj1982.03615995004600050017x
- 653 Heindel RC, Governali FC, Spickard AM, Virginia RA (2019) The role of biological soil  
654 crusts in nitrogen cycling and soil stabilization in Kangerlussuaq, West Greenland.  
655 *Ecosystems* 22:243-256. doi.org/10.1007/s10021-018-0267-8
- 656 Hervé M (2018) RVAideMemoire: Testing and Plotting Procedures for Biostatistics. R  
657 package version 0.9-69
- 658 Hobara S, Mccalley C, Koba K, Giblin AE, Weiss MS, Gettel GM, Shaver GR (2006)  
659 Nitrogen fixation in surface soils and vegetation in an Arctic tundra watershed: a key  
660 source of atmospheric nitrogen. *Arct Antarct Alp Res* 38:363–372.  
661 doi.org/10.1657/1523-0430(2006)38
- 662 Hodkinson ID, Coulson SJ, Webb NR (2003) Community assembly along proglacial  
663 chronosequences in the high Arctic: vegetation and soil development in north-west  
664 Svalbard. *J Ecol* 91:651-663. doi.org/10.1046/j.1365-2745.2003.00786.x
- 665 Johansen BE, Tømmervik H, Karlsen SR (2012) Vegetation mapping of Svalbard utilizing  
666 Landsat TM/ETM+ data. *Polar Rec* 48:47-63. doi.org/10.1017/S0032247411000647
- 667 Johansen BE, Tømmervik H (2014) The relationship between phytomass, NDVI and  
668 vegetation communities on Svalbard. *Int J Appl Earth Obs* 27:20-30.  
669 doi.org/10.1016/j.jag.2013.07.001
- 670 Keller K, Blum JD, Kling GW (2007) Geochemistry of soils and streams on surfaces of  
671 varying ages in Arctic Alaska. *Arct Antarct Alp Res* 39:84-98.  
672 doi.org/10.1657/1523-0430
- 673 Lange OL, Kilian E, Ziegler H (1986) Water vapor uptake and photosynthesis of lichens:  
674 performance differences in species with green and blue-green algae as phycobionts.  
675 *Oecologica* 71:104-110. doi.org/10.1007/BF00377327

- 676 Langhans TM, Storm C, Schwabe A (2009) Community assembly of biological soil crusts of  
677 different successional stages in a temperate sand ecosystem, as assessed by direct  
678 determination and enrichment techniques. *Microb Ecol* 58:394-407.  
679 doi.org/10.1007/s00248-009-9532-x
- 680 Levy E, Madden E (1933). The point method of pasture analysis. *New Zeal J Agr Res*  
681 46:267–279.
- 682 Loisel J, Yu Z, Beilman DW, Camill P, Alm J, Amesbury MJ, Anderson D, Andersson S,  
683 Bochicchio C, Barber K, Belyea LR, Bunbury J, Chambers FM, Charman DJ, De  
684 Vleeschouwer F, Fialkiewicz-Kozie B, Finkelstein S, Galka M, Garneau M, Hammarlund  
685 D, Hinchcliffe W, Holmquist J, Hughes P, Jones MC, Klein ES, Kokfelt U, Korhola A,  
686 Kuhry P, Lamarre A, Lamentowicz M, Large D, Lavoie M, MacDonald G, Magnan G,  
687 Makila M, Mallon G, Mathijssen P, Mauquoy D, McCarroll J, Moore TR, Nichols J,  
688 O'Reilly B, Oksanen P, Packalen M, Peteet D, Richard PJ, Robinson S, Ronkainen T,  
689 Rundgren M, Sannel ABK, Tarnocai C, Thom T, Tuittila ES, Turetsky M, Valiranta M,  
690 van der Linden M, van Geel B, van Bellen S, Vitt D, Zhao Y, Zhou W (2014) A database  
691 and synthesis of northern peatland soil properties and Holocene carbon and nitrogen  
692 accumulation. *The Holocene* 24:1028–1042. doi.org/10.1177/0959683614538073
- 693 Liu XY, Koba K, Koyama LA, Hobbie SE, Weiss MS, Inagaki Y, Shaver GR, Giblin AE,  
694 Hobara S, Nadelhoffer KJ, Sommerkorn M, Rastetter EB, Kling GW, Laundre JA, Yano  
695 Y, Makabe A, Yano M, Liu CQ (2018) Nitrate is an important nitrogen source for Arctic  
696 tundra plants. *Proc Natl Acad Sci USA* 115:3398-3403.  
697 doi.org/10.1073/pnas.1715382115
- 698 Mercado-Díaz JA, Gould WA, González G (2014). Soil  
699 nutrients, landscape age, and *Sphagno-Eriophoretum vaginati* plant communities in  
700 Arctic moist-acidic tundra landscapes. *OJSS* 4:375-387.  
701 doi.org/10.4236/ojss.2014.411038
- 702 Murphy J, Riley JP (1962) A modified single solution method for the determination of  
703 phosphate in natural waters. *Anal Chim Acta* 27:31-36.  
704 doi.org/10.1016/S0003-2670(00)88444-5
- 705 Øvstedal D, Tønsberg T, Elvebakk A (2009) The lichen flora of Svalbard. *Sommerfeltia* 33:3-  
706 393. doi.org/10.2478/v10208-011-0013-5
- 707 Oksanen J, Blanchet FG, Friendly M, Kindt R, Legendre P, McGlenn D, Minchin PR, O'Hara  
708 RB, Simpson GL, Solymos P, Stevens MHH, Szoecs E, Wagner H (2018) Package  
'vegan'. *Community Ecology Package*. Version 2.5-3.

- 709 Pushkareva E, Pessi IS, Willmotte A, Elster J (2015) Cyanobacterial community composition  
710 in Arctic soil crusts at different stages of development. *FEMS Microbiol Ecol* 91:fiv143.  
711 doi: 10.1093/femsec/fiv143
- 712 Pushkareva E, Johansen JR, Elster J (2016) A review of the ecology, ecophysiology and  
713 biodiversity of microalgae in Arctic soil crusts. *Polar Biol* 39:2227–2240.  
714 doi.org/10.1007/s00300-016-1902-5
- 715 Reich PB, Hobbie SE, Lee T, Ellsworth DS, West JB, Tilman D, Knops JMH, Naeem S, Trost  
716 J (2006) Nitrogen limitation constrains sustainability of ecosystem response to CO<sub>2</sub>.  
717 *Nature* 440:922–925. doi.org/10.1038/nature04486
- 718 Rønning OI (1996) *The Flora of Svalbard*, 1st ed. Norwegian Polar Institute, Tromsø, Norway
- 719 Schimel JP, Bennett J (2004) Nitrogen mineralization: challenges of a changing paradigm.  
720 *Ecology* 85:591–602. doi.org/10.1890/03-8002
- 721 Screen JA, Simmonds I (2010) The central role of diminishing sea ice in recent Arctic  
722 temperature amplification. *Nature* 464:1334–1337. doi:10.1038/nature09051
- 723 Shaver GR, Chapin III FS (1991) Production:biomass relationships and element cycling in  
724 contrasting Arctic vegetation types. *Ecol Monogr* 61:1–31. doi.org/10.2307/1942997
- 725 Van Cleve K, Alexander V (1981) Nitrogen cycling in tundra and boreal ecosystems. In:  
726 Clarke FE, Rosswall T (eds) *Terrestrial Nitrogen Cycles*, *Ecol Bull* 33:375–404.
- 727 Vimal SR, Singh JS, Arora NK, Singh S (2017) Soil-plant-microbe interactions in stressed  
728 agriculture management: a review. *Pedosphere* 27:177–192.  
729 doi:10.1016/S1002-0160(17)60309-6
- 730 Walker DA (2000) Hierarchical subdivision of Arctic tundra based on vegetation response to  
731 climate, parent material and topography. *Glob Change Biol* 6:19–34.  
732 doi.org/10.1046/j.1365-2486.2000.06010.x
- 733 Walker DA, Raynolds MK, Daniëls FJ, Einarsson E, Elvebakk A, Gould WA, Katenin AE,  
734 Kholod SS, Markon CJ, Melnikov ES, Moskalenko NG, Talbot SS, Yurtsev BA, and the  
735 other members of the CAVM Team (2009) The circumpolar Arctic vegetation map. *J*  
736 *Veg Sci* 16:267–282. doi.org/10.1111/j.1654-1103.2005.tb02365.x
- 737 Weber B, Büdel B, Belnap J (2016) *Biological soil crusts: an organizing principle in drylands*,  
738 1st ed. Springer, Berlin, Heidelberg, Germany.
- 739 Weber HE, Moravec J, Theurillat JP (2000) *International Code of Phytosociological*  
740 *Nomenclature*, 3rd edition. *J Veg Sci* 11:739–768. doi.org/10.2307/3236580
- 741 Wild B, Schnecker J, Bárta J, Capek P, Guggenberger G, Hofhansl F, Kaiser C, Lashchinsky  
742 N, Mikutta R, Mooshammer M, Santracková H, Shibistova O, Urich T, Zimov S, Richter

Formatiert: Englisch (USA)

- 743 A (2013) Nitrogen dynamics in turbid cryosols from Siberia and Greenland. *Soil Biol*  
 744 *Biochem* 67:85–93. doi.org/10.1016/j.soilbio.2013.08.004
- 745 Williams L, Borchhardt N, Colesie C, Baum C, Komsic-Buchmann K, Rippin M, Becker B,  
 746 Karsten U, Büdel B (2017) Biological soil crusts of Arctic Svalbard and of Livingston  
 747 Island, Antarctica. *Polar Biol* 40:399-411. doi.org/10.1007/s00300-016-1967-1
- 748 Yoshitake S, Uchida M, Koizumi H, Kanda H, Nakatsubo T (2010) Production of biological  
 749 soil crusts in the early stage of primary succession on a High Arctic glacier foreland. *New*  
 750 *Phytol* 186:451-460. doi.org/10.1111/j.1469-8137.2010.03180.x
- 751 Zamin TJ, Grogan P (2012) Birch shrub growth in the low Arctic: the relative importance of  
 752 experimental warming, enhanced nutrient availability, snow depth and caribou exclusion.  
 753 *Environ Res Lett* 7:034027. doi.org/10.1088/1748-9326/7/3/034027
- 754 Zhang X, Friedl MA, Schaaf CB, Strahler AH (2004) Climate controls on vegetation  
 755 phenological patterns in northern mid- and high latitudes inferred from MODIS data.  
 756 *Glob Change Biol* 10:1133-1145. doi.org/10.1111/j.1529-8817.2003.00784.x

757

758

## 759 **Figure Legends**

760 **Fig. 1 A-D.** Maps of Svalbard (A, B) and the two study sites Knudsenheia (C) and Ossian-  
 761 Sarsfjellet (D). The symbols on the magnified maps of Knudsenheia (C) K.1.1. to K.1.3,  
 762 K.2.1. to K.2.3 and K.3.1 to K.3.3 indicate the dry, intermediate and wet plots, respectively.  
 763 Accordingly, the symbols on the magnified maps of Ossian-Sarsfjellet (D) O.1.1. to O.1.3,  
 764 O.2.1. to O.2.3 and O.3.1 to O.3.3 label the dry, intermediate and wet plots, respectively. All  
 765 plot details are summarized in Table 1.

766

767 **Fig. 2 A-B.** Photographs of the two study sites Knudsenheia (A) and Ossian-Sarsfjellet (B).

768

769 **Fig. 3.** Dendrogram of each replicate from dry (d), intermediate (i) and wet (w) plots from  
 770 Knudsenheia (KH) and Ossian-Sarsfjellet (OS) based on environmental soil data. The  
 771 dendrogram was drawn after a cluster analysis using total C/N, labile phosphorus,  
 772 ammonium, nitrate, pH, moisture, organic matter content and texture of the upper soil  
 773 samples (Table 1). ediate; KH w: Knudsenheia wet; OS d: Ossian-Sarsfjellet dry; OS i:  
 774 Ossian-Sarsfjellet intermediate; OS w: Ossian-Sarsfjellet wet.

775

776 **Fig 4.** Summary of the vegetation survey. Percentage of area covered by the different  
 777 functional groups as determined with the point intercept method on (A) the whole transect and  
 778 (B) in the subsites along the soil moisture gradient (Data see Table 1). KH: Knudsenheia; OS:  
 779 Ossian-Sarsfjellet; KH d: Knudsenheia dry; KH i: Knudsenheia intermediate; KH w:  
 780 Knudsenheia wet; OS d: Ossian-Sarsfjellet dry; OS i: Ossian-Sarsfjellet intermediate; OS w:  
 781 Ossian-Sarsfjellet wet.

782

783 **Fig 5.** Non-metric multidimensional scaling (nMDS) plot visualizing the similarity and  
 784 dissimilarity of the ground coverage in Knudsenheia (KH) and Ossian-Sarsfjellet (OS) in the  
 785 different plots (d-dry, i-intermediate, w-wet). Ellipses correspond to 95% confidence interval.  
 786 The black arrows indicate the influence direction of the only three significantly correlated soil  
 787 parameters: ammonium, pH and moisture. Stress=0.15

788

789 **Fig. 6.** Number of vascular plant species (A) in the two field sites and (B) in the subsites.  
 790 Venn diagram showing the total number of vascular plant species in each field site,  
 791 Knudsenheia (KH, yellow-brown) and Ossian-Sarsfjellet (OS, turkis-green), as well as the  
 792 number of species that are present in both sites (intersection). (B) Number of different  
 793 vascular plant species per plot (dry, intermediate, wet) in Knudsenheia (KH) and Ossian-  
 794 Sarsfjellet (OS).

795

796 **Fig. 7.** Scheme of both Arctic vegetation toposequences along the soil moisture gradients in  
 797 Knudsenheia (KH) and Ossian-Sarsfjellet (OS) (d: dry; i: intermediate; w: wet). Only the  
 798 dominant vegetation is indicated. Adapted after Elvebakk (1994) and Walker (2000).

799

800

### 801 **Figure Legends for Online Resource 1**

802

803 **Fig. 8.1-8.18.** Digital photographs of the 18 permanent sampling plots showing the dominant  
 804 vegetation. K: Knudsenheia; O:Ossian-Sarsfjellet; K1.1-K1.3: dry plots; K2.1-K2.3:  
 805 intermediate plots; K3.1-K3.3: wet plots; O1.1-O1.3: dry plots; O2.1-O2.3: intermediate  
 806 plots; O3.1-O3.3: wet plots.

807

808 **Fig. 9.** Non-metric multidimensional scaling (nMDS) plot visualizes the similarity and  
 809 dissimilarity of the vascular plant diversity in Knudsenheia (KH) and Ossian-Sarsfjellet (OS)



810 at different plots (d-dry, i-intermediate, w-wet). The black arrows indicate the influence  
811 direction of the only three significantly correlated soil parameters: ammonium, pH, sand  
812 content and moisture. Ellipses correspond to 95% confidence interval. Stress=0.17  
813



**TABLE 2** Mineralogical soil properties. For sample details see Table 1. Values are given as percentage of dry weight with  $3\sigma$  error (3 replicates). n.d.: not detected.

Study site	Subsite	Dolomite Ankerite	Calcite	Muscovite	Biotite	Chlorite	Na-Plagioclase	K-Feldspar	Kaolinite	Quartz	Hematite	Sillimanite
Knudsenheia	Dry	8.0±0.9	0.8± 0.1	2.9± 0.2	n.d. 0.2	5.4± 1.2	5.0± 0.3	1.6± 0.1	1.5± 0.2	73.8± 1.0	1.0± 0.1	n.d.
		Intermediate	31.7±1.5	2.4± 0.2	3.8± 0.2	n.d. 0.2	2.4± 0.4	6.9± 0.5	4.0± 0.3	0.6± 0.2	47.6± 0.5	0.6± 0.1
	Wet		25.7±0.9	0.6± 0.1	2.4± 0.2	<0.6 0.2	2.2± 0.6	9.2± 0.5	3.7± 0.2	n.d. 0.2	55.6± 0.5	n.d.
		Dry	4.8±0.6	6.8± 0.2	6.2± 0.3	3.0± 0.3	3.7± 0.6	13.3± 0.6	5.4± 0.5	n.d. 0.5	56.8± 0.7	n.d.
	Intermediate		10.8±0.6	5.5± 0.2	8.6± 0.3	4.0± 0.5	6.6± 0.6	11.1± 0.7	5.0± 0.4	0.9± 0.3	44.8± 0.6	0.4± 0.1
		Wet	22.6±0.5	12.5± 0.2	9.6± 0.3	5.2± 0.3	5.1± 0.5	7.1± 0.4	3.1± 0.3	1.2± 0.3	33.6± 0.5	n.d.

**TABLE 3** Plant taxa found at the study sites Knudsenheia (KH) and Ossian Sarsfjellet (OS) along a moisture gradient defined by three subsites dry (d), intermediate (i), and wet (w). The taxonomic level of clade and family for each species is given.

Species	Family	Clade	Growth form	Occurrence						
				KHd	KHi	KHw	OSd	OSi	OSw	
<i>Equisetum variegatum</i>	Equisetaceae	Equisetopsida	Forb		x					x
<i>Carex</i> sp.	Cyperaceae	Monocotyledonae	graminoid		x	x				
<i>Festuca</i> sp.	Poaceae	Monocotyledonae	graminoid				x	x		
<i>Luzula confuse</i>	Juncaceae	Monocotyledonae	graminoid	x	x					
<i>Luzula nivalis</i>	Juncaceae	Monocotyledonae	graminoid	x						
<i>Poa</i> sp.	Poaceae	Monocotyledonae	graminoid	x			x	x		x
<i>Arenaria</i> sp.	Caryophyllaceae	Dicotyledonae	Forb		x	x				
<i>Bistorta vivipara</i>	Polygonaceae	Dicotyledonae	Forb		x	x				
<i>Cardamine pratensis</i> ssp. <i>angustifolia</i>	Brassicaceae	Dicotyledonae	Forb		x	x				x
<i>Cassiope tetragona</i>	Ericaceae	Dicotyledonae	Dwarf shrub					x		x
<i>Cerastium arcticum</i>	Caryophyllaceae	Dicotyledonae	Forb	x						
<i>Draba alpina</i>	Brassicaceae	Dicotyledonae	Forb	x						
<i>Dryas octopetala</i>	Rosaceae	Dicotyledonae	Dwarf shrub				x	x		x
<i>Minuartia rubella</i>	Caryophyllaceae	Dicotyledonae	Forb	x						
<i>Minuartia</i> sp.	Caryophyllaceae	Dicotyledonae	Forb		x	x				
<i>Oxyria digyna</i>	Polygonaceae	Dicotyledonae	Forb	x	x					
<i>Papaver dahlianum</i>	Papaveraceae	Dicotyledonae	Forb	x						
<i>Pedicularis hirsuta</i>	Scrophulariaceae	Dicotyledonae	Forb	x						
<i>Sagina nivalis</i>	Caryophyllaceae	Dicotyledonae	Forb		x	x				x
<i>Salix polaris</i>	Salicaceae	Dicotyledonae	Dwarf shrub	x	x	x		x		x

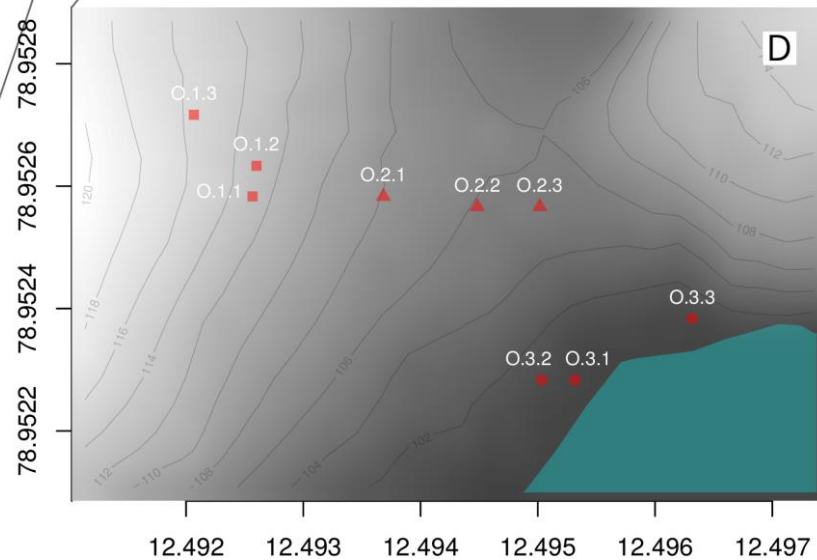
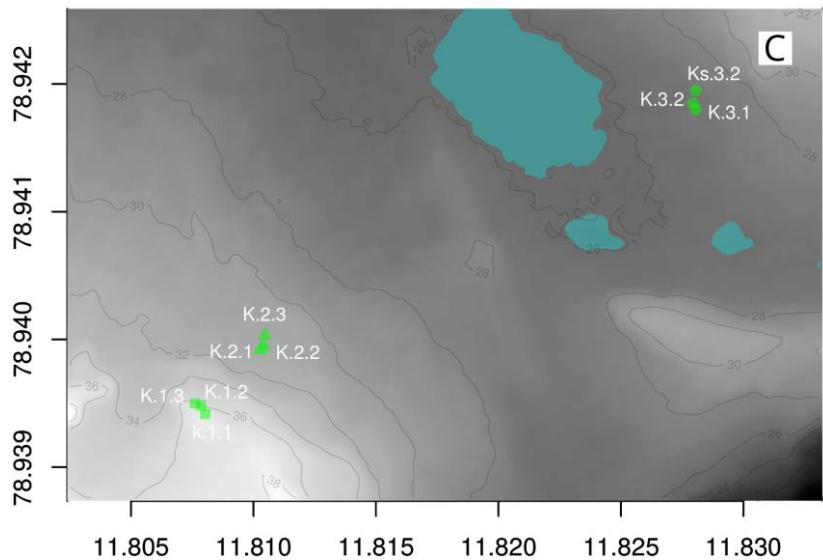
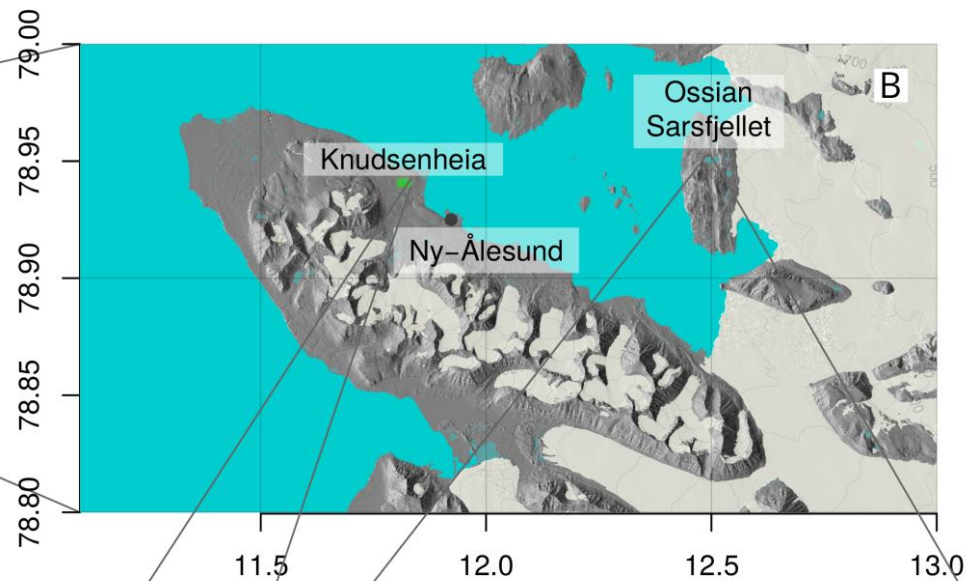
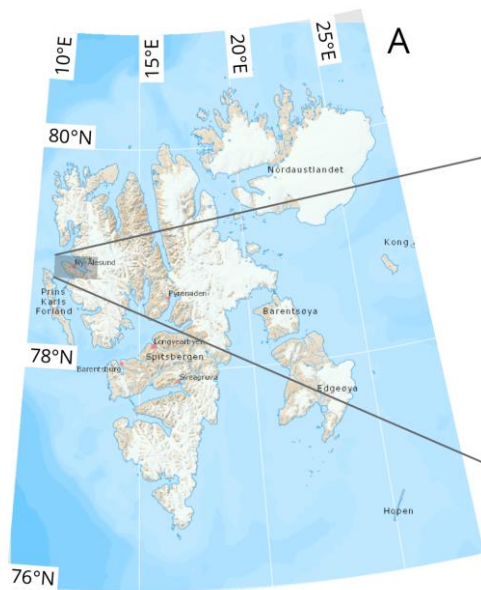
<i>Saxifraga cernua</i>	Saxifragaceae	Dicotyledonae	Forb	x	x	x			
<i>Saxifraga cespitosa</i>	Saxifragaceae	Dicotyledonae	Forb	x	x	x			
<i>Saxifraga oppositifolia</i>	Saxifragaceae	Dicotyledonae	Forb	x	x		x	x	x
<i>Silene acaulis</i>	Caryophyllaceae	Dicotyledonae	Forb				x	x	

---

## Supplementary data

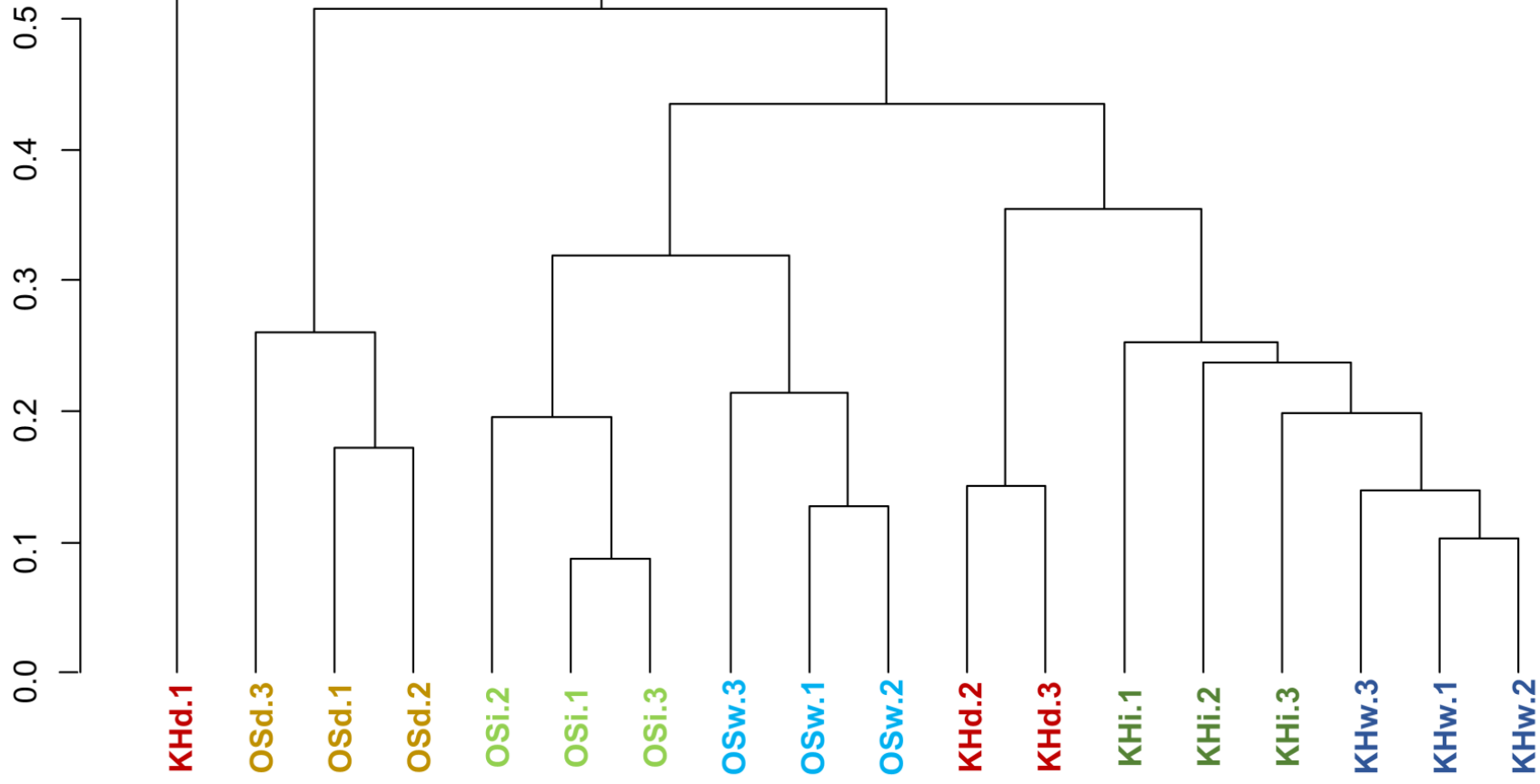
**TABLE 4** Summary of the vegetation surveys on the individual plots. Percentage of area covered by the different functional vegetation groups as determined with the point intercept method on all plots along the soil moisture gradient (dry, intermediate, wet) at Knudsenheia and Ossian Sarsfjellet.

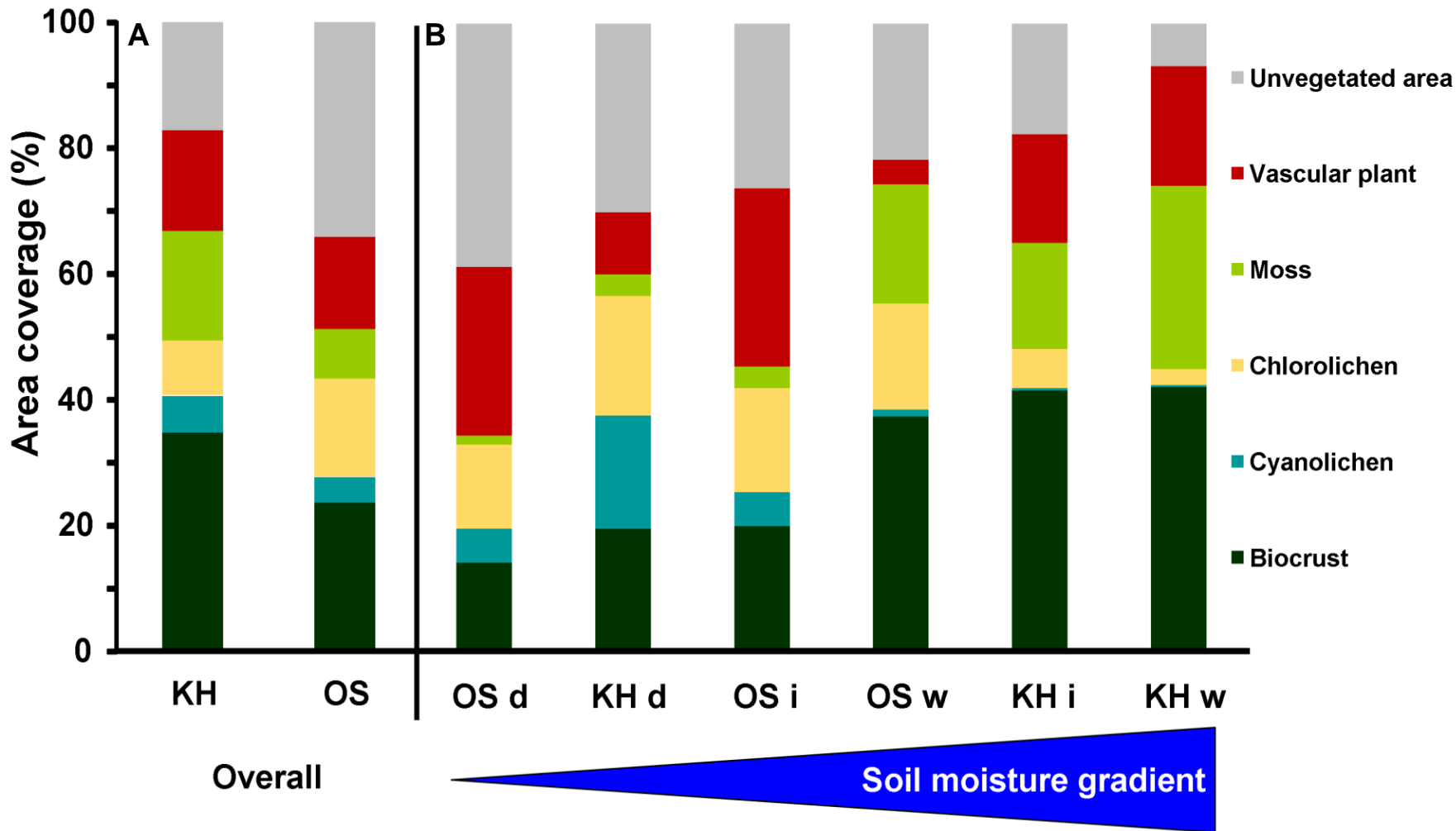
Study site	Subsite	Plot	Vegetation ground cover (%)				
			Higher plants	Mosses	Lichens	Dark BSCs	Unvegetated
Knudsenheia	Dry	K1.1	25	0	50	10	15
		K1.2	25	1	62	2	10
		K1.3	15	1	66	3	15
	Intermediate	K2.1	15	8	5	57	15
		K2.2	40	20	4	30	6
		K2.3	8	5	15	50	22
	Wet	K3.1	7	8	16	65	4
		K3.2	4	2	4	70	20
		K3.3	4	8	6	67	15
Ossian Sarsfjellet	Dry	O1.1	15	0	1	6	78
		O1.2	45	0	1	10	54
		O1.3	25	0	1	14	60
	Intermediate	O2.1	90	1	3	1	5
		O2.2	55	0	15	5	25
		O2.3	85	1	2	10	2
	Wet	O3.1	2	7	0	65	26
		O3.2	1	3	0	66	30
		O3.3	4	8	8	65	15

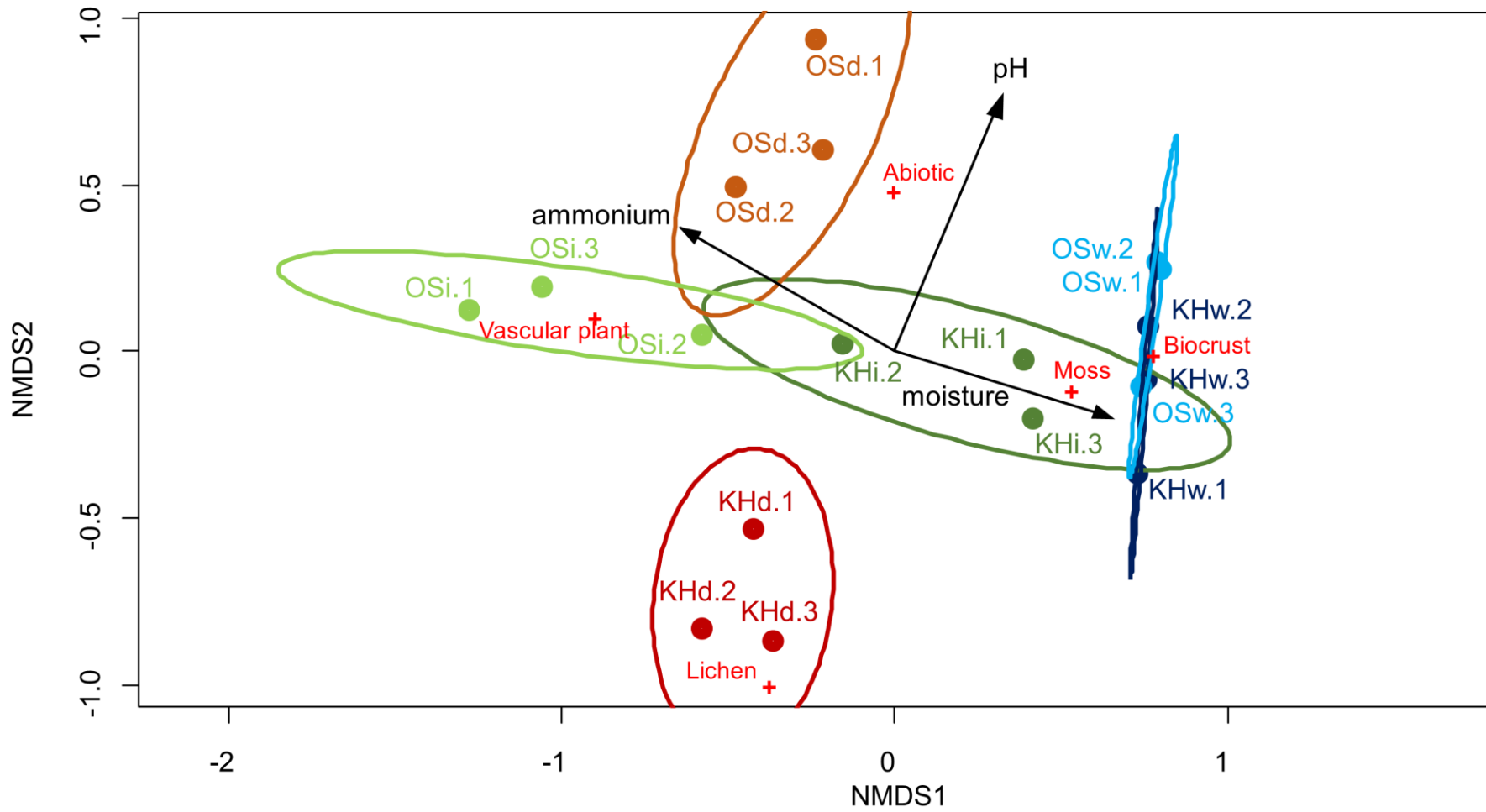


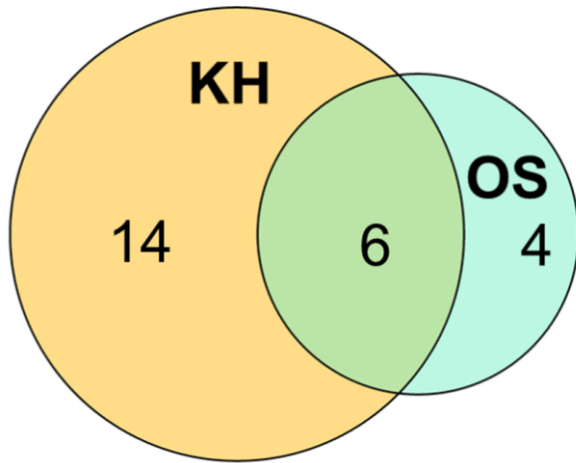
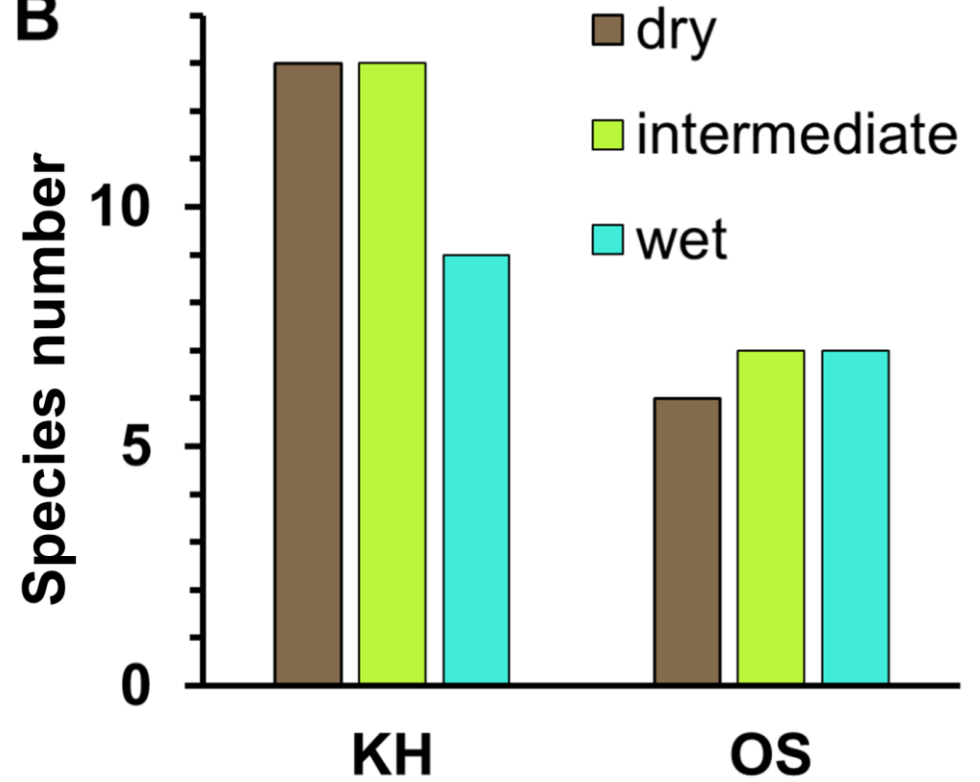




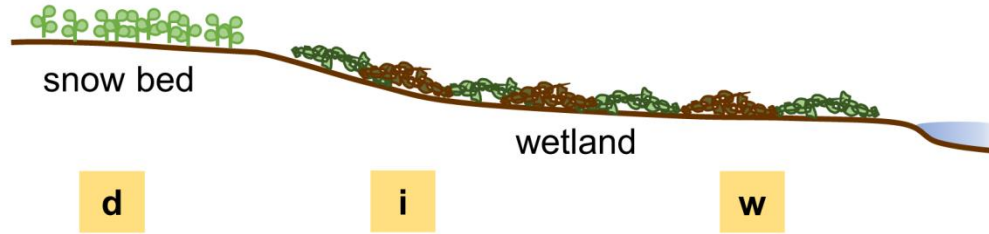











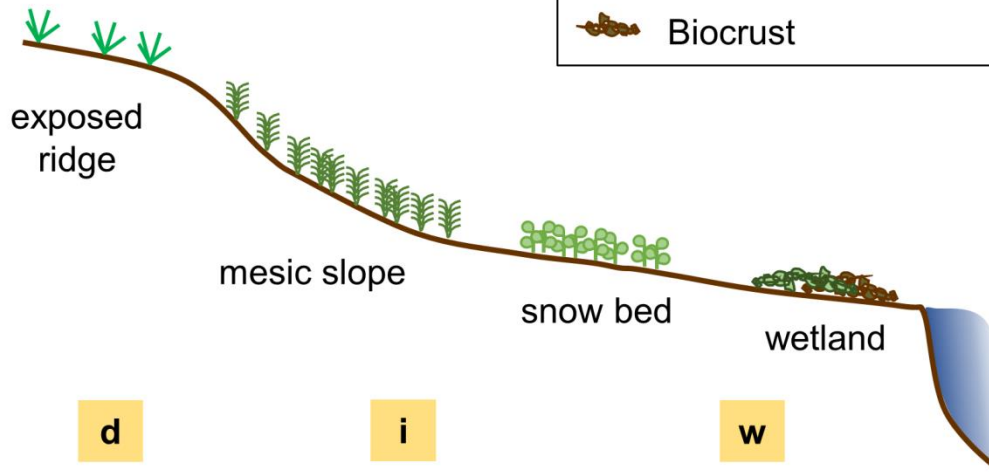
**A****B**

KH



-  Prostrate dwarf shrub
-  *Cassiope tetragona*
-  Herb
-  Moss
-  Biocrust

OS



## Supplementary Figures (8.1-8.18)

### Vascular plants on plots – Knudsenheia

\* dominant species

#### K 1.1



- *Salix polaris*
- *Saxifraga oppositifolia* \*
- *Luzula confusa*
- *Cerastium arcticum*

K 1.2



- *Salix polaris*
- *Saxifraga oppositifolia* \*
- *Draba alpina*
- *Minuartia* sp.
- *Sagina nivalis*
- *Cerastium arcticum*
- *Arenaria* sp.

K 1.3



- *Silene acaulis* \*
- *Saxifraga oppositifolia* \*
- *Salix polaris*
- *Minuartia* sp.
- *Luzula confusa*
- *Pedicularis hirsuta*
- *Arenaria* sp.



K 2.1



- *Salix polaris* \*
- *Luzula confusa*
- *Saxifraga oppositifolia*
- *Silene acaulis*
- *Minuartia* sp.
- *Bistorta vivipara*

K 2.2



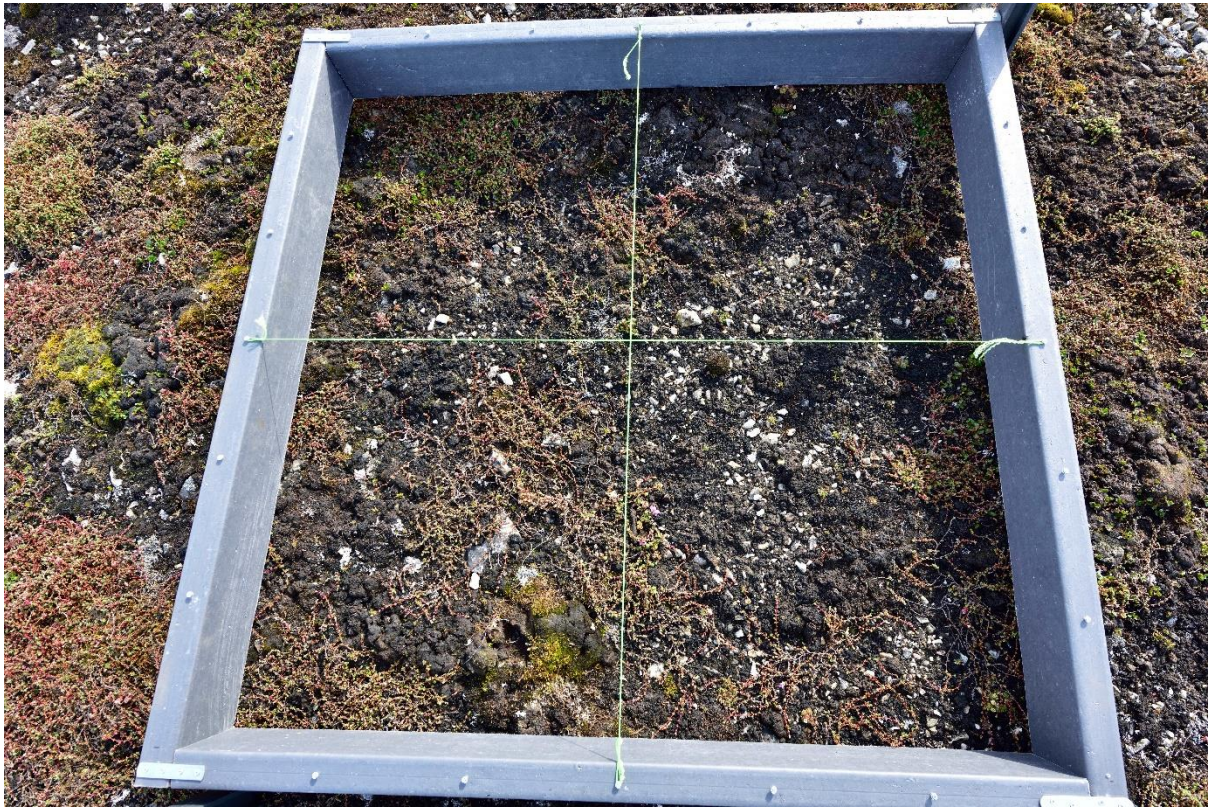
- *Saxifraga oppositifolia* \*
- *Salix polaris*
- *Luzula confusa*
- *Draba alpina*
- *Bistorta vivipara*
- *Oxyria digyna*
- *Saxifraga cespitosa*

K 2.3



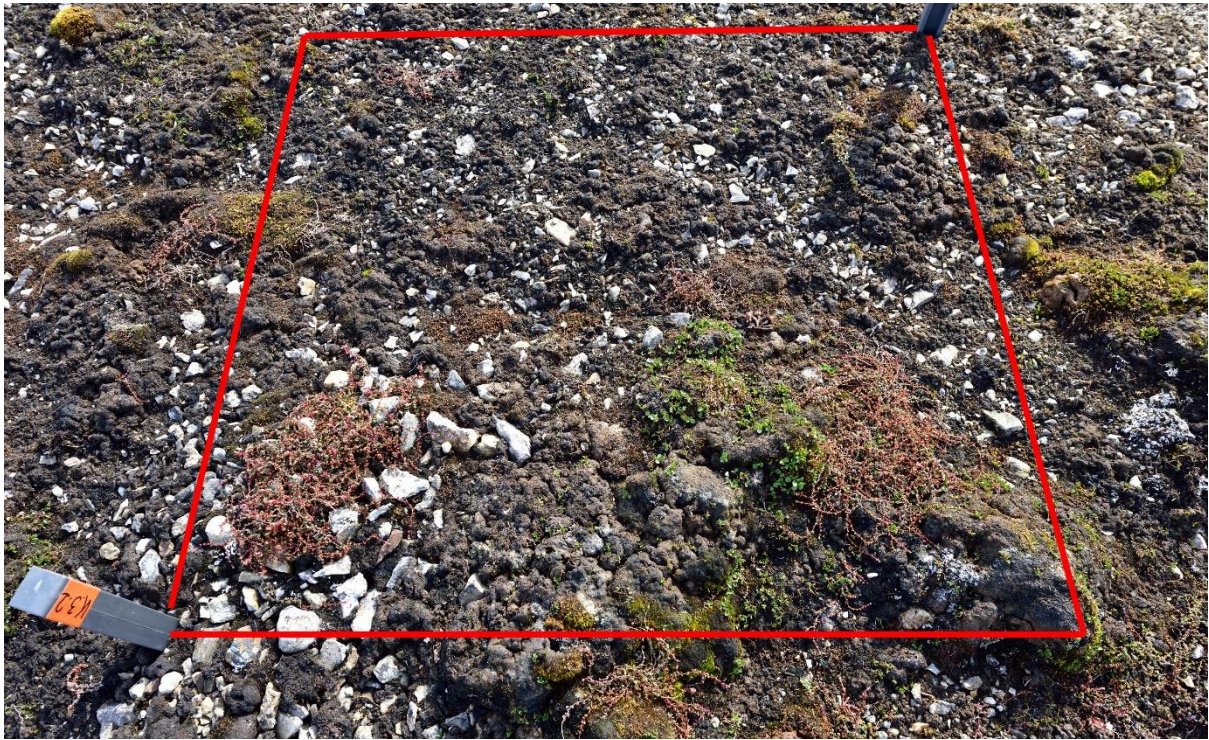
- *Saxifraga oppositifolia* \*
- *Salix polaris*
- *Draba alpina*
- *Bistorta vivipara*
- *Oxyria digyna*
- *Minuartia* sp.
- *Carex* sp.
- *Arenaria* sp.

K 3.1



- *Sagina nivalis*
- *Saxifraga oppositifolia* \*
- *Oxyria digyna*
- *Saxifraga cespitosa*
- *Salix polaris*
- *Luzula nivalis*

K 3.2



- *Saxifraga oppositifolia* \*
- *Salix polaris*
- *Sagina nivalis*
- *Saxifraga cespitosa*
- *Arenaria* sp.
- *Poa* sp.

K 3.3



- *Saxifraga oppositifolia* \*
- *Salix polaris*
- *Carex* sp.
- *Cardamine pratensis* ssp. *angustifolia*
- *Saxifraga cespitosa*
- *Sagina nivalis*
- *Arenaria* sp.

## Vascular plants on plots – Ossian Sarsfjellet

\* dominant species

### O 1.1



- *Dryas octopetala* \*
- *Luzula nivalis*
- *Saxifraga oppositifolia*

O 1.2



- *Dryas octopetala* \*
- *Luzula nivalis*
- *Saxifraga oppositifolia*
- *Silene acaulis*



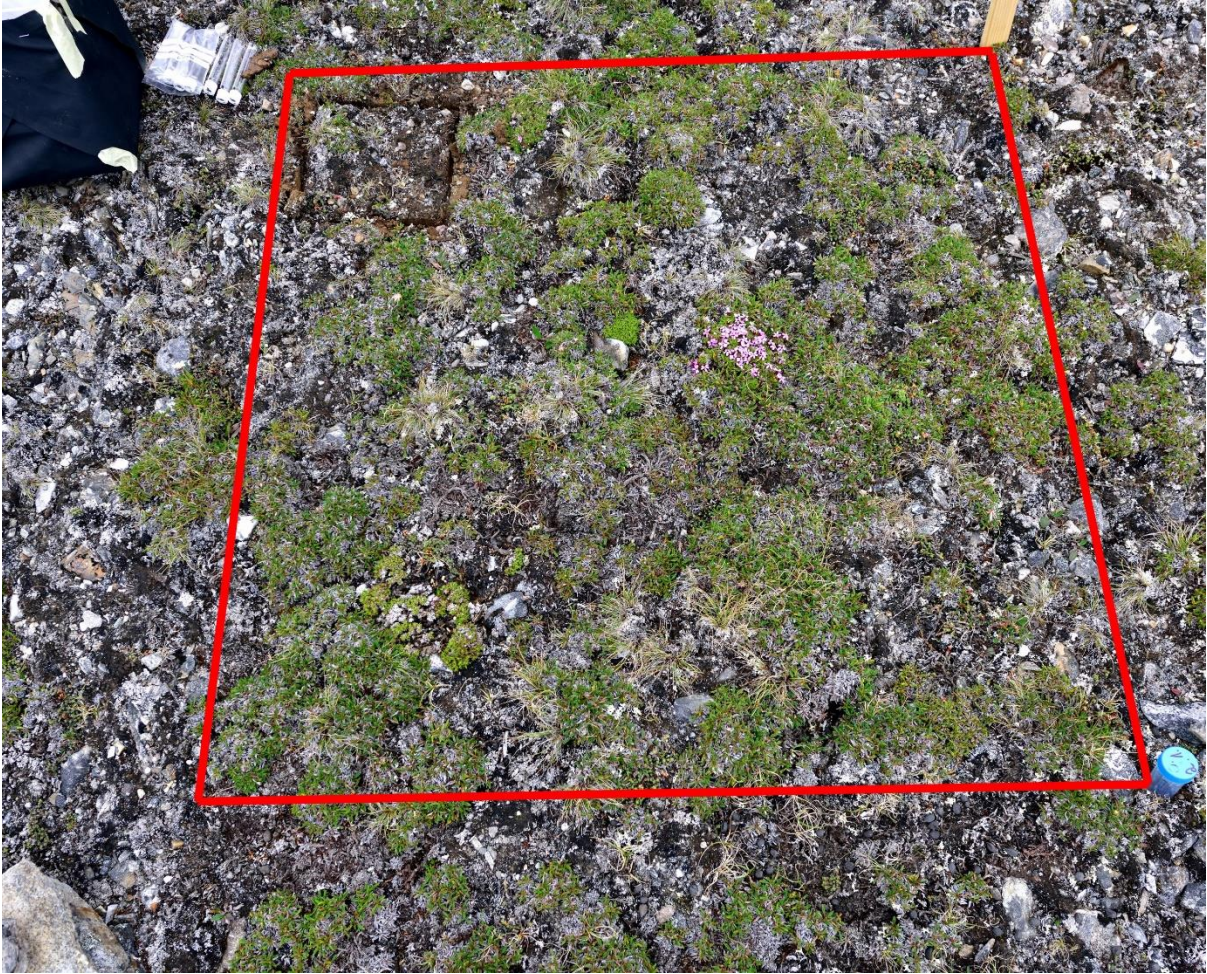
O 1.3



- *Dryas octopetala* \*
- *Luzula nivalis* \*
- *Silene acaulis*
- *Bistorta vivipara*
- *Saxifraga oppositifolia*

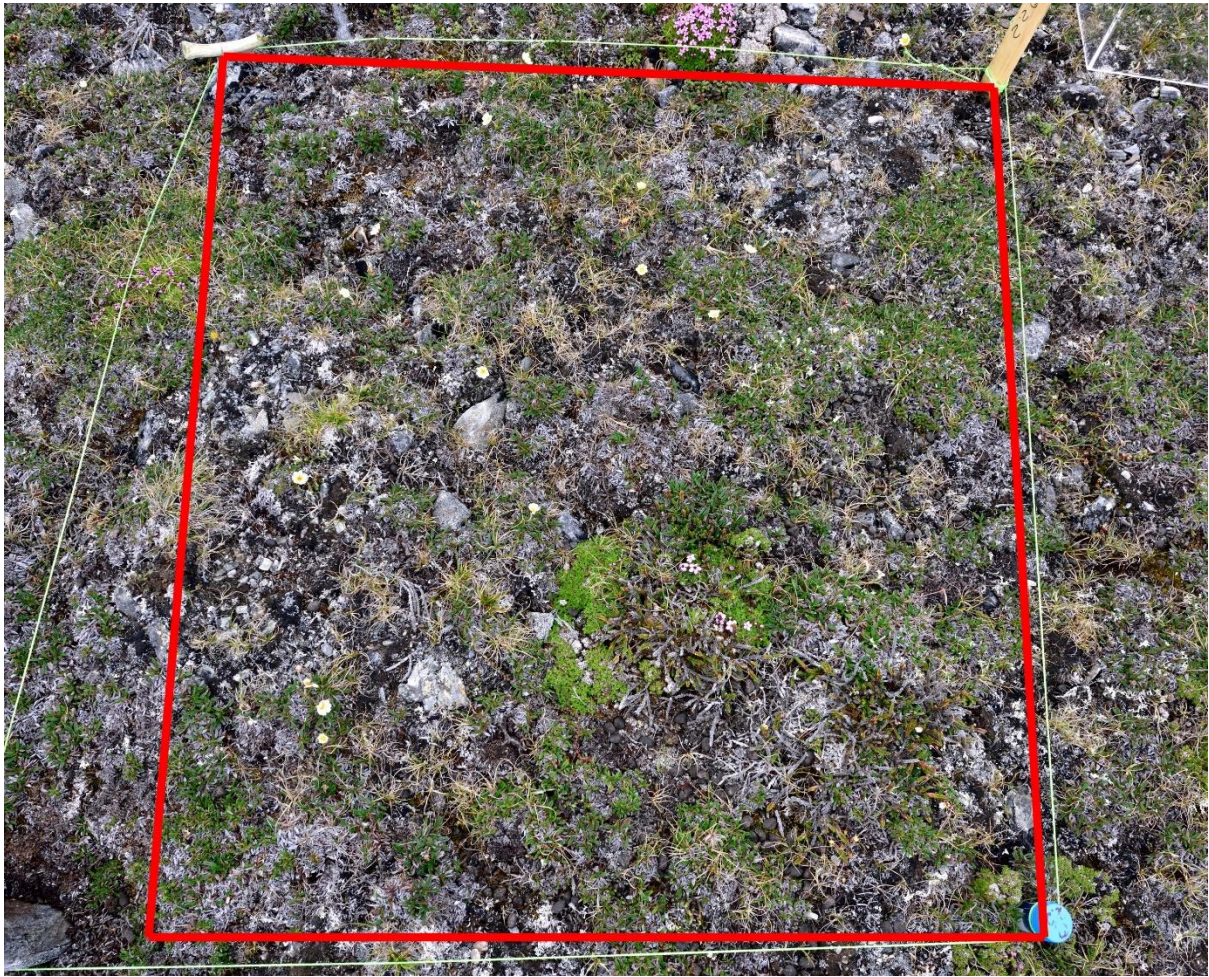
O 2.1

O 2.1



- *Silene acaulis*
- *Dryas octopetala* \*
- *Bistorta vivipara*
- *Poa* sp.
- *Saxifraga oppositifolia*

O 2.2



- *Silena acaulis*
- *Dryas octopetala*
- *Saxifraga oppositifolia*
- *Festuca* sp. \*
- *Cassiope tetragona*

O 2.3



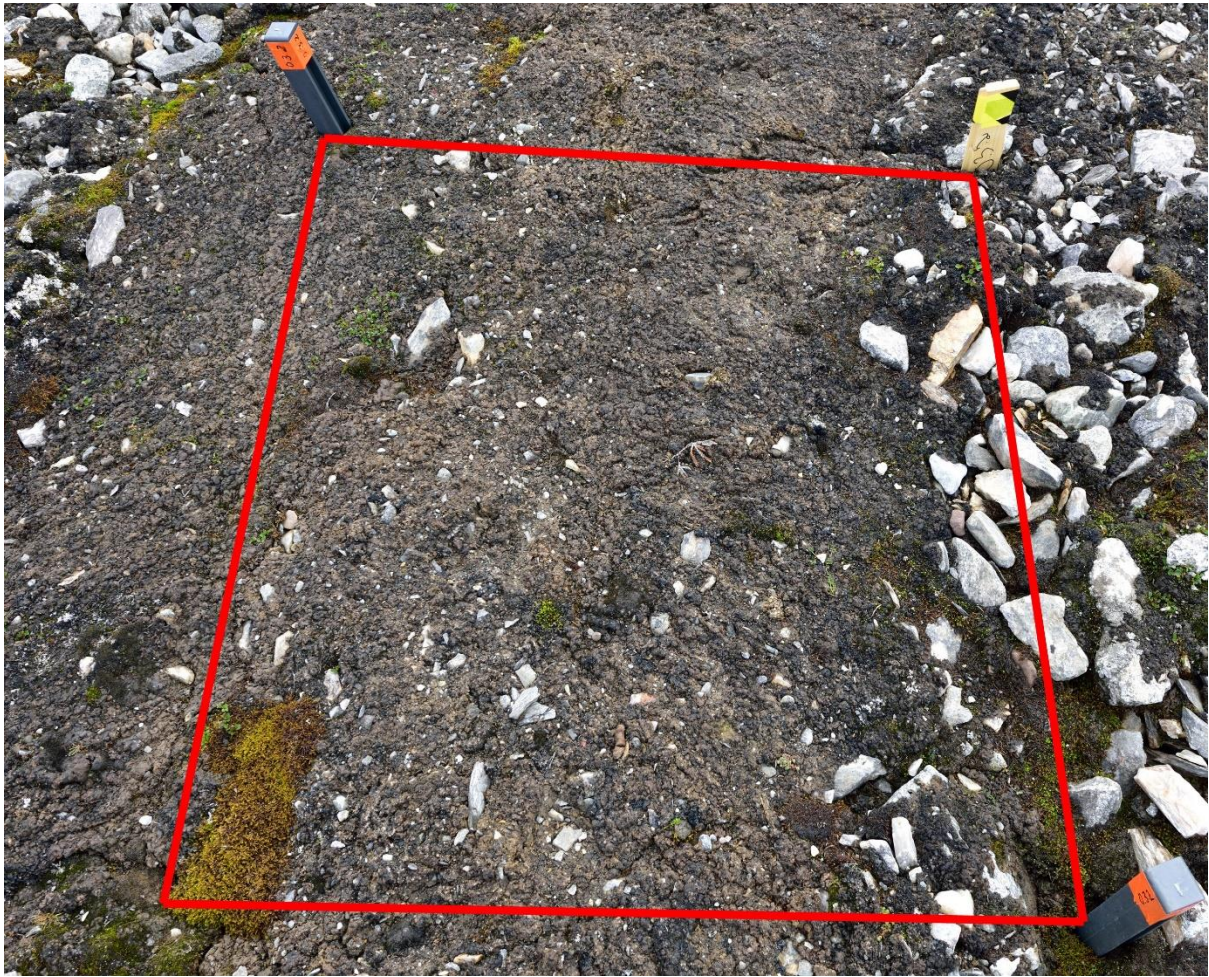
- *Dryas octopetala* \*
- *Saxifraga oppositifolia*
- *Luzula confua*
- *Poa* sp.
- *Festuca* sp.

O 3.1



- *Salix polaris* \*
- *Saxifraga oppositifolia*

O 3.2



- *Salix polaris* \*
- *Cardamine pratensis* ssp. *angustifolia*
- *Saxifraga oppositifolia*

O 3.3



- *Salix polaris*
- *Cassiope tetragona*
- *Saxifraga cernua*
- *Saxifraga oppositifolia*
- *Arenaria* sp.

Supplementary Figure 9

